

図 4.3. ポリオキシエチレンアルキルエーテル硫酸塩 (AES, エマール 170J (ポリオキシエチレン(1) ラウリルエーテル硫酸ナトリウム), エマール 270J (ポリオキシエチレン(2) ラウリルエーテル硫酸ナトリウム), エマール 20C (ポリオキシエチレン(3) ラウリルエーテル硫酸ナトリウム)) の FIA-MS の ESI スペクトル

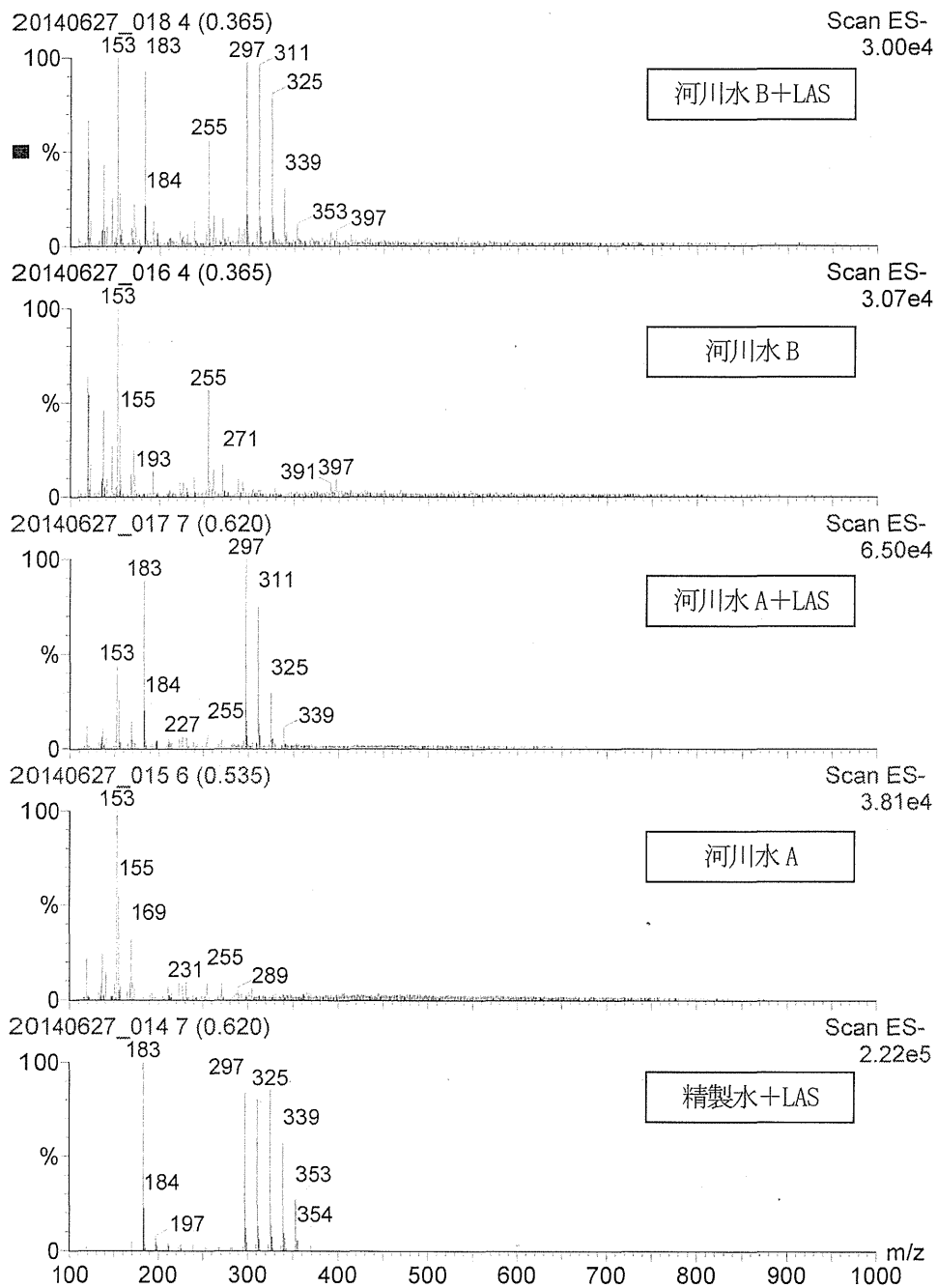


図 4-4. 市販の洗剤を河川水に LAS (C10-C14) 添加した時の FIA-MS の ESI スペクトル
 河川水：フレッシュ度 50%，LAS 添加濃度：1 mg/L

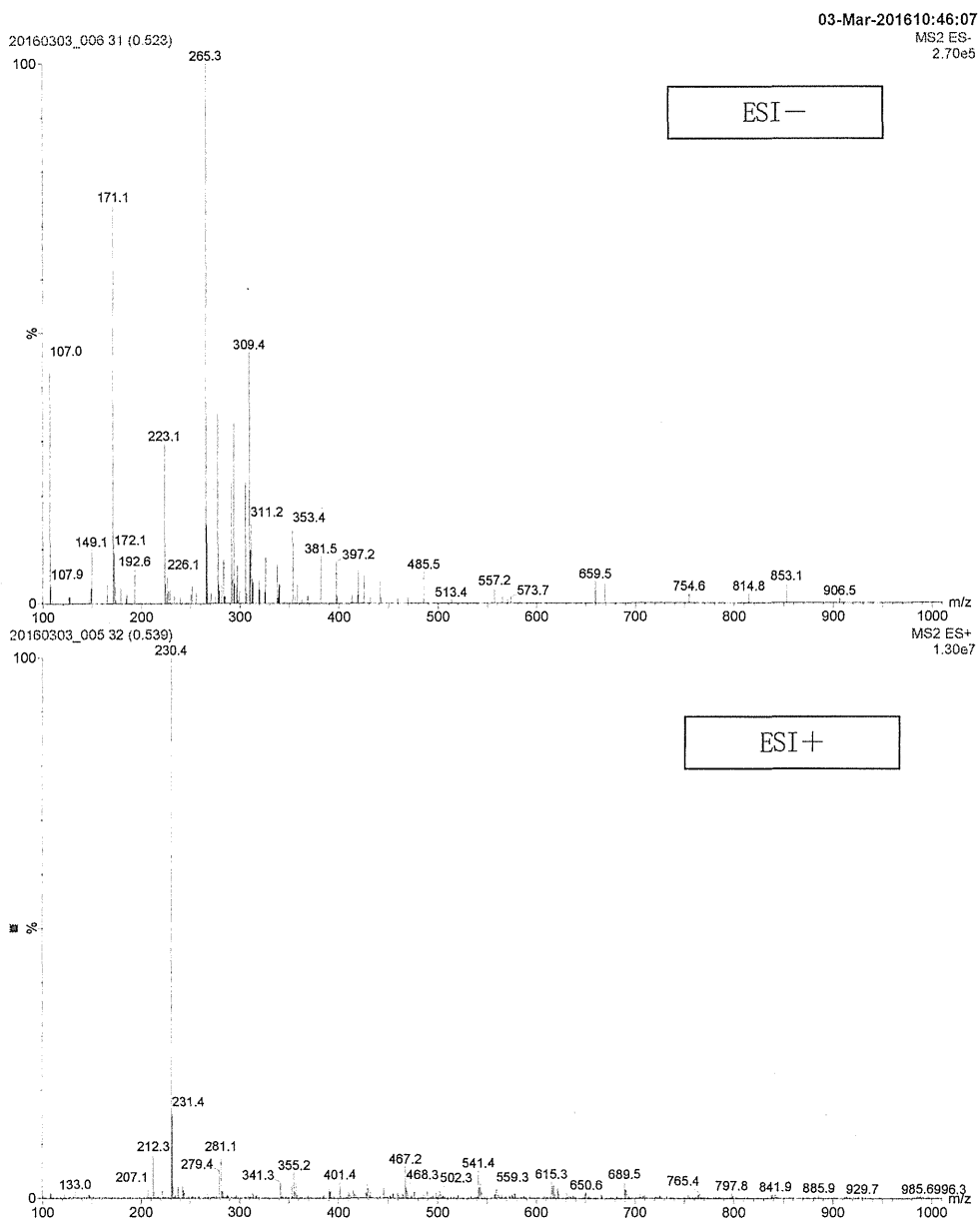


図 4-5. 市販の家庭用洗剤の FIA-MS の ESI スペクトル

家庭用洗剤の濃度 : 10 mg/L, 成分表示 (31%, AO, AES, POE, AS)

