

administered materials in toxicity studies is fundamental, and characterizing delivered nanomaterials after administration in a test system or model provides the best quality data on dose and material properties that are related to observed responses, but this is limited by current methodological capabilities [2]. Further studies, especially *in vivo*, using different types of characterized materials, relevant routes of administration, and doses closely reflecting expected levels of exposure are needed to adequately evaluate the reproductive and developmental toxicity of nanomaterials.

**Conflict of interest**

None.

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**References**

[1] Oberdörster G, Maynard A, Donaldson K, Castranova V, Fitzpatrick J, Ausman K, et al. Principles for characterizing the potential human health effects from exposure to nanoparticles: elements of a screening strategy. *Part Fibre Toxicol* 2005;2:8, doi:10.1186/1743-8977-2-8.

[2] Oberdörster G, Oberdörster E, Oberdörster J. Nanotoxicology: an emerging discipline evolution from studies of ultrafine particles. *Environ Health Perspect* 2005;113:823–39.

[3] Penn SG, He L, Natan MJ. Nanoparticles for bioanalysis. *Curr Opin Chem Biol* 2003;7:609–15.

[4] Liu WT. Nanoparticles and their biological and environmental applications. *J Biosci Bioeng* 2006;102:1–7.

[5] Colborn T, vom Saal F, Soto A. Developmental effects of endocrine-disrupting chemicals in wildlife and humans. *Environ Health Perspect* 1993;101:378–84.

[6] Brooking J, Davis SS, Illum L. Transport of nanoparticles across the rat mucosa. *J Drug Target* 2001;9:267–79.

[7] Wang J, Liu Y, Jiao F, Lao F, Li W, Gu Y, et al. Time-dependent translocation and potential impairment on central nervous system by intranasally instilled TiO<sub>2</sub> nanoparticles. *Toxicology* 2008;254:82–90.

[8] Lee KP, Trochimowicz HJ, Reinhardt CF. Pulmonary response of rats exposed to titanium dioxide (TiO<sub>2</sub>) by inhalation for two years. *Toxicol Appl Pharmacol* 1985;79:179–92.

[9] Lindenschmidt RC, Driscoll KE, Perkins MA, Higgins JM, Maurer JK, Belfiore KA. The comparison of a fibrogenic and two nonfibrogenic dusts by bronchoalveolar lavage. *Toxicol Appl Pharmacol* 1990;102:268–81.

[10] Muhle H, Bellmann B, Creutzenberg O, Dasenbrock C, Ernst H, Kilpper R, et al. Pulmonary response to toner upon chronic inhalation exposure in rats. *Fundam Appl Toxicol* 1991;17:280–99.

[11] Warheit B, Sayes CM, Reed KL, Swain KA. Health effects related to nanoparticle exposures: environmental, health and safety considerations for assessing hazards and risks. *Pharmacol Ther* 2008;120:35–42.

[12] Heinrich U, Fuhst R, Rittinghausen S, Cretzenberg O, Bellmann B, Koch W, et al. Chronic inhalation exposure of Wistar rats and two different strains of mice to diesel engine exhaust, carbon black, and titanium dioxide. *Inhal Toxicol* 1995;7:533–56.

[13] Baan R, Staif K, Grosse Y, Secretan B, Ghissassi F, Coglianò V, on behalf of the WHO International Agency for Research on Cancer Monograph Working Group. Carcinogenicity of carbon black, titanium dioxide, and talc; 2006 [cited June 22, 2009], available from: <http://oncology.thelancet.com>.

[14] Landsiedel R, Kapp MD, Schulz M, Wiench K, Oesch F. Genotoxicity investigations on nanomaterials: methods, preparation and characterization of test material, potential artifacts and limitations—many questions, some answers. *Mutat Res* 2009;681:241–58.

[15] Ema M, Kobayashi N, Naya M, Hanai S, Nakanishi J. Genotoxicity evaluation of titanium dioxide nanoparticles. *Jpn J Environ Toxicol* 2009;12:71–84 [in Japanese].

[16] Fedulov AV, Leme A, Yang Z, Dahl M, Lim R, Mariani TJ, Kobzik L. Pulmonary exposure to particles during pregnancy causes increased neonatal asthma susceptibility. *Am J Respir Mol Biol* 2008;38:57–67.

[17] CEPA/OEHHA (California Environmental Protection Agency/Office of Environmental Health Hazard Assessment). Safe drinking water and toxic enforcement ACT of 1986 (Position 65). Notice to interested parties, February 21, 2003. Chemical listed effective February 21, 2003 as known to the State of California to cause cancer; 2003 [cited October 13, 2009], available from: <http://oehha.ca.gov/prop65/prop65.list/files/22103not.pdf>.

[18] Shimizu M, Tainaka H, Oba T, Mizuo K, Umezawa M, Takeda K. Maternal exposure to nanoparticle titanium dioxide during the prenatal period alters gene expression related to brain development in the mouse. *Part Fibre Toxicol* 2009;6:20, doi:10.1186/1743-8977-6-20.

[19] Takeda K, Suzuki K, Ishihara A, Kubo-Irie M, Fujimoto R, Tabata M, et al. Nanoparticles transferred from pregnant mice to their offspring can damage the genital and cranial nerve system. *J Health Sci* 2009;55:95–102.

[20] Komatsu T, Tabata M, Kubo-Irie M, Shimizu T, Suzuki K, Nihei Y, et al. The effects of nanoparticles on mouse testis Leydig cells *in vitro*. *Toxicol in Vitro* 2008;22:1825–31.

[21] Shelley SA. Nanotechnology: turning basic science into reality. In: Theodore L, Kunz RG, editors. *Nanotechnology—environmental implications and solution*. Hoboken: John Wiley and Sons, Inc. Publication; 2005. p. 61–107.

[22] Challier JC, Panigel M, Meyer E. Uptake of colloidal <sup>198</sup>Au by fetal liver in rats, after direct intrafetal administration. *Int J Nucl Med Biol* 1973;1:103–6.

[23] Takahashi S, Matsuoka O. Cross placental transfer of <sup>198</sup>Au-colloid in near term rats. *J Radiat Res* 1981;22:242–9.

[24] Sadauskas E, Wallin H, Stoltenberg M, Vogel U, Doering P, Larsen A, et al. Kupper cells are central in the removal of nanoparticles from the organism. *Part Fibre Toxicol* 2007;4:10, doi:10.1186/1743-8977-r-4-10.