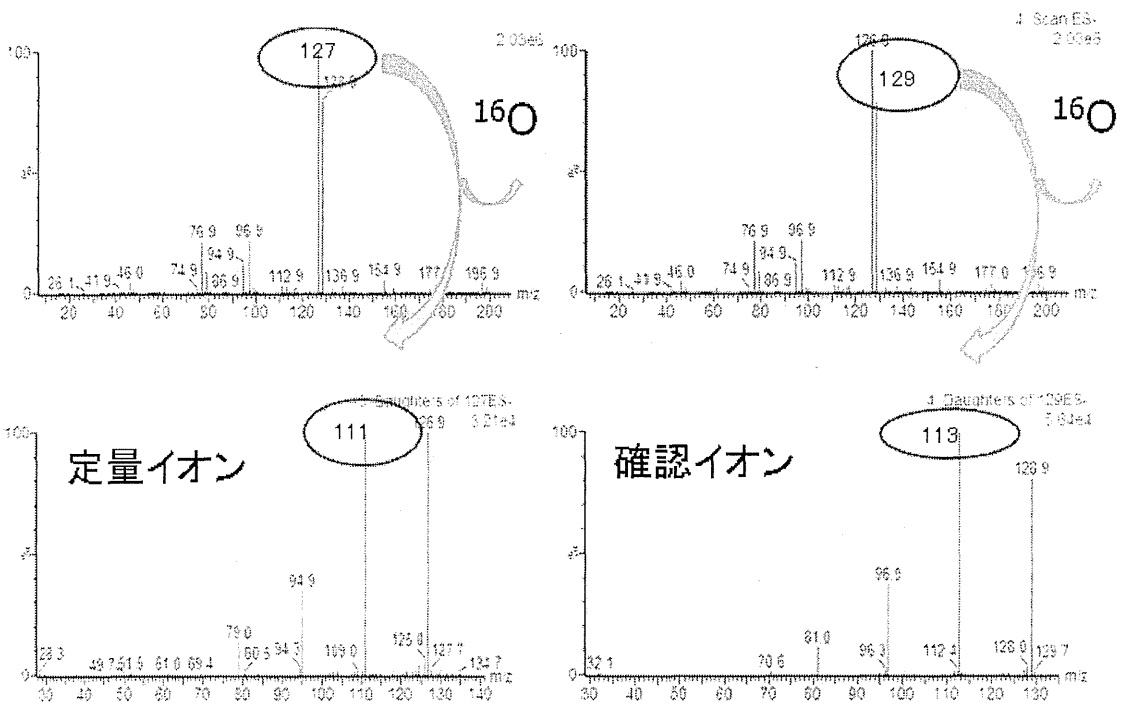


図 5-1 臭素酸イオンの MS/MS スペクトル

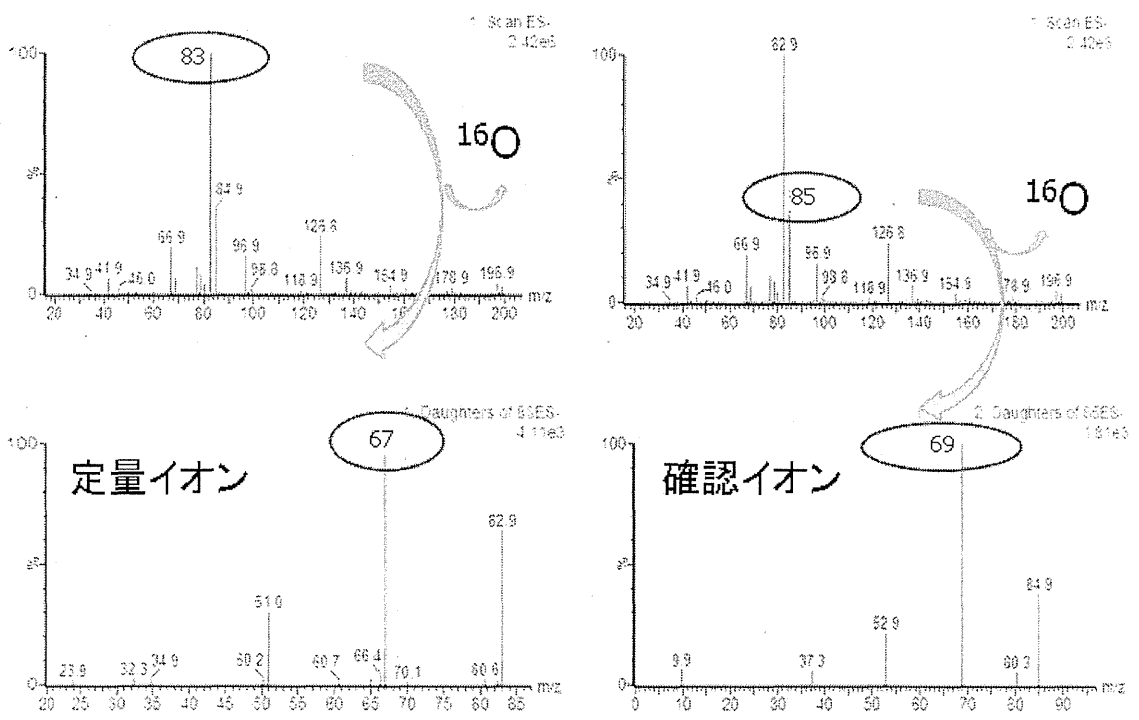


図 5-2 塩素酸イオンの MS/MS スペクトル

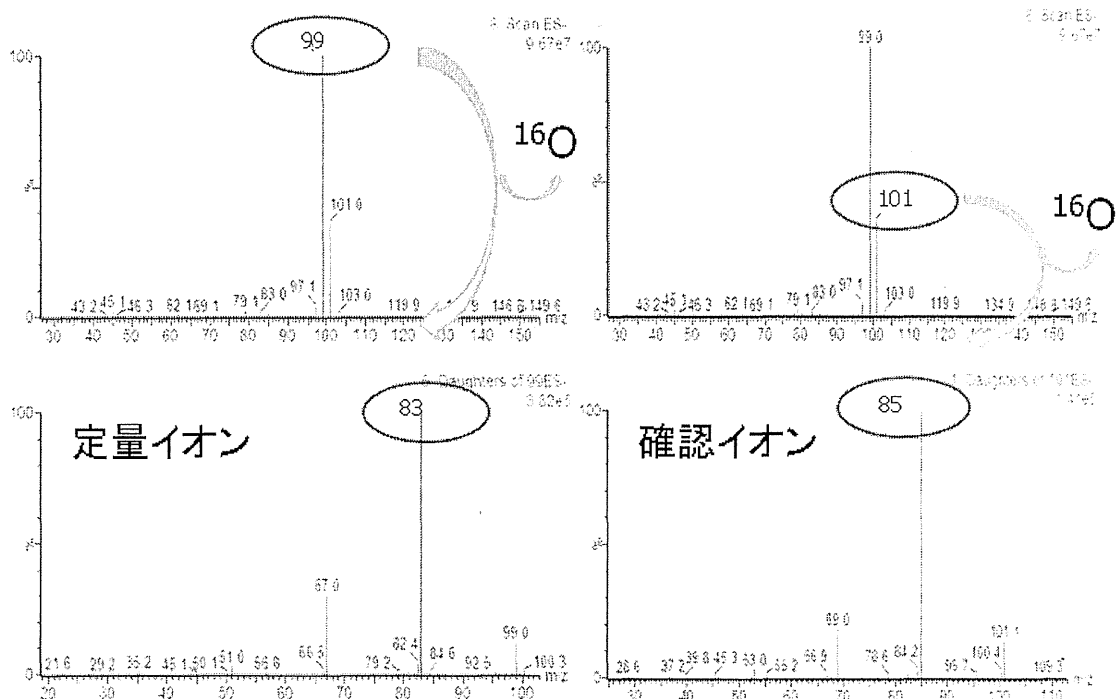


図 5-3 過塩素酸イオンのMS/MS スペクトル

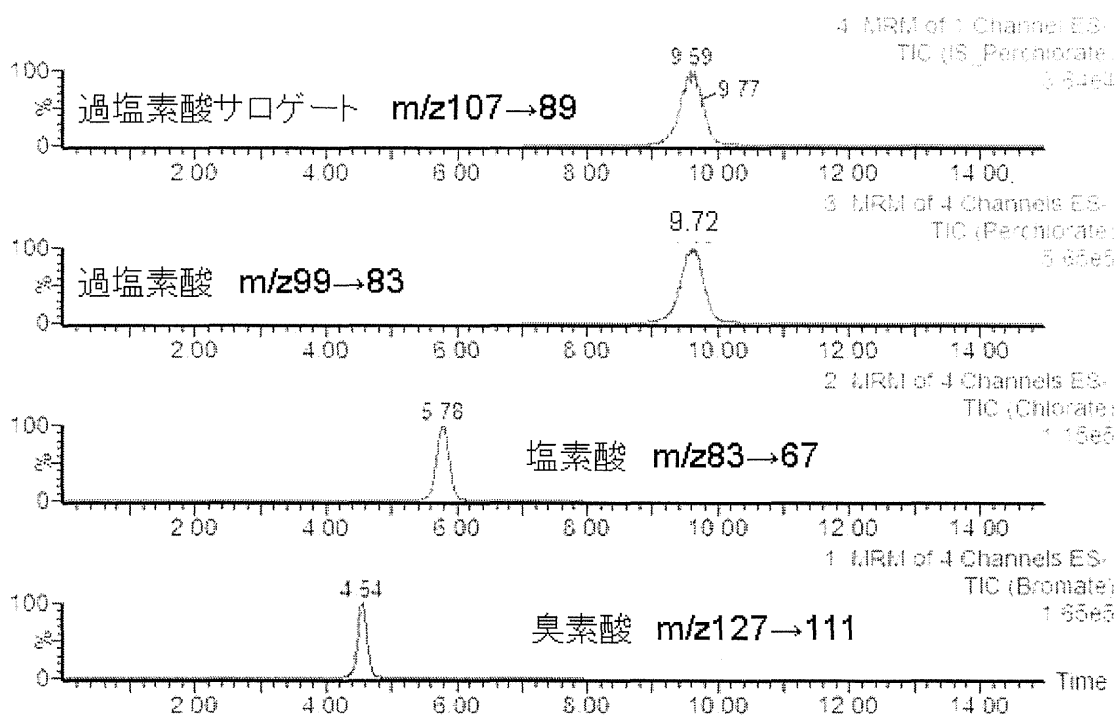


図 5-2 臭素酸、塩素酸および過塩素酸イオンのLC-MS/MS クロマトグラム

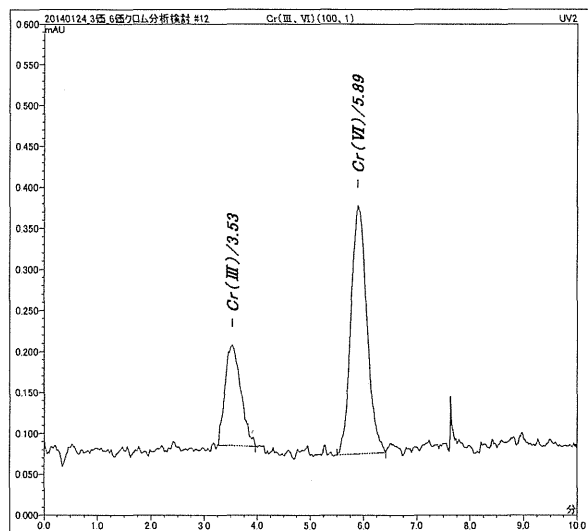


図 6-1. Cr (III) 100 μ g/L、Cr (VI) 1 μ g/L の混合液のクロマトグラム

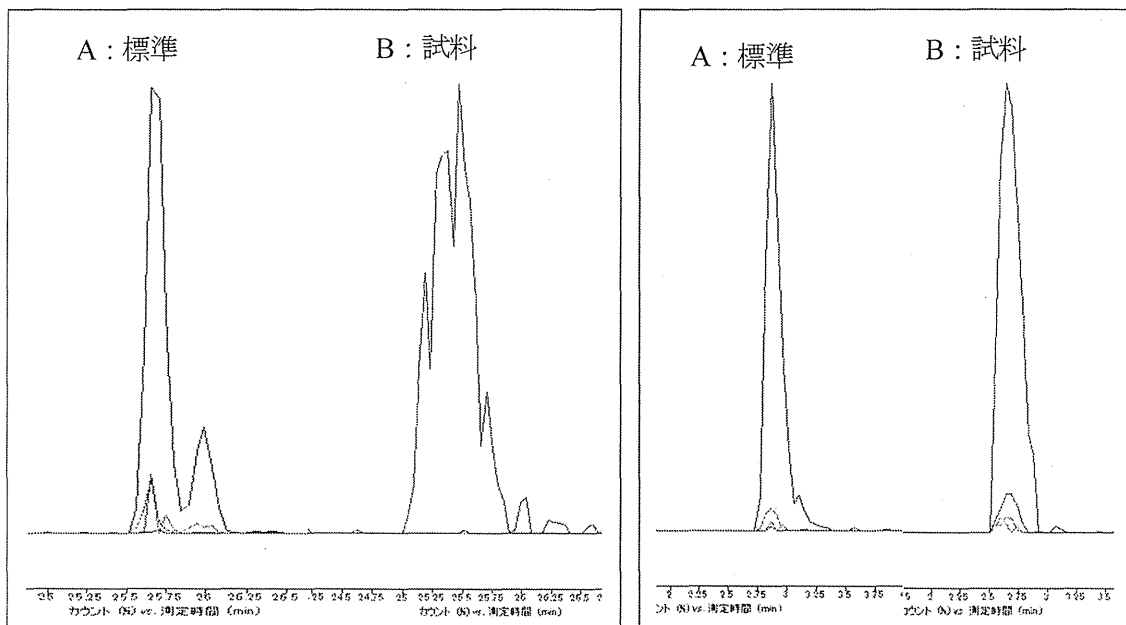


図 8-1. インソースフラグメントイオンによる誤検出の例 (Siduron) (左)

図 8-2. インソースフラグメントイオンによる検出の例 (Metformin) (右)

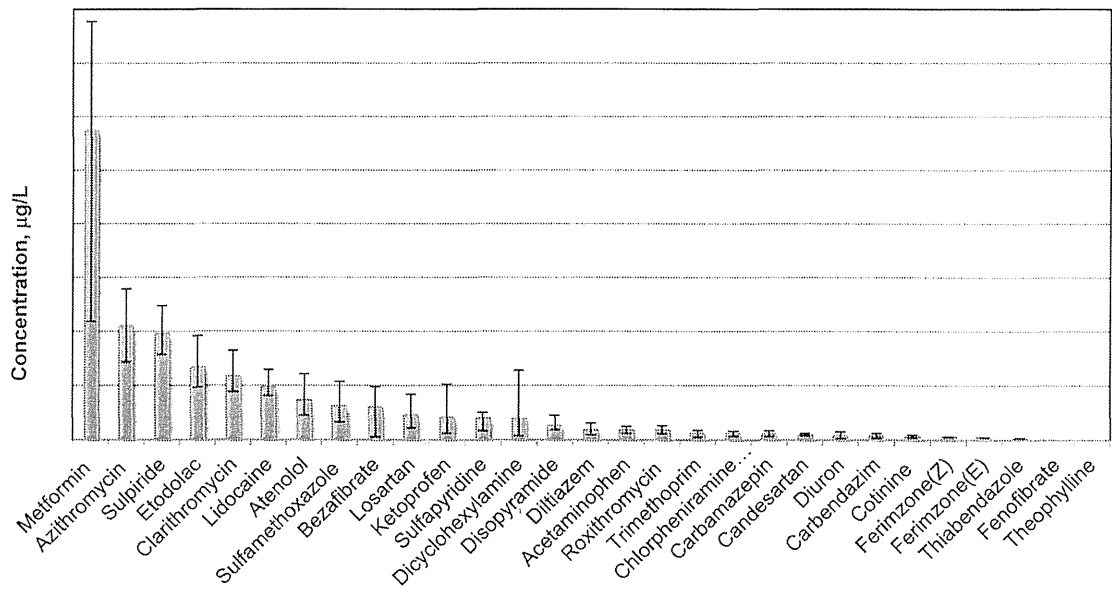


図 8-3. 北九州市内の 5 下水処理場での検出物質

研究成果の刊行に関する一覧表

書籍

著者氏名	論文タイトル名	書籍全体の編集者名	書籍名	出版社名	出版地	出版年	ページ
大野 浩一	知っておきたい新公衆衛生	一般財団法人 日本環境衛生 センター	公衆衛生リスク	一般財団 法人 日 本環境衛 生センタ ー	東京	2015	15-21
Kosaka, K., Asami M., and Kunikane S.	Perchlorate: Origin and occurrence in drinking water	Elias S.	Reference Module in Earth Systems and Environmental Sciences	Elsevier	Burlington	2013	1-11
浅見真理	第1編 水環境 を脅かす新た な汚染物質の 実態、第5章 1,4-ジオキサン	エヌ・ティー ・エス東京	排水汚水処理 技術集成	エヌ・テ ィー・エ ス東京	東京	2013	59-68
小坂浩司	第1編 水環境 を脅かす新た な汚染物質の 実態、第2章 過塩素酸・ NDMA	エヌ・ティー ・エス東京	排水・汚水処 理技術集成	エヌ・テ ィー・エ ス東京	東京	2013	29-38

雑誌

発表者氏名	論文タイトル名	発表誌名	巻号	ページ	出版年
Shirasaki, N., Matsushita, T., Matsui, Y. and Marubayashi, T.	Effect of aluminum hydrolyte species on human enterovirus removal from water during the coagulation process	Chemical Engineering Journal	284(1)	786-793	2016
泉山信司, 遠藤卓郎	水道における人への危害が問 題となる病原微生物とその対 策	水環境学会誌	39A (2)	54-58	2016
浅見真理, 松井佳彦	水道における化学物質・放射 性物質の管理と制御	水環境学会誌	39A (2)	48-53	2016

中井喬彦, 小坂浩司, 浅見真理, 秋葉道宏	LC- MS/MS法による水道水中の2,6-ジクロロ-1,4-ベンゾキノンの測定法検討と実態調査	水環境学会誌	38	67-73	2015
Kosaka, K., Asami, M., Ohkubo, K., and Akiba M.	Determination of a N-nitrosodimethylamine precursor in water using ultra-high performance liquid chromatography- tandem mass spectrometry	Analytical Sciences	31	769-772	2015
小坂浩司, 福田圭佑, 中村怜奈, 浅見真理, 越後信哉, 秋葉道宏	トリクロラミン生成能におよぼすオゾン処理の影響	土木学会論文集 G (環境)	71(7)	III_361-III_369	2015
Phattarapattamawong, S., Echigo, S. and Itoh, S.	Characterization of organic precursors for chlorinous odor before and after ozonation by a fractionation technique	Water Research	88(1)	836-843	2016
越後信哉	浄水処理過程における人為由来化合物からの消毒副生成物の生成機構	水環境学会誌	38(3)	89-94	2015
大野浩一	続：米国ウエストバージニア州における化学物質河川流出事故時の対応－事故後の調査について	水道	60(6)	32-40	2015
小熊久美子	イギリスの水道水質基準超過に対する枠組みと対応	水道	60(4)	28-34	2015
広瀬明彦	化学物質の毒性評価に基づいた水質基準値の設定	水環境学会誌	39	59-63	2016
Hirata-Koizumi, M., Fujii, S., Hina, K., Matsumoto, M., Takahashi, M., Ono, A., and Hirose, A.	Repeated dose and reproductive/developmental toxicity of long-chain perfluoroalkyl carboxylic acids in rats: perfluorohexadecanoic acid and perfluorotetradecanoic acid	Fundam Toxicol Sci.	2	177-190	2015
Chau, H., T., C., Kadokami, K., Duong, H., T., Kong, L., Nguyen, T., T., Nguyen, T., Q., and Y., Ito	Occurrence of 1153 organic micropollutants in the aquatic environment of Vietnam	Environmental Science Pollution Research	-	1-13	2015
Allinson, G., Allinson, M., and Kadokami, K.	Combining Passive Sampling with a GC-MS-Database Screening Tool to Assess Trace Organic Contamination of Rivers: a Pilot Study in Melbourne, Australia	Water Air Soil Pollution	-	226-230	2015