[25] Myllynen P, Loughran MJ, Howard CV, Sormunen R, Walsh A, Vähäkangas KH. Kinetics of gold nanoparticles in the human placenta. *Reprod Toxicol* 2008;26:130–7.

[26] Wiwantikitt V, Sereemasump A, Rojanathanes R. Effect of gold nanoparticles on spermatzoa: the first world report. *Fertil Steril* 2009;91:e7–8.

[27] Kim WS, Kim HC, Hong SH. Gas sensing properties of MoO<sub>3</sub> nanoparticles synthesized by solvothermal method. *J Nanopart Res* 2009, doi:10.1007/s11051-009-9751-6.

[28] Braydich-Stolle L, Hussain S, Schlager JJ, Hofmann MC. *In vitro* cytotoxicity of nanoparticles in mammalian germline stem cells. *Toxicol Sci* 2005;88:412–9.

[29] MakhluufSBD, Qasem R, Rubinstein S, Gedanken A, Breibart H. Loading magnetic nanoparticles into sperm cells does not affect their functionality. *Langmuir* 2006;22:9480–2.

[30] Bhabra G, Sood A, Fisher B, Cartwright L, Saunders M, Evans WH, et al. Nanoparticles can cause DNA damage across a cellular barrier. *Nat Nanotechnol* 2009;4:876–83.

[31] Lin YS, Haynes CL. Impact of mesoporous silica nanoparticle size, pore ordering, and pore integrity on hemolytic activity. *J Am Chem Soc* 2010;132:4834–42.

[32] Park MVDZ, Annema W, Salvati A, Lesniak A, Elsaesser A, Barnes C, et al. *In vitro* developmental toxicity test detects inhibition of stem cell differentiation by silica nanoparticles. *Toxicol Appl Pharmacol* 2009;240:108–16.

[33] Loutfy RO, Lowe TP, Moravsky AP, Katagiri S. Commercial production of fullerenes and carbon nanoparticles. In: *Perspectives of fullerene nanotechnology*. Netherlands: Springer; 2002. p. 35–46.

[34] Tsuchiya T, Oguri I, Yamakoshi YN, Miyata N. Novel harmful effects of [60]fullerene on mouse embryos *in vitro* and *in vivo*. *FEBS Lett* 1996;393:139–45.

[35] Sumner SCJ, Fennell TR, Snyder RW, Taylor GF, Lewin AH. Distribution of carbon-14 labeled C<sub>60</sub>([<sup>14</sup>C]C<sub>60</sub>) in the pregnant and in the lactating dam and the effect of C<sub>60</sub> exposure on the biochemical profile of urine. *J Appl Toxicol* 2009, doi:10.1002/jat.1503.

[36] Donaldson K, Tran L, Jimenez LA, Duffin R, Newby DE, Mills N, et al. Combustion-derived nanoparticles: a review of their toxicology following inhalation exposure. *Part Fibre Toxicol* 2005;2:10, doi:10.1186/1743-8977-2-10.

[37] Yoshida S, Hiyoshi K, Ichinose T, Takano H, Oshio S, Sugawara I, et al. Effect of nanoparticles on the male reproductive system of mice. *Int J Androl* 2008;32:337–42.

[38] Chan WH, Shiao NH. Cytotoxic effect of CdSe quantum dots on mouse embryonic development. *Acta Pharmacol Sin* 2008;28:259–66.

[39] Derfus AM, Chan WCW, Bhatia SN. Probing the cytotoxicity of semiconductor quantum dots. *Nano Lett* 2004;4:11–8.

[40] Bangs LB. New developments in particle-based immunoassays: introduction. *Pure Appl Chem* 1996;68:1873–9.

[41] Bosman SJ, Niet SP, Patton WC, Jacobson JD, Corselli JU, Chan PJ. Development of mammalian embryos exposed to mixed-size nanoparticles. *Clin Exp Obst Gyn* 2005;32:222–4.

[42] Shepard TH, Lemire RJ. *Catalog of teratogenic agents*. 11th ed. Baltimore: The Johns Hopkins University Press; 2004.

[43] Nielsen E, Thorup I, Schnipper A, Hass U, Meyer O, Ladefoged J, et al. Children and the unborn child. Exposure and susceptibility to chemical substances—an evaluation. Environmental project No. 589. Miljøprojekt; 2001 [cited October 18, 2009], available from: <http://www2.mst.dk/udgiv/Publications/2001/87-7909-574-7/pdf/87-7909-333-7.pdf>.

[44] Klimisch HJ, Andreae M, Tillmann U. A systemic approach for evaluating the quality of experimental toxicological and ecotoxicological data. *Regl Toxicol Pharmacol* 1997;25:1–5.

[45] Zhu X, Zhu L, Li Y, Duan Z, Chen W, Alvarez PJJ. Developmental toxicity in zebrafish (*Danio rerio*) embryos after exposure to manufactured nanomaterials: buckminsterfullerene aggregates (nC<sub>60</sub>) and fullerol. *Environ Toxicol Chem* 2007;26:976–9.

# ラットを用いた短期反復投与毒性試験の低用量群に統計学的有意差が検出される割合

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## 1 要約

反復投与毒性試験の試験責任者は、低用量を無影響量として設定する。しかし、低用量群に統計学的有意差が散見される毒性試験は、少なくない。化審法によるラットを用いた109の28日間反復投与毒性試験を調査し、低用量群に統計学的有意差 ( $P < 0.05$ ) が認められる数とその割合を調査した。その結果、低用量群には205/12167 (1.5%) の有意差が認められた。次に検査項目別では、尿検査が3.3%と高く、その他の検査項目の血液学的検査、血液生化学的検査、器官重量および器官重量・体重比は、1.1~1.8%程度であった。以上の調査結果、我々は低用量群の有意差検出率が2% (最大<5%)程度であれば無毒性量を評価できると判断する。

## 2 はじめに

げっ歯類などを用いた毒性試験は、医薬品、農薬、動物用医薬品および一般化学物質についてそれぞれの所轄の官庁が定めたガイドラインによって実施されている。これら毒性試験の目的は、無影響量 (NOEL) または無毒性量 (NOAEL) の把握である。試験責任者は、用量設定予備試験の結果から本試験の各用量を決定する。そ