Matsui, Y., Nakao, S., Sakamoto, A., Taniguchi, T., Pan, L., Matsushita, T., and Shirasaki, N.	Adsorption capacities of activated carbons for geosmin and 2-methylisoborneol vary with activated carbon particle size : effects of adsorbent and adsorbate characteristics	Water Research	85	95-102	2015
Duong, H., T., Kadokami, K., Chau, H., T., C., Nguyen, T., Q., Nguyen, T., T., Kong, L.	Groundwater screening for 940 organic micro-pollutants in Hanoi and Ho Chi Minh City, Vietnam	Environmental Science Pollution Research	22(24)	19835-19847	2015
Shirasaki, N., Matsushita, T., Matsui, Y., Oshiba, A., Marubayashi, T., and Sato, S.	Improved virus removal by high-basicity polyaluminum coagulants compared to commercially available aluminum-based coagulants	Water Research	48	375-386	2014
Shirasaki, N., Matsushita, T., Matsui, Y., Urasaki, T., Kimura, M. and Ohno, K.	Virus removal by an in-line coagulation-ceramic microfiltration process with high-basicity polyaluminum coagulation pretreatment	Water Science and Technology: Water Supply	14(3)	429-437	2014
Kishida, N., Noda, N., Haramoto, E., Kawaharasaki, M., Akiba, M. and Sekiguchi, Y.	Quantitative detection of human enteric adenoviruses in river water by microfluidic digital polymerase chain reaction	Water science and technology	70(3)	555-60	2014
国立感染症研究所, 感染研感染症疫学センター	クリプトスポリジウム症およびジアルジア症 2014年7月現在	病原微生物検出情報月報(IASR) <特集>	35(8)	185-202	2014
浅見真理	水質事故対応の現状	水道	59(4)	30-35	2014
浅見真理	水道水源汚染に注意すべき物質の管理について	水環境学会誌	38(3)	84-88	2015
Narita, K., Matsui, Y., Iwao, K., Kamata, M., Matsushita, T. and Shirasaki, N.	Selecting pesticides for inclusion in drinking water quality guidelines on the basis of detection probability and ranking	Environment International	63	114-120	2014
小坂浩司, 福田圭佑, 浅見真理, 越後信哉, 秋葉道宏	トリクロラミン生成能におよぼす2段階塩素処理条件の影響	土木学会論文集 G (環境)	70 (7)	III_9-III_16	2014
Kosaka, K., Asami, M., Ohkubo, K., Iwamoto, T., Tanaka, Y., Koshino, H., Echigo, S. and Akiba, M.	Identification of a new N-nitrosodimethylamine precursor in sewage containing industrial effluents	Environmental Science & Technology	48	11243-11250	2014

浅見真理, 小坂浩司, 島崎大, 武井佳奈子	塩水電解による塩素酸・過塩 素酸の生成特性と電極材質に 関する検討	水環境学会誌	14 (5)	189-195	2014
Kosaka, K., Kobashigawa, N., Nakamura, R., Asami, M., Echigo, S. and Akiba, M.	Control of trichloramine formation by two-step chlorination in water purification processes	Water Science and Technology: Water Supply	14 (4)	650-656	2014
Kosaka, K., Asami, M., Nakai, T., Ohkubo, K., Echigo, S. and Akiba, M.	Formaldehyde formation from tertiary amine derivatives during chlorination	Science of the Total Environment	488- 489	325-332	2014
Sakuma, M., Matsushita, T., Matsui, Y., Aki, T., Isaka, M. and Shirasaki, N.	Mechanisms of trichloramine removal with activated carbon: Stoichiometric analysis with isotopically labelled trichloramine and theoretical analysis with a diffusion-reaction model	Water Research	68	839-848	2015
松浦直紀, 川瀬敬三, 片山慎介, 杜旭, 李 雪花, 陳景文, 門上 希和夫	網羅分析法を用いた中国長江 の化学物質調査	環境化学	24 (3)	83-92	2014
川瀬敬三, 門上希和 夫	GC-MS向け汎用ターゲットス クリーニングシステムの検討	分析化学	64 (1)	43-50	2015
Kong, L., Kadokami, K., Wang, S., Duong, H. T. and Chau, H. T. C.	Monitoring of 1300 organic micro-pollutants in surface waters from Tianjin, North China	Chemosphere	122	125-130	2015
門上希和夫	化学物質網羅分析法の開発と 途上国の環境調査	地球環境	19(2)	165-172	2014
Matsushita, T., Hirai, S., Ishikawa, T., Matsui, Y. and Shirasaki, N.	Decomposition of 1,4-dioxane by vacuum ultraviolet irradiation: Study of economic feasibility and by-product formation	Process Safety and Environmental Protection	94	528-541	2015
Kimura, M., Matsui, Y., Saito, S., Takahashi, T., Nakagawa, M., Shirasaki, N. and Matsushita, T.	Hydraulically irreversible membrane fouling during coagulation-microfiltration and its control by using high-basicity polyaluminum chloride	Journal of Membrane Science	477	115-122	2015
Ikari, M., Matsui, Y., Suzuki, Y., Matsushita, T. and Shirasaki, N.	Removal of iodide from water by chlorination and subsequent adsorption on powdered activated carbon	Water Research	68	227-237	2015

Kato, H., Fujii, S., Takahashi, M., Matsumoto, M., Hirata-Koizumi, M., Ono, A. and Hirose, A.	Repeated dose and reproductive/developmental toxicity of perfluorododecanoic acid in rats.	Environmental Toxicology	doi: 10.1002/tox.21996. [Epub ahead of print]		2015
Matsui, Y., Sakamoto, A., Nakao, S., Taniguchi, T., Matsushita, T., Shirasaki, N., Sakamoto, N. and Yurimoto, H.	Isotope microscopy visualization of the adsorption profile of 2-methylisoborneol and geosmin in powdered activated carbon	Environmental Science & Technology	48	10897-10903	2014
Takahashi, M., Ishida, S., Hirata-Koizumi, M., Ono, A., and Hirose, A.	Repeated dose and reproductive/developmental toxicity of perfluoroundecanoic acid in rats	J Toxicol Sci.	39(1)	97-108	2014
Matsushita T., Suzuki H., Shirasaki N., Matsui Y, and Ohno K.	Adsorptive virus removal with super-powdered activated carbon	Separation and Purification Technology	107	79-84	2013
Matsushita, T., Shirasaki, N., Tatsuki, Y. and Matsui, Y.	Investigating norovirus removal by microfiltration, ultrafiltration, and precoagulation-microfiltration processes using recombinant norovirus virus-like particles and real-time immuno-PCR	Water Research	47	5819-5827	2013
岸田直裕、原本英司、今野祥頭、泉山信司、浅見真理、秋葉道宏	水中のクリプトスポリジウム・ジアルジア検査における遺伝子検査法の実用性に関する検討	土木学会論文集 G (環境)	69 (7)	III_631-III_637	2013
泉山信司、黒木俊郎	水系感染する病原微生物 (クリプトスポリジウムおよびレジオネラ) への対策	水環境学会誌	36(A)N 0.5	161-164	2013
小坂浩司、浅見真理、佐々木万紀子、松井佳彦、秋葉道宏	水道統計を基にした水道事業者等における農薬類の測定計画の解析	土木学会論文集 G (環境)	69 (7)	III_363-III_374	2013
越後信哉、森田悠斗、伊藤禎彦	琵琶湖・淀川流域の水道水におけるヨウ素系消毒副生成物の存在実態	土木学会論文集 G (環境)	69 (7)	III_385-III_392	2013
Quan, D., Okashita, R., Yanagibashi, Y., Echigo, S., Ohkouchi, Y., Itoh, S., and Jinno, H.	Exposure to haloacetic acids via typical components of the Japanese diet and their allocations of drinking water ingestion to total exposure	J. Water Supply: Res. Technol.-AQUA	62 (4)	224-233	2013

Echigo, S., Itoh, S., Ishihara S., Aoki, Y., and Hisamoto, Y.	Reduction of chlorinous odor by the combination of oxidation and ion-exchange treatment	J. Water Supply: Res. Technol.- AQUA	63 (2)	106-113	2014
伊藤禎彦	消毒副生成物の飲用寄与率と 水道水質基準	環境衛生工学研 究	27 (4)	3-22	2013
Asami, M., Yoshida, N., Kosaka, K., Ohno, K., and Matsui, Y.	Contribution of tap water to chlorate and perchlorate intake: A market basket study	Science of the Total Environment	463-464	199-208,	2013
Niizuma, S., Matsui, Y., Ohno, K., Itoh, S., Matsushita, T., and Shirasaki, N.	Relative source allocation of TDI to drinking water for derivation of a criterion for chloroform: a Monte-Carlo and multi-exposure assessment	Regulatory Toxicology and Pharmacology	67	98-107	2013
Matsumoto, M., Yamaguchi, M., Yoshida, Y., Senuma, M., Takashima, H., Kawamura, T., Kato, H., Takahashi, M., Hirata-Koizumi, M., Ono, A., Yokoyama, K., and Hirose, A.	An antioxidant, N,N'-diphenyl-p- phenylenediamine (DPPD), affects labor and delivery in rats: A 28-day repeated dose test and reproduction/developmental toxicity test	Food Chem Toxicol.	56	290-296	2013
Matsui, Y., Ishikawa, T. B., Kimura, M., Machida, K., Shirasaki, N. and Matsushita, T.	Aluminum concentrations of sand filter and polymeric membrane filtrates: a comparative study	Separation and Purification Technology	119	58-65	2013
金見 拓	利根川ホルムアルデヒド水質 事故の概要と提起された課題	日本リスク研究 学会誌	23 (2)	57-64	2013
浅見真理、小坂浩司 大野浩一	水道側から見たホルムアルデ ヒド水質事故関連のリスク管 理制度とその課題	日本リスク研究 学会誌	23 (2)	71-76	2013
大野浩一	利根川水系ホルムアルデヒド 水質事故をめぐる考察と給水 停止に対する住民のパーセプ ションについて	日本リスク研究 学会誌	23 (2)	81-85	2013
門上希和夫	微量化学物質による環境汚染 の安全性評価－ノンターゲッ ト分析の必要性と手法－（総 説）	安全工学	52(3)	155-161	2013

研究成果の刊行物・別刷



Effect of aluminum hydrolyte species on human enterovirus removal from water during the coagulation process



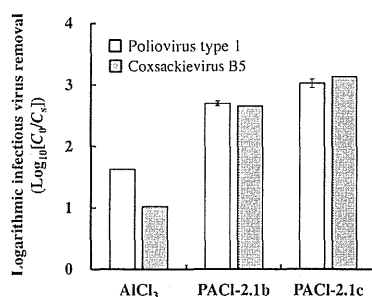
N. Shirasaki*, T. Matsushita, Y. Matsui, T. Marubayashi

Division of Environmental Engineering, Faculty of Engineering, Hokkaido University, N13W8, Sapporo 060-8628, Japan

HIGHLIGHTS

- Differences in aluminum species in coagulant affected enterovirus removal.
- High Al_c content PACI with basicity 2.1 effectively removed enteroviruses.
- Al_{30} species in PACI probably played a major role in enterovirus removal.
- PV and CV removal ratios were almost the same during coagulation.
- Viruses were removed mainly by coprecipitation into growing aluminum hydroxide.

GRAPHICAL ABSTRACT



ARTICLE INFO

Article history:

Received 18 August 2015

Received in revised form 5 September 2015

Accepted 7 September 2015

Available online 25 September 2015

Keywords:

Aluminum hydrolyte species

Basicity

Charge neutralization

Coagulation

Coxsackievirus

Poliovirus

ABSTRACT

We prepared different types of aluminum-based coagulants, consisting of mainly monomeric aluminum species, polymeric aluminum species, or colloidal aluminum species, to investigate the effect of aluminum hydrolyte species on the removal of two types of human enteroviruses, poliovirus (PV) type 1 and the free-chlorine-resistant virus coxsackievirus (CV) B5, from lake and river water samples during the coagulation process. We found that differences in the distribution of the aluminum hydrolyte species in the coagulant affected the removal of these enteroviruses during coagulation: the removal ratios of PV and CV observed with polyaluminum chloride (PACI) with a high colloidal aluminum content and a basicity of 2.1 (i.e., PACI-2.1c) were larger than those observed with high monomeric aluminum content coagulant (i.e., AlCl₃ solution) and with high polymeric aluminum content coagulant (PACI-2.1b). Unlike AlCl₃ or PACI-2.1b, PACI-2.1c contains Al₃₀ species, indicating that Al₃₀ species probably play a major role in the removal of enteroviruses. The PV and CV removal ratios were almost identical, regardless of the coagulant type or viral quantification method used (plaque-forming unit method or real-time polymerase chain reaction method), suggesting that PV and CV behaved similarly during the coagulation process. We also experimentally confirmed that the main mechanism for virus removal was coprecipitation into growing aluminum hydroxide during charge neutralization; virus adsorption onto formed aluminum hydroxide flocs also contributed to virus removal, but played a limited role.

© 2015 Elsevier B.V. All rights reserved.

1. Introduction

Coagulation is an important process in water treatment to remove contaminants. Of the various coagulants used for the coagulation process, prepolymerized aluminum coagulants, such as polyaluminum chloride (PACI), have been widely used in drinking

* Corresponding author. Tel./fax: +81 11 706 7282.

E-mail address: nobutaka@eng.hokudai.ac.jp (N. Shirasaki).

water treatment because of their superiority to traditional aluminum coagulants [e.g., aluminum chloride (AlCl_3) and alum]; PACls are more efficient, are less dependent on temperature and pH, can be used at a lower dosage, and produce less sludge [1–3].

The hydrolysis reactions of aluminum species are very complicated, depending on water quality and coagulant type, which contribute to the formation of various aluminum hydrolyte species [4,5]. The ferron method has been employed extensively to categorize the aluminum hydrolyte species into the following three fractions on the basis of the kinetic differences between the reactions of the aluminum species and the ferron reagent: monomeric species, fast-reacting polymeric species, and slow-reacting colloidal species, denoted as Al_a , Al_b , and Al_c , respectively [6]. Al_a is composed mainly of monomeric species such as Al^{3+} , $\text{Al}(\text{OH})^{2+}$, and $\text{Al}(\text{OH})_2^+$; dimeric and trimeric species such as $[\text{Al}_2(\text{OH})_2]^{4+}$ and $[\text{Al}_3(\text{OH})_4]^{5+}$ are also classified as Al_a [1]. Al_b is the intermediate polymeric species, and many researchers have suggested that Al_b could include the Al_{13} species $[\text{AlO}_4\text{Al}_{12}(\text{OH})_{24}(\text{H}_2\text{O})_{12}]^{7+}$, which was identified by using liquid ^{27}Al nuclear magnetic resonance (NMR) spectroscopy [1,4,7]. Al_c is the large polymer or colloidal species, and the Al_{30} species $[\text{Al}_{30}\text{O}_8(\text{OH})_{56}(\text{H}_2\text{O})_{24}]^{18+}$, which was also identified by using liquid ^{27}Al NMR spectroscopy, is classified into Al_c [7].

The distribution of aluminum hydrolyte species in PACls and the formation of specific aluminum hydrolyte species such as Al_{13} and Al_{30} might be controlled by basicity ($[\text{OH}^-]/[\text{Al}^{3+}]$), aluminum concentration, rate and mode of neutralization, and reaction temperature and time during PACl preparation [7–9]. In addition, the aluminum species originally contained in the PACls or formed *in situ* during the coagulation process play an important role in determining the behavior and efficiency of coagulants [4,5]. For these reasons, many researchers have paid particular attention to the effects of aluminum species on coagulation efficiency. For example, Yan *et al.* [1] reported that the removal efficiencies of turbidity and dissolved organic carbon (DOC) correlate well with the Al_c content and Al_b content in PACls, respectively, whereas both Al_b and Al_c content influence the removal of ultraviolet absorbance at 254 nm (UV254, an indicator of natural organic matter concentration). Our research group [10] and Duan *et al.* [11] reported that the residual aluminum concentration in treated water is closely related to the Al_a content in the PACl: low Al_a content in the PACl yields a low residual aluminum concentration. Moreover, among the identified aluminum hydrolyte species in PACls, Al_{13} and Al_{30} are believed to be effective coagulation species due to their strong charge neutralization capability and high structural stability [12]. In fact, Lin *et al.* [13] showed that DOC removal by PACl with a high Al_{13} content is superior to that by commercially available PACl with a low Al_{13} content at about pH 6, and that a lower dosage of high Al_{13} content PACl than of commercially available PACl is required for optimal DOC removal. Zhang *et al.* [2] reported that compared to AlCl_3 , PACl with a high Al_{30} content exhibits greater UV254 removal efficiency and leads to a lower residual aluminum concentration across a broad pH range and a wide coagulant dosage range. In addition, Hu *et al.* [4] showed the value of PACl with a high Al_{13} content for the removal of arsenic (As) from As-spiked tap water over a broad pH range, and concluded that the removal efficiencies of As(V) correlated with the amount of Al_{13} species present. Moreover, Mertens *et al.* [14] reported that PACl with a high Al_{30} content contributed to the efficient removal of As(III) and As(V) from As-contaminated groundwater at pH 7–8.

Human enteric viruses, which are frequently present in contaminated drinking water sources and do not settle from suspension by gravity, can also be removed by the coagulation process with PACl [15]. However, the relationship between the removal efficiencies of human enteric viruses and the aluminum hydrolyte species in PACl remains unclear, and the role of Al_{13} and Al_{30} species in the

removal of human enteric viruses has not been investigated. We recently reported that laboratory-prepared PACl with a high Al_c content removed bacteriophages (i.e., viruses that infect bacteria) more efficiently than did laboratory-prepared PACl with a high Al_b content or commercially available PACls, suggesting that the Al_{30} species in high Al_c content PACl probably play a major role in bacteriophage removal during the coagulation process [16]. Because the removal ratios of human enteric viruses including poliovirus were different to and smaller than that of bacteriophage MS2 in the coagulation process with aluminum-based coagulant [17], the effect of aluminum hydrolyte species on virus removal, and specifically the role of Al_{13} and Al_{30} , may also differ between human enteric virus and bacteriophage removal. In addition, because human enteric viruses have high resistance to free-chlorine disinfection compared with human enteric bacteria [18], and the increase in free-chlorine dosage needed for sufficient disinfection of human enteric viruses sometimes results in the formation of high levels of toxic disinfection by-products [19], improvements in the coagulation efficiency of virus removal are highly desired. Therefore, the identification of aluminum hydrolyte species in aluminum-based coagulants that efficiently remove human enteric viruses from drinking water would improve coagulation efficiency and support the development of novel aluminum-based coagulants for the prevention and control of waterborne disease caused by exposure to such viruses through drinking water.