して本試験では、低用量群を無影響量と推定し試験を開始する。本試験終了後、各測定項目について統計学および毒性学的有意差から無毒性量を決定する。しかし、これら毒性試験では、低用量群に統計学的有意差が認められる測定値が散見される。この場合は、背景値との比較、用量依存性の有無および関連項目との整合性などを吟味して無毒性量などを決定する。本調査報告の目的は、既存化学物質に対するラットを用いた短期28日間反復投与毒性試験から低用量群にどの程度の統計学的有意差 ( $P < 0.05$ ) が検出されているかを調査し、若干の考察を加え、試験責任者の知見にしていだきたい。

## 3 調査材料および方法

化審法ガイドライン (NITE, 2007) に従ったラットを用いた109の28日間反復投与毒性試験をインターネット (厚生労働省, 2009) から取得した。調査した化審法28日反復投与毒性試験の基本的群構成を表1に示した。主な解析データは、投与28日後の定量値を用いた。各群の構成は、対照群を含めて4および5群 (37試験) であった。多くの試験の1群内動物数は、5匹程度、用量の公比は、3が最も多かった。被験物質の投与は、全て胃ゾンデによる強制経口投与であった。

表1 試験群の構成と供試動物数

群	供試動物数	
	28日間の投与期間	14日間の回復期間
対照	10	5
低用量	5	—
中用量	5	—
高用量	10	5

調査に用いた検査項目（定量値）は、表2に示した。体重、飼料摂取量および飲水量などのデータはグラフによる開示のみで、これらのデータは、もし統計学的有意差が認められても経時的変化によって考察ができることから除外した。また尿検査および病理解剖・組織学的検査成績の定性データは、例数が少ないことから統計学的分析に問題があることから除外した。

表2 調査に用いた主な検査項目（定量値）

検査項目	主な測定項目
行動機能観察 (FOB)	握力、閉鎖極および自発運動量
尿検査	尿量、比重、浸透圧、Na, K, Cl
血液学的検査	白血球数およびその分画比、赤血球数、ヘモグロビン量、ヘマトクリット値、平均赤血球容積、平均赤血球色素量、平均赤血球色素濃度、血小板数および白血球百分率、プロトロンビン時間、活性化部分トロンボプラスチン時間など
血液生化学的検査	総蛋白濃度、総コレステロール濃度、ブドウ糖濃度、尿酸窒素濃度、クレアチニン濃度、アルカリフォスファターゼ活性、GOT活性、GPT活性、 $\gamma$ -GTP活性、トリグリセライド濃度、無機リン濃度、カルシウム濃度、A/G、ナトリウム濃度、カリウム濃度および塩素濃度など 一部の試験は骨髄検査および電気泳動による蛋白分画検査を実施している。
器官重量および体重に対する相対重量比	脳、甲状腺、胸腺、心臓、肺、肝臓、脾臓、腎臓、副腎、精巣、卵巣など

表3に統計解析に使用した主な解析手法（Kobayashi *et al.*, 2008）を示した。調査した報告書は、片側検定、両側検定および記載なしが混在していた。この解析によって有意差（ $P < 0.05$ ）が認められた検査項目を用量毎に集計した。

表3 げっ歯類を用いた28日間の反復投与毒性試験に使用された統計解析ツール

Bartlettの等分散検定、分散分析、Dunnettの多重比較検定、Schefféの多重比較検定、Duncanの多重範囲検定、Kruskal-Wallisの検定、Dunnett型ノンパラメトリック検定、Scheffé型ノンパラメトリック検定、Steelの多重比較検定など
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## 4 調査結果および考察

調査試験109報ごとの対照群に対する低用量、中用量、高用量および最高用量群の5%水準による有意差検出数とその割合を表4に示した。1試験の調査項目は、平均111であった。対照群に対する有意差検出数とその割合は、低用量群、中用量群、高用量群および最高用量群がそれぞれ1.6、3.4、10および17%と当たり前であるが用量依存性が顕著に認められた。無影響量として設定した低用量群は、最大8.4%、最小0.0%、最頻値0.0%で平均1.6%の有意差が認められた。試験によって大きなばらつきが認められた。低用量群に全く統計学的有意差が認められなかった試験は、30/109であった。

次に109試験の調査項目別に対照群に対する各用量群の統計学的有意差数およびその割合を表5に示した。表下段の合計は、表4の合計と一致する。検査項目中、尿検査は、他の検査項目に比較して群間を通して有意差が検出されやすいことが認められた。尿検査以外の検査項目の血液学的検査、血液生化学的検査、器官重量および器官重量・体重比は、各群内で大きな差が無かった。

定期検査・解剖などが投与開始後26、52、78および104週で設定されている長期の慢性毒性試験など大規模試験は、極めて多くの測定値が得られる。このため低用量群には、偶発的な統計学的有意差が検出されやすい。低用量群に統計学的有意差が認められた場合の無毒性量は、毒性学的有意差の有無を検討し設定したい。以上の調査結果から、我々は、全調査項目中、低用量群に統計学的有意差が1~2%（ $< 5\%$ ）程度認められても無毒性量を評価できると判断する。

表4 化審法28日間反復投与毒性試験から得られた定量値に対する有意差 ( $P < 0.05$ ) 検出数およびその割合 (その1, 試験別からの結果)