Here, we conducted batch coagulation experiments to investigate the effect of aluminum hydrolyte species on the removal of human enteric viruses, specifically human enteroviruses, by comparing five coagulants with different distributions of aluminum hydrolyte species. In addition, we investigated the mechanism of human enterovirus removal during the coagulation processes by examining the coprecipitation of enteroviruses into growing aluminum hydroxide during charge neutralization and the absorption of enteroviruses onto preformed aluminum floc particles. Because poliovirus (PV) type 1 is commonly used as representative of human enteric viruses [17,20], and coxsackievirus (CV) B5 has high resistance to free-chlorine disinfection compared with other types of CVs (e.g., CV B3 and CV B4) and other human enteric viruses including PVs (types 1, 2 and 3), echoviruses (EV types 1 and 11), and adenoviruses (AdV types 40 and 41) [21,22], we chose these two human enteroviruses for use in our study.

2. Materials and methods

2.1. Source water and coagulants

Lake water and river water were sampled from Lake Imbanuma in Chiba, Japan, and the Toyohira River in Sapporo, Japan, on 17 November 2014 (water quality data are shown in Table 1). The source water samples were stored at 4 °C until use, and the temperature was adjusted to 20 °C prior to use.

Five aluminum-based coagulants were used for the coagulation experiments (Table 2). AlCl_3 solution was prepared by dilution of reagent-grade aluminum(III) chloride hexahydrate ($\text{AlCl}_3 \cdot 6\text{H}_2\text{O}$, Wako Pure Chemical Industries, Osaka, Japan) dissolved in Milli-Q water (Milli-Q Advantage, Millipore Corp., Billerica, MA, USA).

Table 1
Water quality data for the source water samples.

	Lake water	River water
pH	8.6	7.4
Turbidity (NTU)	40.5	1.5
DOC (mg/L)	2.2	0.9
UV260 (cm^{-1})	0.052	0.024
Alkalinity (mg- CaCO_3/L)	84.1	17.5

Table 2
Characteristics of the aluminum-based coagulants used in the present study.

Coagulants	Basicity	Aluminum concentration (g-Al/L)	Relative density at 20 °C	Aluminum species distribution		
				Al _a (%)	Al _b (%)	Al _c (%)
AlCl ₃	0.0	13.5	1.1	88.6	11.4	0.0
PACI-2.1b	2.1	1.5	1.0	15.5	73.9	10.6
PACI-2.4b	2.4	1.6	1.0	5.3	87.2	7.5
PACI-2.1c	2.1	10.4	1.1	11.8	6.0	82.2
PACI-2.4c	2.4	10.3	1.1	3.6	8.7	87.7

Four nonsulfated high-basicity PACIs (PACI-2.1b, PACI-2.4b, PACI-2.1c, and PACI-2.4c, where 2.1 and 2.4 are basicity values, and “b” and “c” indicate high Al_b and Al_c content, respectively, as measured by the ferron method; see below) were prepared by using a base titration method in our laboratory. A certain amount of 0.25 M AlCl₃ solution was transferred to a glass beaker, and then the solution was titrated with 0.15 M NaOH at a constant rate (4 mL/min) by use of a peristaltic pump to achieve the targeted basicity value to prepare PACI-2.1b and PACI-2.4b. The solution in the beaker was stirred at 630 rpm and the temperature was maintained at 90–95 °C by a hot plate stirrer during the titration. PACI-2.1c and PACI-2.4c were prepared from 1.5 M AlCl₃ solution and 0.9 M NaOH by using the same method just described, and then the solution was continuously stirred at 630 rpm and kept at 90–95 °C for 24 h after the titration. All coagulants were used in batch coagulation experiments immediately after dilution with Milli-Q water.

2.2. Characterization of coagulants

The aluminum hydrolyte species in the coagulants were analyzed by using the ferron method and liquid ²⁷Al NMR spectroscopy after the coagulants were diluted with Milli-Q water to a concentration of 2.7 g-Al/L, that is, 0.1 M-Al (for AlCl₃ solution, PACI-2.1c, and PACI-2.4c), or 1.35 g-Al/L, that is, 0.05 M-Al (for PACI-2.1b and PACI-2.4b), depending on the aluminum concentration in the coagulant (Table 2). The positive colloid charges of the coagulants (diluted with Milli-Q water to 1–3 mg-Al/L) were determined by using a colloid titration method. The details of the ferron method, liquid ²⁷Al NMR analysis, and the colloid titration method were described in our previous report [16].

2.3. Human enteroviruses

The Sabin attenuated LSc/2ab strain of PV type 1 was kindly provided by Dr. Hiroyuki Shimizu, National Institute of Infectious Diseases, Tokyo, Japan. The Faulkner strain of CV B5 (ATCC VR-185) was obtained from American Type Culture Collection (ATCC, Manassas, VA, USA). PV and CV were propagated in buffalo green monkey kidney epithelial cells (BGM cells), kindly supplied by Dr. Daisuke Sano, Hokkaido University, Sapporo, Japan, and were maintained in 1X Eagle's minimum essential medium (EMEM, with phenol red, Nissui Pharmaceutical Co. Ltd., Tokyo, Japan) supplemented with 10% (v/v) heat-inactivated fetal bovine serum (Life Technologies, Carlsbad, CA, USA), 2 mM l-glutamine (Life Technologies), 100 units/mL penicillin, 100 µg/mL streptomycin (Pen Strep, Life Technologies), and 1.125 g/L NaHCO₃. One milliliter of 10-fold diluted PV or CV stock solution [diluted with Dulbecco's phosphate-buffered saline (without Ca and Mg ions, Nissui Pharmaceutical Co. Ltd.)] was inoculated onto a monolayer of approximately 90% confluent BGM cells in a 75 cm² flask. The flask was then incubated in a humidified incubator at 37 °C in 5% CO₂ for 30 min. After this incubation, 20 mL of 1X EMEM was added to the flask, and then the flask was incubated at 37 °C in 5% CO₂ for

2–3 days until 100% cytopathic effects of BGM cells were confirmed, at which point the viruses were harvested. Following incubation, viruses were released from the infected BGM cells by freezing at –83 °C and thawing at 37 °C, three times. The PV or CV culture solution was passed through a membrane filter (nominal pore size 0.2 µm, hydrophilic cellulose acetate; Dismic-25cs, Toyo Roshi Kaisha, Ltd., Tokyo, Japan) to prepare the PV or CV stock solution. The filtrate was further purified by using a tangential-flow filtration cassette (nominal molecular weight cutoff 1,000,000, regenerated cellulose (RC); Pellicon XL, Millipore Corp.), and by using a centrifugal filter device (nominal molecular weight cutoff 100,000, RC; Amicon Ultra-15, Millipore Corp.) twice to ensure that only a minimal amount of DOC components from the culture medium was introduced into the virus-spiked source water (see Section 2.4). The concentrations of the purified PV and CV solutions were approximately 10⁶ and 10⁷ PFU/mL, respectively.

2.4. Coagulation experiments

Batch coagulation experiments were conducted with 300 mL of virus-spiked source water in square plastic beakers at 20 °C. The purified PV or CV solution (see Section 2.3) was added to the beaker at a concentration of approximately 10³ PFU/mL (C₀). Because the purified virus solution was diluted during spiking of the source water, virus addition contributed less than 0.2–0.3 mg/L of unintentional DOC carryover. After enough HCl or NaOH was added to the spiked water to bring the final pH to a target value of 6, 7, or 8, coagulant was injected into the water. Because the dosages used in the actual drinking water treatment plants for the treatments of Lake Imba-numa water (Kashiwai drinking water treatment plant, Chiba, Japan) and Toyohira River water (Moiwa drinking water treatment plant, Sapporo, Japan) were 2.70 and 1.08 mg-Al/L, respectively, on the day we sampled the source water, those coagulant dosages were used in the present study. The water was stirred rapidly for 1 min ($G = 200 \text{ s}^{-1}$, 94 rpm) and then slowly for 10 min ($G = 20 \text{ s}^{-1}$, 20 rpm) with an impeller stirrer. The water was then left at rest for 60 min to allow the generated aluminum floc particles to settle. Supernatants were taken from the beaker after settling for quantification of the PV or CV concentration (C_s).

To investigate the importance of charge neutralization on virus removal during the coagulation process, 300 mL of source water without virus addition was coagulated rapidly for 1 min at pH 7, as described above. Following charge neutralization, the purified PV solution was added to the beaker at approximately 10³ PFU/mL (C₀). The water was stirred slowly for 10 min and then left at rest for 60 min. Supernatants were taken from the beaker after settling for quantification of the PV concentration (C_s). To assess virus adsorption onto preformed aluminum floc particles, 300 mL of source water without virus addition was coagulated rapidly for 1 min and then slowly for 10 min at pH 7, as described above, to generate aluminum floc particles. Following the formation of the aluminum floc particles, the purified PV solution was added to the beaker at approximately 10³ PFU/mL (C₀). The water was stirred slowly for 10 min and then left at rest for 60 min. Supernatants were taken from the beaker after settling for quantification of the PV concentration (C_s).

2.5. Enterovirus assay

Infectious PV or CV was quantified by use of a plaque assay. Approximately 90% confluent BGM cells in a 75 cm² flask were seeded in 6-well (8.96 cm²/well) tissue culture plates. One milliliter of 10-fold serially diluted sample (diluted with 1X EMEM supplemented with 2 mM l-glutamine, 100 units/mL penicillin, 100 µg/mL streptomycin, and 1.125 g/L NaHCO₃) or of 2-fold diluted sample (diluted with 2X EMEM containing 4 mM l-glutamine, 200 units/mL penicillin, 200 µg/mL streptomycin, and

2.25 g/L NaHCO₃) was inoculated onto a monolayer of 100% confluent BGM cells in a 6-well plate, which was then incubated in a humidified incubator at 37 °C in 5% CO₂ for 90 min. After this incubation, the inoculum was removed by inverting the plate, and then 3 mL of agar overlay was applied to the monolayer. The overlay was prepared by combining equal volumes of 2.5% (w/v) agarose (Agar-EPI, Nacalai Tesque, Inc., Kyoto, Japan) and 2X EMEM (without phenol red, Nissui Pharmaceutical Co. Ltd.) supplemented with heat-inactivated fetal bovine serum (2% for PV, 20% for CV), 4 mM l-glutamine, 200 units/mL penicillin, 200 µg/mL streptomycin, and 2.25 g/L NaHCO₃. After the addition of the agar overlay, the plate was incubated at 37 °C in 5% CO₂ for 2 days. After this incubation, the cell monolayer was stained with 1 mL of 0.15 g/L neutral red at 37 °C in 5% CO₂ for 3 h, and then excess stain was discarded by inverting the plate. Plaques of each well were counted for 1–3 days (1 day for PV, 3 days for CV) after discarding the neutral red until no new plaques appeared. The average plaque count of triplicate wells or twelve wells prepared from one sample was considered as the infectious PV or CV concentration for that sample.

The real-time polymerase chain reaction (PCR) method, which detects all viruses regardless of their infectivity or the presence of aggregates, was used to quantify viral RNA. PV and CV RNA was specifically quantified by using the real-time reverse transcription-PCR (RT-PCR) method. Viral RNA was extracted from 200 µL of sample with a QIAamp MinElute Virus Spin Kit (Qiagen K. K., Tokyo, Japan) to obtain a final volume of 20 µL. The extracted RNA solution was added to a High-Capacity cDNA Reverse Transcription Kit with RNase Inhibitor (Applied Biosystems Japan, Tokyo, Japan) for the RT reaction, which was conducted at 25 °C for 10 min, 37 °C for 120 min, and 85 °C for 5 s, with subsequent cooling to 4 °C in the thermal cycler (Thermal Cycler Dice Model TP600, Takara Bio Inc., Otsu, Japan). The cDNA solution was then amplified with a TaqMan Universal PCR Master Mix, No AmpErase UNG (Applied Biosystems Japan) with 400 nM primers (HQ-SEQ grade, Takara Bio, Inc.) and 250 nM TaqMan probe (Applied Biosystems Japan). The oligonucleotide sequences of the primers and the probe used for the PV and CV quantifications were taken from previous reports [23,24]. Amplification was conducted at 50 °C for 2 min, 95 °C for 10 min, and then 40 cycles of 95 °C for 15 s and 60 °C for 1 min in an Applied Biosystems 7300 Real-Time PCR System (Applied Biosystems Japan). The standard curve for the real-time PCR method was based on the relationship between the infectious PV or CV concentration of a freshly prepared purified PV or CV solution measured by the plaque assay described above, and the number of cycles (C_t value) of PCR amplification.

2.6. Aluminum species distribution analysis during the coagulation process

The effect of pH on the aluminum hydrolyte species transfer of coagulants during the coagulation process was analyzed mostly by following the method of Yan *et al.* [25]. Batch coagulation experiments were conducted with 1000 mL of prepared Milli-Q water containing 0.5 mM NaHCO₃ and NaNO₃ in square plastic beakers at 20 °C. After enough HCl or NaOH was added to the prepared Milli-Q water to bring the final pH to a target value of 6, 7, or 8, 2.7 mg-Al/L of coagulant was injected in the water. The water was stirred rapidly for 1 min ($G = 339 \text{ s}^{-1}$, 200 rpm), and then the solution was analyzed by using the ferron method.

3. Results and discussion

3.1. Characterization of coagulants

To investigate the influence of the type of aluminum hydrolyte species present in coagulants on virus removal, five

aluminum-based coagulants containing different aluminum hydrolyte species were prepared and used in this study. Application of the ferron method showed that the major aluminum species in AlCl₃ was a monomeric aluminum species (Al_a) (Table 2); ²⁷Al NMR spectroscopy confirmed that the major aluminum species in AlCl₃ was monomeric because only two peaks, (the aluminum monomer peak at 0 ppm and the internal standard substance peak at 80 ppm) were observed in the AlCl₃ spectrum (Fig. 1). In contrast, the major aluminum species in PACI-2.1b and PACI-2.4b was polymeric aluminum species (Al_b) (Table 2). Moreover, unlike AlCl₃, these PACIs contained Al₁₃ species as shown by the central tetrahedral Al peak in the Al₁₃ species at 63 ppm in the ²⁷Al NMR spectra of PACI-2.1b and PACI-2.4b (Fig. 1). In contrast, PACI-2.1c and PACI-2.4c contained a high proportion of colloidal aluminum species (Al_c) (Table 2), and these PACIs contained not only Al₁₃ species but also Al₃₀ species, which was not the case for PACI-2.1b and PACI-2.4b, because the broad peaks of the octahedral Al of the external shells in the Al₁₃ and Al₃₀ species at 10–12 ppm and the central tetrahedral Al in the Al₃₀ species at 70 ppm were observed in the ²⁷Al NMR spectra of PACI-2.1c and PACI-2.4c, in addition to the peak of the Al₁₃ species (Fig. 1). These results concur with those of other researchers who have reported that aluminum monomer can be classified as Al_a, Al₁₃ species can be classified as Al_b, and Al₃₀ species can be classified as Al_c [1,4,7].

3.2. Effect of aluminum species in coagulants on virus removal

The effect of aluminum species in the coagulants on the infectious PV removal ratio ($\log_{10}(C_0/C_5)$) during the coagulation process was evaluated by using the PFU method after settling (Fig. 2a). Because of the small size of PV, no PV removal (0.1- \log_{10}) was observed in the absence of coagulant at approximately pH 7. In contrast, the coagulation process removed infectious PV over the pH range of 6–8 no matter what type of coagulant was used. The addition of the coagulant destabilized the dispersion of the virus particles in the source water. The particles were then entrapped in or adsorbed onto the aluminum floc particles generated during coagulation, and then the floc particles with the entrapped or adsorbed viruses settled from suspension under the influence of gravity

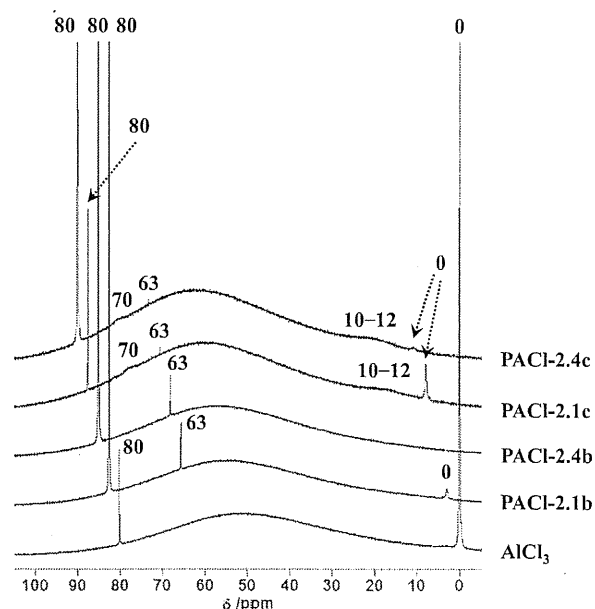


Fig. 1. ²⁷Al NMR spectra of the coagulants used in the present study.

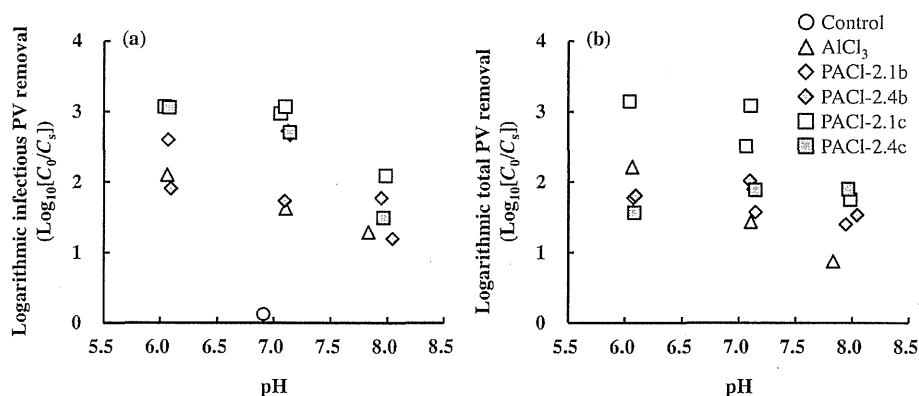


Fig. 2. Effect of aluminum hydrolyte species in coagulants on (a) infectious PV removal as evaluated by using the PFU method and on (b) total PV removal as evaluated by using the PCR method after settling during the coagulation process. Source water, lake water; coagulant dosage, 2.70 mg-Al/L.

during the settling process. The removal ratios of the infectious PV depended on the coagulant type: whereas coagulation with AlCl₃ and PACI-2.4b resulted in approximately 2-log₁₀ removal over a pH range of 6–7, approximately 3-log₁₀ removal was achieved with PACI-2.1b, PACI-2.1c, and PACI-2.4c. In addition, the removal ratios observed with PACI-2.1c were somewhat larger than those observed with the other aluminum-based coagulants, and approximately 2-log₁₀ removal was obtained with PACI-2.1c even at pH 8. Taken together, these data indicate that the ability of the coagulants to remove infectious PV from virus-spiked lake water followed the order: PACI-2.1c > PACI-2.4c ≥ PACI-2.1b > PACI-2.4b = AlCl₃. The removal ratios of total PV, evaluated by using the PCR method, were also observed to be somewhat larger with PACI-2.1c than those observed with the other aluminum-based coagulants tested (Fig. 2b). The ability of PACI-2.1c to remove infectious PV from virus-spiked river water was also higher than that of the other aluminum-based coagulants, in agreement with the result from the virus-spiked lake water experiments (Fig. 3a), although the removal ratios obtained with the lake water and the river water were different, due to differences in water quality (Table 1). Moreover, the total PV removal ratios were also found to be somewhat larger with PACI-2.1c than those with the other aluminum-based coagulants tested in the river water sample (Fig. 3b). These results indicate that differences in the distribution of the aluminum species in the coagulants affected virus removal performance during coagulation and that PACI-2.1c effectively removed PV, particularly at pH 6–7.