試験番号	測定項目数	有意差検出数および (%)			
		低用量群	中用量群	高用量群	最高用量群
1	120	0 (0.0)	2 (1.7)	4 (3.3)	10 (11)
2	104	6 (5.7)	5 (4.8)	4 (3.8)	—
3	108	1 (0.9)	2 (1.8)	3 (2.7)	—
4	128	2 (1.5)	10 (7.8)	23 (18)	—
5	98	4 (4.8)	10 (10)	17 (17)	—
6	120	0 (0.0)	4 (3.3)	10 (8.3)	24 (20)
7	146	2 (1.3)	6 (4.1)	42 (28)	—
8	118	1 (0.8)	1 (0.8)	0 (0.0)	1 (0.8)
9	118	1 (0.8)	2 (1.6)	27 (22)	—
10	116	1 (0.8)	10 (8.6)	26 (22)	—
11	116	2 (1.7)	10 (8.6)	26 (22)	—
12	156	3 (1.9)	11 (7.0)	36 (23)	—
13	106	2 (1.8)	5 (4.7)	30 (28)	—
14	100	7 (7.0)	11 (11)	23 (23)	—
15	100	2 (2.0)	2 (2.0)	2 (2.0)	—
16	128	3 (2.3)	10 (7.8)	44 (34)	—
17	96	5 (5.2)	0 (0.0)	4 (4.1)	—
18	118	2 (1.6)	0 (0.0)	1 (0.8)	4 (3.3)
19	114	0 (0.0)	1 (0.8)	1 (0.8)	—
20	102	4 (3.9)	1 (0.9)	6 (5.8)	—
21	94	3 (3.1)	6 (6.3)	2 (2.1)	—
22	128	0 (0.0)	2 (1.5)	23 (17)	—
23	136	3 (2.2)	7 (5.1)	11 (8.0)	—
24	76	1 (1.3)	2 (2.6)	5 (6.5)	21 (27)
25	110	2 (1.8)	5 (4.5)	17 (15)	—
26	118	0 (0.0)	0 (0.0)	0 (0.0)	4 (3.3)
27	116	0 (0.0)	2 (1.7)	9 (7.7)	23 (19)
28	100	2 (2.0)	4 (4.0)	37 (37)	—
29	158	5 (3.1)	12 (7.5)	31 (19)	—
30	120	0 (0.0)	2 (1.6)	10 (8.3)	—
31	134	0 (0.0)	0 (0.0)	4 (2.9)	16 (11)
32	126	1 (0.7)	6 (4.7)	12 (9.5)	30 (23)
33	106	3 (2.8)	1 (0.9)	23 (21)	—
34	128	0 (0.0)	3 (2.3)	3 (2.3)	38 (29)
35	118	0 (0.0)	9 (7.6)	24 (20)	40 (33)
36	54	0 (0.0)	2 (3.7)	9 (16)	—
37	106	0 (0.0)	2 (1.8)	6 (5.6)	—
38	122	1 (0.8)	5 (4.0)	44 (36)	—
39	100	6 (6.0)	17 (17)	27 (27)	—
40	104	1 (0.9)	3 (2.8)	7 (6.7)	—
41	104	2 (1.9)	4 (3.8)	9 (8.6)	—
42	106	3 (2.8)	1 (0.9)	25 (23)	—
43	151	2 (1.3)	9 (2.9)	31 (20)	—
44	106	2 (1.8)	9 (8.4)	3 (2.8)	—
45	112	3 (2.6)	9 (8.0)	18 (16)	—
46	86	0 (0.0)	5 (5.8)	14 (16)	30 (34)
47	104	2 (1.9)	0 (0.0)	7 (6.7)	—
48	70	0 (0.0)	1 (1.4)	0 (0.0)	—
49	50	0 (0.0)	2 (4.0)	9 (18)	30 (60)
50	74	0 (0.0)	1 (1.3)	0 (0.0)	—
51	132	3 (2.2)	5 (3.7)	9 (6.8)	—
52	142	12 (8.4)	11 (7.7)	7 (4.9)	—
53	94	1 (1.0)	2 (2.1)	2 (2.1)	—
54	116	5 (4.3)	7 (6.0)	3 (2.5)	—
55	98	7 (7.1)	3 (3.0)	23 (23)	—
56	104	6 (5.7)	12 (11)	21 (20)	—
57	88	2 (2.2)	1 (1.1)	6 (6.8)	—
58	98	0 (0.0)	0 (0.0)	8 (8.1)	—

試験番号	測定項目数	有意差検出数および (%)			
		低用量群	中用量群	高用量群	最高用量群
59	100	0 (0.0)	0 (0.0)	1 (1.0)	8 (8.0)
60	104	2 (1.9)	8 (7.6)	37 (35)	58 (55)
61	96	1 (1.0)	7 (7.2)	16 (16)	—
62	128	1 (0.7)	3 (2.3)	25 (19)	—
63	138	2 (1.4)	6 (4.3)	49 (35)	—
64	128	1 (0.7)	1 (0.7)	12 (9.3)	—
65	130	0 (0.0)	3 (2.3)	9 (6.9)	33 (25)
66	104	0 (0.0)	6 (5.7)	42 (40)	—
67	94	6 (6.3)	3 (3.1)	4 (4.2)	—
68	106	3 (1.8)	3 (1.8)	3 (1.8)	—
69	110	0 (0.0)	0 (0.0)	2 (1.8)	21 (19)
70	122	0 (0.0)	2 (1.6)	4 (3.2)	0 (0.0)
71	162	1 (0.6)	2 (1.2)	9 (5.5)	—
72	136	1 (0.7)	3 (2.2)	18 (13)	—
73	104	1 (0.9)	0 (0.0)	1 (0.8)	4 (3.8)
74	106	0 (0.0)	0 (0.0)	7 (6.6)	19 (25)
75	112	7 (6.2)	1 (0.8)	1 (0.8)	—
76	118	1 (0.8)	3 (2.5)	6 (5.0)	13 (11)
77	90	0 (0.0)	0 (0.0)	2 (2.2)	14 (15)
78	116	0 (0.0)	4 (3.4)	2 (1.7)	5 (4.3)
79	106	1 (0.9)	0 (0.0)	4 (3.7)	8 (7.5)
80	62	2 (3.2)	2 (3.2)	2 (3.2)	3 (4.8)
81	96	1 (1.0)	1 (1.0)	0 (0.0)	—
82	124	3 (2.4)	3 (2.4)	30 (24)	53 (42)
83	110	1 (0.9)	3 (2.7)	13 (11)	—
84	100	1 (1.0)	1 (1.0)	2 (2.0)	—
85	102	1 (0.9)	2 (1.9)	2 (1.9)	14 (13)
86	120	4 (3.3)	12 (10)	15 (12)	29 (24)
87	162	1 (0.6)	4 (2.4)	16 (9.8)	—
88	134	3 (2.2)	1 (0.7)	3 (2.2)	17 (12)
89	90	1 (1.1)	2 (2.2)	2 (2.2)	—
90	108	1 (0.9)	2 (1.8)	1 (0.9)	—
91	100	2 (2.0)	2 (2.0)	4 (4.0)	—
92	92	0 (0.0)	0 (0.0)	0 (0.0)	2 (2.1)
93	110	2 (1.8)	5 (4.5)	21 (19)	33 (30)
94	100	0 (0.0)	0 (0.0)	1 (1.0)	—
95	104	1 (0.9)	1 (0.9)	1 (0.9)	—
96	104	5 (4.8)	11 (10)	22 (21)	—
97	98	4 (4.0)	5 (5.1)	19 (19)	—
98	114	0 (0.0)	0 (0.0)	8 (7.0)	43 (37)
99	118	3 (2.5)	3 (2.5)	0 (0.0)	—
100	88	1 (1.1)	1 (1.1)	1 (1.1)	—
101	118	3 (2.5)	3 (2.5)	12 (10)	—
102	120	1 (0.8)	0 (0.0)	3 (2.5)	18 (15)
103	116	0 (0.0)	1 (0.8)	1 (0.8)	14 (12)
104	150	1 (0.6)	6 (4.0)	15 (10)	—
105	134	5 (3.7)	7 (5.2)	14 (10)	—
106	100	0 (0.0)	0 (0.0)	6 (6.0)	11 (11)
107	136	1 (0.7)	1 (0.7)	4 (2.9)	—
108	130	2 (1.5)	4 (3.0)	16 (12)	31 (23)
109	116	1 (0.8)	1 (0.8)	2 (1.7)	9 (7.7)
試験数	109	109	109	109	37
合計	12167	205	414	1318	731/4074
平均値	111	1.60%	3.40%	10%	17%
最大値	162	8.40%	17%	36%	55%
最小値	54	0.00%	0.00%	0.00%	0.00%
最頻値	—	0.00%	—	—	—

表5 化審法28日間反復投与毒性試験から得られた定量値に対する有意差 ( $P < 0.05$ ) 検出数およびその割合 (その2, 測定項目からの結果)

測定項目	測定項目数	有意差検出数および (%)			
		低用量群	中用量群	高用量群	最高用量群
FOB	68	1 (1.4)	3 (4.4)	8 (11)	1/10 (10)
尿検査	392	13 (3.3)	30 (7.6)	81 (20)	40/96 (40)
血液学的検査	3586	56 (1.5)	106 (2.9)	318 (8.8)	176/1198 (14)
血液生化学的検査	4285	79 (1.8)	163 (3.8)	455 (10)	267/1426 (18)
器官重量	1928	22 (1.1)	44 (2.2)	188 (9.7)	103/672 (15)
器官重量・体重比	1908	34 (1.7)	68 (3.5)	268 (14)	144/672 (21)
合計	12167	205 (1.6)	414 (3.4)	1318 (10)	731/4074 (17)

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### 引用論文および資料

- 1) 厚生労働省 (2009) :  
[http://dra4.nihs.go.jp/mhlw\\_data/jsp/SearchPage.jsp](http://dra4.nihs.go.jp/mhlw_data/jsp/SearchPage.jsp)
- 2) Katsumi Kobayashi, K. Sadasivan Pillai, Yuki Sakuratani, Takemaru Abe, Eiichi Kamata and Makoto Hayashi (2008) : "Evaluation of statistical tools used in short-term repeated dose administration toxicity studies with rodents" : *J. Toxicol. Sci.*, Vol. 33 : No. 1, 97-104
- 3) NITE (2007) :  
<http://www.safe.nite.go.jp/kasinn/pdf/28test.pdf>

# 毒性試験から得られる定量値に対する有効数値の桁数の差違および Mann-Whitney の *U* 検定と Wilcoxon の検定の有意差検出の違い

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抄録：12 受託試験機関による 106 のラットを用いた 28 日反復投与毒性試験（化審法）から得られた定量値の有効桁数を検討するため心臓の実重量について調査した。定量データの有効値の桁数の設定は、各試験機関で異なることがわかった。多くの試験機関は、3 桁が最も多く次いで 5, 4 桁の順であった。次に、分散などを用いない順位和検定の Mann-Whitney の *U* 検定を用いて桁数の違いによる有意差検出パターンを検索した。桁数が 3 と小さい場合は、桁数 4 および 5 に比較して有意差が検出しにくいことがわかった。この理由は、桁数が小さいと同一順位が増加する傾向を示すことにある。*t*-検定系は、桁数に大きく影響されない。動物数が 30～40 程度の場合、統計数値表を利用する Mann-Whitney の *U* 検定と計算による Wilcoxon の検定は、ほぼ同様の検出力を示す。したがって、短期毒性試験の場合、もし順位和検定を採用する場合は、片側検定の Mann-Whitney の *U* 検定または Wilcoxon の検定を用い、平均値および標準偏差の表示は、有効桁数を 3 で表示し、計算値（生データ）は、4 または 5 桁を採用したい。

## 1 はじめに

げっ歯類を用いた反復投与毒性試験から得られる定量データは、体重および飼料摂取量などの飼育管理データ、赤血球数およびヘモグロビン量などの血液学的検査、AST・ALT 活性値および総蛋白濃度などの血液生化学的検査、尿比重および尿量などの尿検査および肝および腎などの器官重量検査などその全項目は、50 以上 (Kobayashi, 1997) である。これらの解析法のほとんどは、決定樹 (Kobayashi *et al.*, 2008) によって対照群と用量群間差を吟味している。解析途中で、全群間に等分散が認められない場合は、順位和検定 (rank sum test) となる。また比率・グレード値およびはじめから分散などの分布を利用しない順位和検定を採用する場合も少なくない。

このノンパラメトリック検定の順位和検定は、群間差を全群の順位の違いで解析している。群間に同一値が存在すると検出力は、低下する。したがって、個体値の桁数の設定によって有意差検出が変化することが推測できる。有効数値の桁数が多いと同一順位は、極めて少なくなると推測できる。

そこで我々は、化審法による公開試験報告書 109 について定量値の有効数値の桁数を調査し、有意差検出

パターンを調査した。同時に検出力の差を Student の  $t$ -検定, Mann-Whitney の  $U$  検定および Wilcoxon の検定と比較検討した。

## 2 調査材料および結果

### 2.1 有効数値の桁数

化審法 (NITE, 2007) 対応による 12 試験機関で実施されたラットを用いた 28 日間反復投与毒性試験 (MHLW, 2009) 149 試験をインターネットから検索し、その内、試験機関によって測定値の桁数が 3~5 と大きく異なる心実重量 (雄) を取得し、その心実重量個体値を試験機関別に、その有効数値の桁数を調査した (表 1)。