In the present study, similar removal ratios of PV were observed with AlCl₃ and PACI-2.4b, even though the original compositions of the aluminum species in these coagulants were quite different (Fig. 4, the analytical pH condition was approximately 4–5, these are the same data as those shown in Table 2). Because further transformation of aluminum species after coagulant dosing is known to occur, and the distribution of aluminum species depends largely on the original composition of the coagulants and the pH of the treated water [4,25,26], we investigated the effect of pH on the transformation of the aluminum species after coagulant dosing across a pH range of 6–8 (Fig. 4). The distribution of the aluminum species of the four PACIs during coagulation tended to be maintained (i.e., the original composition of the aluminum species was retained): the Al_b fraction, which represented a large proportion of PACI-2.1b and PACI-2.4b, and the Al_c fraction, which represented a large proportion of PACI-2.1c and PACI-2.4c, were almost unchanged between the pH ranges of 4–5 and 6–8 (Fig. 4), although the Al_c fraction, which represented a small proportion of PACI-2.1b and PACI-2.4b, was smaller when the pH was increased from 4–5 to 6. This result indicates that most of the

original Al_b and Al_c content of the PACIs was relatively stable over the pH range of 4–8, in agreement with previous studies [4,26]. In contrast, the distribution of the aluminum species of AlCl₃ changed markedly during coagulation compared with the original composition: the Al_a fraction, which represented a large proportion of AlCl₃ decreased and the Al_b fraction increased when the pH was increased from 4–5 to 6–8 (Fig. 4). This result indicates that a large proportion of the original Al_a content of AlCl₃ transformed into Al_b during coagulation, and that the formation of Al_b *in situ* probably led to the similar removal ratios of PV that were observed with PACI-2.4b.

The removal ratios of PV observed with high Al_c content PACIs tended to be larger than those observed with high Al_b content PACIs (Figs. 2 and 3). Because PACI-2.1c and PACI-2.4c contained not only Al₁₃ species but also Al₃₀ species, which was not the case for PACI-2.1b and PACI-2.4b, we hypothesize that the Al₃₀ species in PACI played a major role in PV removal during the coagulation process. These results, taken together with our previous findings [16], suggest that Al₃₀ species in aluminum-based coagulants are the important species to achieve efficient removal of not only bacteriophages but also human enteric viruses during the coagulation process.

When the basicity was increased from 2.1 to 2.4, the Al_c content in the high Al_c content PACIs increased slightly and the Al_a content in those PACIs decreased slightly (Fig. 4, the analytical pH condition was approximately 4–5). A similar trend was observed for the distribution of the aluminum species of PACIs during coagulation (Fig. 4, the analytical pH condition was approximately 6–8). However, the removal ratios of PV observed with PACI-2.4c tended to be lower than those with PACI-2.1c (Figs. 2 and 3). This result suggests that the coagulation efficiency of virus during the coagulation process is not determined simply by the amount of Al_c in the coagulant.

To elucidate why PACI-2.1c effectively removed PV, we used the colloid titration method to determine the positive colloid charges of the coagulants (Fig. 5). The colloid charge densities of AlCl₃ were almost zero and constant, regardless of the aluminum concentration. In contrast, the colloid charge densities of the four PACIs increased with increasing aluminum concentration. However, we observed no large differences in the colloid charge densities of the PACIs, whereas the removal ratios observed with PACI-2.1c were somewhat larger than those observed with the other aluminum-based coagulants, as described above. Therefore, the virus coagulation efficiency was not solely dependent on either the amount of Al_c in the coagulant or the positive colloid charge of the coagulant. Ye *et al.* [12] recently reported that the amount of Al₃₀ species in the coagulant increases with increasing basicity

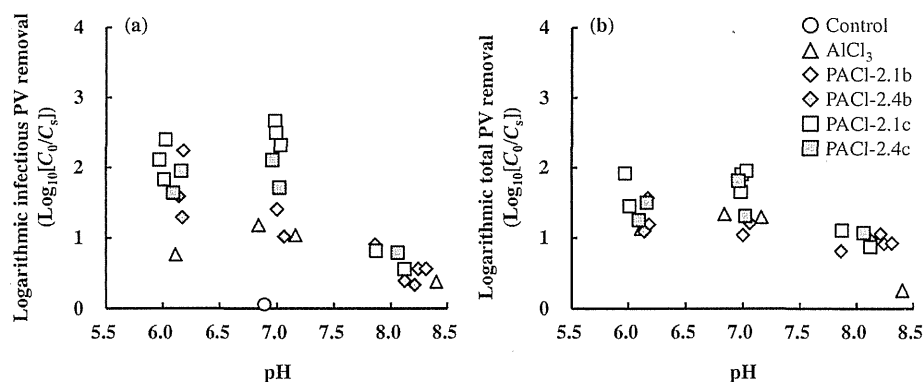


Fig. 3. Effect of aluminum hydrolyte species in coagulants on (a) infectious PV removal as evaluated by using the PFU method and on (b) total PV removal as evaluated by using the PCR method after settling during the coagulation process. Source water, river water; coagulant dosage, 1.08 mg-Al/L.

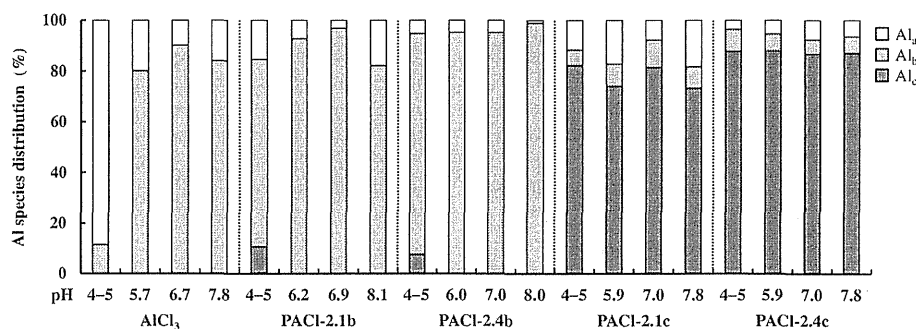


Fig. 4. Effect of pH on the transformation of aluminum species after coagulant dosing, as evaluated by using the ferron method.

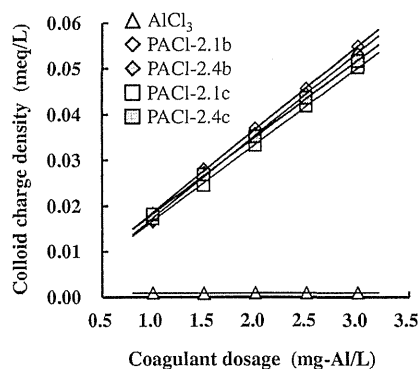


Fig. 5. Positive colloid charges of the coagulants used in the present study, as evaluated by using the colloid titration method.

from 1.0 to 2.0, but decreases when the basicity increases from 2.0 to 2.5. Therefore, the difference in the amount of Al₃₀ species between PACI-2.1c and PACI-2.4c due to the difference in basicity probably led to the difference in virus removal performance during coagulation. Further investigation is needed to determine why the removal ratios of PV observed with PACI-2.4c tended to be lower than those with PACI-2.1c, even though the Al_c content of PACI-2.4c was slightly higher than that of PACI-2.1c.

3.3. Comparison of PV and CV removal ratios during coagulation

As described above, PACI-2.1c removed PV more efficiently than did the other aluminum-based coagulants tested, especially at

approximately pH 7. To confirm that PACI-2.1c actually removed viruses more effectively than did AlCl₃ or PACI-2.1b, we also evaluated the removal ratios of a CV that has high resistance to free-chlorine disinfection relative to other human enteric viruses, including PV [21,22]. We evaluated the CV removal ratios by using the PFU and PCR methods after settling during the coagulation process, and then compared the results with those for PV (Fig. 6). For both enteroviruses, the removal ratios observed with PACI-2.1c were larger than those observed with AlCl₃ and PACI-2.1b at about pH 7: approximately 3-log₁₀ removals were achieved with PACI-2.1c. This result indicates that coagulation with PACI-2.1c more effectively removes human enteroviruses, including a virus with high resistance to free-chlorine disinfection, relative to coagulation with AlCl₃ and PACI-2.1b.

The removal ratios of infectious PV and CV as evaluated by use of the PFU method were almost identical during coagulation regardless of the coagulant type used (Fig. 6). A similar trend was observed for the removal ratios of total PV and CV, as evaluated by using the PCR method. In contrast, Mayer *et al.* [20] reported that the removal ratio of CV B6, evaluated by using the 50% tissue culture infectious dose (TCID₅₀) method, was consistently larger than that of PV type 1 evaluated by the same method during enhanced coagulation with FeCl₃; for example, the optimal enhanced coagulation conditions of 40 mg/L FeCl₃ and a pH of between 5.5 and 6.5 resulted in 1.0–2.5-log₁₀ and 1.5–3.0-log₁₀ removals of infectious PV and CV, respectively. This result is not consistent with our result, although we did observe similar ranges of removal ratios for infectious PV and CV. In general, the removal ratios of microorganisms including viruses during coagulation are influenced by several factors such as source water quality, coagulant type, coagulant dosage, and coagulation pH [27]. Several of these factors, in particular coagulant type (aluminum-based