表 1 公開化審法 28 日間反復投与毒性試験から得られた心重量の試験機関別有効数値の桁数

試験機関名	調査試験数	心重量測定試験数	有効数値の桁数
A	24	7	3
B	19	9	3
C	18	13	3
D	25	17	5 および 4 (2 試験)
E	15	15	3
F	13	10	3
G	14	14	3
H	10	10	3
I	4	4	3
J	4	4	4
K	2	2	4
L	1	1	4
合計	149	106	

表 2 有効数値の桁数別公開試験数

有効数値の桁数	試験数 (%)
3	82 (77)
4	9 (8.5)
5	15 (14)
合計	106 (100)

その結果、149 の調査試験の中で心実重量を測定している試験数は 106 であった。各試験機関は、決まった有効数値の桁数が設定されていた。その桁数は、3 が最も多く全試験の 77%、次いで 5 桁が 14%、4 桁が 8.5% であった (表 1 および表 2)。

心実重量は、雄が 1 g 以上で、この場合の有効数値の桁数の表示が 3 桁 (例 1.12 g) であるが、雌は 1 g 以下で、この場合の有効数値の桁数の表示が 2 桁 (例 0.91 g) が殆どであった。すなわち、雄と雌で有効数値の桁数が異なる場合が多かった。

### 2.2 有効数値の桁数の違いによる統計学的有意差の差違

調査試験報告書の中で高用量群に有意差が認められた肝実重量 (g) の 1 試験を抽出し、有効桁数の違いによる有意差検出パターンを調査した。桁数を 5, 4 および 3 に設定し、桁数の変化に影響がないパラメトリック検定の Student  $t$ -検定と同一順位があると有意差が変化する Mann-Whitney の  $U$  検定と比較検討した。Mann-Whitney の  $U$  検定を選択した理由は、各群内動物数が 6 であることによる。両者の解析は、Excel 2008 および AOKI (2010) を用いて片側検定を採用した。

その結果は、表 3 に示した。このデータは、対照群に対して高用量群は、約 12% 重量の増加を示している。Student の  $t$ -検定は、全桁数とも 5% 水準で有意差を示した。しかし、Mann-Whitney の  $U$  検定は、同一順位が各群に 1 つ以上存在すると有意差が認められない。したがって、有効桁数を多く設定した場合は、同一順位の発生が少ないことから検出力が高くなる。

表3 化審法の28日間反復投与毒性試験から得られた雄の肝重量指数の違いによる有意差検出パターン

有効桁数	群		確率 (P)	
	対照 (N = 6)	高用量 (N = 6)	Student t-test	Mann-Whitney U test
5	10.391, 11.442, 13.653, 10.224, 10.783, 10.414	13.194, 11.444, 13.701, 11.572, 12.683, 12.661	0.0279	< 0.05 (U = 5)
平均値±標準偏差	11.151 ± 1.301	12.543 ± 0.889		
平均順位	4.3	8.6		
4	10.39, 11.44, 13.65, 10.22, 10.78, 10.41	13.19, 11.44, 13.70, 11.57, 12.68, 12.66	0.0279	有意差なし (U = 5.5)
平均値±標準偏差	11.14 ± 1.30	12.54 ± 0.88		
平均順位	4.4	8.5		
3	10.4, 11.4, 13.7, 10.2, 10.8, 10.4	13.2, 11.4, 13.7, 11.6, 12.7, 12.7	0.0286	有意差なし (U = 6)
平均値±標準偏差	11.1 ± 1.3	12.5 ± 0.8		
平均順位	4.5	8.5		

### 2.3 t-検定, Mann-WhitneyのU検定およびWilcoxonの検定の検出力

一般的に2群間検定の場合、F-testの結果、不等分散の場合は、Welchの検定となる。または始めからMann-WhitneyのU検定を採用する場合がある。Mann-WhitneyのU検定は、一般的に標本数が50程度までは統計数値表(AOKI, 2010)が用意されている。標本数がそれ以上の場合は、正規化検定を計算式からZ値を算出して有意差の判断を行う。この検定法は、Wilcoxonの検定と呼ばれる。したがって、標本数が35の場合は、Mann-WhitneyのU検定かWilcoxonの検定かその使用に迷う。すなわち、動物数が小数例の場合は、統計数値表(簡易表)によるMann-WhitneyのU検定で、

また大多数例は、計算によるWilcoxonの検定で解析するといわれている。

表4に標本数38のデータを示し種々の解析法による検出力の比較を示した。データは、38試験のFibrinogen量の雄雌別変動係数(%)である。解析プログラムは、Excel 2008およびAOKI(2010)を用いた。その結果、片側検定によるStudent t-検定とMann-WhitneyのU検定とWilcoxonの検定は、同一の検出力を示した。

表4 Fibrinogen量の変動係数に対するStudentのt-検定, Mann-WhitneyのU検定およびWilcoxonの検定の検出力(P)

性	N	平均値±標準偏差	F-test	Student t-test	Mann-Whitney U test		Wilcoxon test	
					Two-side	One-side	Two-side	One-side
雄	38	6.28 ± 2.42	P = 0.1030	P = 0.0117	Not slg.	P < 0.05	P = 0.0608	P = 0.0304
雌	38	7.78 ± 3.16			U = 541.5		Z = 1.8752	



### 3 考察

受託試験機関 12 による 106 の反復投与毒性試験から得られた定量値の有効桁数を調査するため心実重量について調査した。定量データの有効値の桁数の設定は、各試験機関で異なることがわかった。多くの試験機関は、3 桁が多かった。次に、分散などを用いない順位和検定の Mann-Whitney の *U* 検定を用いて桁数の違いによる有意差検出パターンを検索した。桁数が 3 と小さい場合は、桁数 4 および 5 に比較して有意差が検出しにくいことがわかった。この理由は、桁数が小さくなると同一順位が増加する傾向することである。しかし、*t*-検定系は桁数に大きく影響されない。

動物数が 30～40 程度の場合、統計数値表を利用する Mann-Whitney の *U* 検定と計算による Wilcoxon の検定は、ほぼ同様の検出力を示す。したがって、短期毒性試験の場合、もし順位和検定を採用する場合は、片側検定の Mann-Whitney の *U* 検定または Wilcoxon の検定を用い、平均値の表示は、有効桁数を 3 で表示し、計算値（生データ）は、4 または 5 桁を採用したい。

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#### 引用論文および資料

- ・大阪大学大学院薬学研究科・医薬情報解析学分野および大阪大学遺伝情報実験センター、大阪大学微生物病研究所（MEPHAS）（2010）：  
<http://www.gen-info.osaka-u.ac.jp/testdocs/tomocom/wilc1.html>
- ・AOKI (2010)：  
<http://aoki2.sl.gunma-u.ac.jp/lecture/Average/u-tab.html>
- ・Kobayashi, K. (1997) : A comparison of one- and two-sided tests for judging significant differences in quantitative data obtained in toxicological bioassay of laboratory animals, *Journal of Occupational Health*,

39, 29-35

- ・Kobayashi, K., Pillai, K. S., Sakuratani, Y., Abe, T., Kamata, E. and Hayashi, M. (2008) : Evaluation and statistical tools used in short-term repeated dose administration toxicity studies with rodent, *Journal of Toxicological Science*, 33, (1), 97-104
- ・NITE (2007) :  
<http://www.safe.nite.go.jp/kasinn/pdf/28test.pdf>
- ・Ministry of Health, Labour and Welfare, (MHLW) (2009) :  
[http://dra4.nihs.go.jp/mhlw\\_data/jsp/SearchPage.jsp](http://dra4.nihs.go.jp/mhlw_data/jsp/SearchPage.jsp)

Original Article

Relation between statistics and treatment-related changes  
obtained from toxicity studies in rats: if detected a  
significant difference in low or middle dose for quantitative  
values, this change is considered as incidental change?

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**ABSTRACT** — The purpose of a toxicity test is to determine the no-observed-effect level (NOEL) of test substance through biological and pharmacological techniques. If the low dose not does show statistically significant and biologically relevant changes in the data evaluated in a study, the usual practice is to consider this dose as the NOEL. To overcome this, 6 types of techniques that seemed to be appropriate are presented in this paper by investigating the results of several domestic and foreign theses on toxicology. The most appropriate techniques appear to be the trend test, comparison between treatment group and historical control by *t*-test, and confirmation that all individual values lie within the 95% confidence interval (2 SD) of the historical control value, if a significant difference is admitted in the low dose.

**Key words:** Toxicity, Rodents, Statistics, Historical control data, Incidental change, Standard deviation and error

INTRODUCTION

Toxicity test is necessary for evaluating the safety of industrial chemicals according to the Chemical Substances Control Law (1986). Quantitative data obtained from toxicity studies with test substance in rodents are analyzed by using decision tree procedure (Hamada *et al.*, 1998; Kobayashi *et al.*, 2008). If statistical analysis of such data reveals that low or mid dose data are significantly different as compared to the control group used as a reference for that particular study, it may not be possible to determine the no-observed-effect level (NOEL) of that test substance. This significant difference obtained in the low and mid dose groups is usually not considered as incidental. This can be confirmed by comparing the data of these groups with the historical control data obtained from the testing facilities or by confirming that there was no dose-related pattern observed, statistically or visually. Usually the comparison with the historical control data is

made by checking the data of the treatment groups (low and mid dose), whether they lie within the width of 2 standard deviations (S.D.) of the historical control data. If the data of the treatment groups lie within the width of 2 S.D. of the historical control, it is assumed that the changes shown by these groups are incidental. However, the S.D. plotted shows the distribution pattern of all the individual historical control data, whereas the data of the treatment group is obtained by calculating the mean of 5 to 35 animals used in an experiment. Hence, the comparison between these 2 data sets may be erroneous since the quality of data of both the groups is different. In order to resolve these disagreements and to confirm incidental finding, this paper presents six alternative techniques. We would have liked to have documented the uniformity of each testing facility in the form of a thesis; however, the documentation of this kind is problematic since there are restrictions in publishing the toxicity results of a non-government organization and since the investigation

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lie within the width of  $\pm 2$  S.D. of the historical control value (Fig. 1).

(3) Also the difference between the mean value of the historical control group and that of the treatment group is analyzed by the *t*-test (Table 1).

#### The NOEL judgments

The NOEL is judged by the study director on the basis of his/her knowledge and experience in addition to the data presented in the thesis. In this case, priority may be given to the clinical relevance of the data as compared to the statistical significant difference. For instance, a significant difference might be detected by using statistics when considering parameters that have a small variance, such as the electrolyte concentration or specific gravity of urine samples. In this case, statistical results are ignored, and the clinical relevance is considered.

#### Checking the reliability of the control group

It is sometimes necessary to check whether or not the control group used in a study is in the normal range. The comparison of the treatment group with double controls (usually used in basal diet experiments) is also made by statistical analysis. Alternatively, the control group can be compared with the historical control value.

#### Investigation using published studies

##### 28-day repeated dosing study in accordance with the Chemical Substance Control Law

The examples of the 28-day repeated dose toxicity study in rats are available in the public domain ([http://dra4.nihs.go.jp/mhlw\\_data/jsp/SearchPage.jsp](http://dra4.nihs.go.jp/mhlw_data/jsp/SearchPage.jsp)) and are presented in Table 1.

When the data of the treatment groups were compared with the historical data, it was found that the mean values

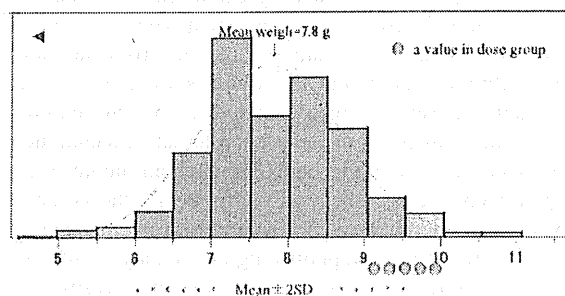


Fig. 1. Is there an individual of the examination group within mean  $\pm 2$  S.D. of the historical value?

of the treatment groups were within the range of  $\pm 2$  S.D. of the historical control data, but not within the range of  $\pm 2$  S.E. of it (except GPT of CAS No. 97-52-9) (Table 1). In the above studies, the authors judged whether the significant differences shown by the treatment groups were incidental or not by examining the data of the treatment groups with regard to the range of  $\pm 2$  S.E. of the respective historical control data. If the data of the treatment groups fall within the range of  $\pm 2$  S.E. of the historical control, they were considered as incidental. However, when the data were analyzed using *t*-test, adrenal weight (CAS No. 7756-94-7) and total protein (CAS No. 56-93-9) of the high dose were significant. Though the calculation procedure of *t*-test is based on the SEs of the treatment and historical control groups, it should be borne in mind that the number of animals used in these groups is different, being much larger in the latter, since the source of historical control data is several studies. But, while comparing the values within the range of the standard error, the number of animals used in each group is not taken into consideration.

In the repeated dose toxicity studies, when a significant difference between the treated and control values is detected at the low or medium doses, the determination of NOEL is difficult. An attempt was made to find out a solution to this problem by investigating 28-day repeat dose toxicity gavage studies in Sprague-Dawley (SD) rats wherein 126 test substances were analyzed in accordance with the Chemical Substance Control Law guidelines ([http://dra4.nihs.go.jp/mhlw\\_data/jsp/SearchPage.jsp](http://dra4.nihs.go.jp/mhlw_data/jsp/SearchPage.jsp)) (Table 2). Among the studies investigated, only one testing facility in Japan described historical control mean value  $\pm$  S.D. and the number of animals clearly in the report. The statements, such as "No dose-related pattern or dose dependency", made in the reports clearly show that there is no significant difference in the high and/or medium dose groups. It seems that the dose-response pattern and/or the dose dependency in most of the studies have been evaluated solely by a macroscopic decision. The following or similar statement from several report, supports our view, "this change was within the physiological range and/or the historical control range or is minimal change." In these reports, the authors have not given the range of values for the historical controls. Thus, the study director assumed that the change in this dosage group to be an incidental change without a statistics or solid scientific support.

#### Studies published in journals

On investigating the changes of parameters of dosage groups of several repeated dose toxicity studies in rodents

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**Table 3.** Grounds considered as incidental changes to the significant differences in dosed group of repeated dose toxicity studies

Reason assumed to be incidental change	Reference
Presence of dose-related pattern, visually	Nishiguchi <i>et al.</i> , 1997; Nishiguchi <i>et al.</i> , 1994; Takagi <i>et al.</i> , 1992b; Nakano <i>et al.</i> , 1992; Tamura <i>et al.</i> , 1983; Yamazaki <i>et al.</i> , 2005; Chemical product Safety Center, 1994; Griffiths <i>et al.</i> , 2007; Topping <i>et al.</i> , 2007; Guijie <i>et al.</i> , 2006; Poon <i>et al.</i> , 1998; McClain <i>et al.</i> , 2006; Sato <i>et al.</i> , 2007; Hellwig <i>et al.</i> , 1993; Webb <i>et al.</i> , 1993; Mellert <i>et al.</i> , 2002; Bär <i>et al.</i> , 1995; Arterburn <i>et al.</i> , 2000; Goldsmith, 2000; Lee <i>et al.</i> , 2004; Janssen <i>et al.</i> , 2000; Kanki <i>et al.</i> , 2003; O'Hagan and Menzel, 2003; Nakamura <i>et al.</i> , 2001; Thomas <i>et al.</i> , 1991; Abdo <i>et al.</i> , 1986; Morgan <i>et al.</i> , 1989; Dunnick <i>et al.</i> , 1987; Okazaki <i>et al.</i> , 2002; Okazaki <i>et al.</i> , 1993; Kato <i>et al.</i> , 1993; Jeong <i>et al.</i> , 2006; Shim <i>et al.</i> , 2003
No change in a related parameter	Inui <i>et al.</i> , 1997; Takeuchi <i>et al.</i> , 1985; Jonker <i>et al.</i> , 1993; MacKenzie <i>et al.</i> , 1992a; Barber and Topping, 1995; Oshima <i>et al.</i> , 1999; Suzuki <i>et al.</i> , 1997; Graça <i>et al.</i> , 2007; Shimpō <i>et al.</i> , 1990
High or low value of control group	Inui <i>et al.</i> , 1997; Mellert <i>et al.</i> , 2002; Macrì <i>et al.</i> , 1987
Study director's judgment (no significant biological difference effect, negligible change, within physiological change or sporadic change)	Takagi <i>et al.</i> , 1994a; Nakano <i>et al.</i> , 1992; Kato <i>et al.</i> , 1991; Takahashi <i>et al.</i> , 1986; Omosu <i>et al.</i> , 2003; Griffiths <i>et al.</i> , 2007; Topping <i>et al.</i> , 2007; Inui <i>et al.</i> , 1997; Cho <i>et al.</i> , 2006; MacKenzie <i>et al.</i> , 1992a, 1992b; Juberg <i>et al.</i> , 1998; Horváth <i>et al.</i> , 2002; Kotkoskie <i>et al.</i> , 1998; Oshima <i>et al.</i> , 1999; Yi <i>et al.</i> , 2007
Within historical range (background data) or normal range (not described to macroscopic or statistic significant)	Takahashi <i>et al.</i> , 1986; Tamura <i>et al.</i> , 1983; Omosu <i>et al.</i> , 2003; Griffiths <i>et al.</i> , 2007; Guijie <i>et al.</i> , 2006; McClain <i>et al.</i> , 2006; Horváth <i>et al.</i> , 2002; Kitamura <i>et al.</i> , 2003; Hart, 1988; Suzuki <i>et al.</i> , 1994; Shiraiishi <i>et al.</i> , 2006
Compared with another control other than vehicle control (using by double control)	Takahashi <i>et al.</i> , 1986; Webb <i>et al.</i> , 1993; Arterburn <i>et al.</i> , 2000
Statistics processing with normal values (historical data) range	Cerdá <i>et al.</i> , 2003

1. There is no significant difference if the mean value of if all the individual values of the dosage group fall within mean  $\pm$  2 S.E. for the historical control values.

2. There is no significant difference if all the individual values of the dosage group fall within mean  $\pm$  2 S.D. of the historical control values.

3. The statistical significant difference of the mean values between the dosage and historical control values group may be analyzed using the *t*-test.

4. The dose-response pattern/dose dependency may be analyzed using statistical techniques, for instance, Jonckheere trend test can be applied.

We recommend one in the above-mentioned four techniques to be used. However we suggest that the decision may not be made entirely on the basis of statistical analysis, but biological relevance of the statistical analysis may also be looked into.

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#### REFERENCES

- Abdo, K.M., Rao, G., Montgomery, C.A., Dinowitz, M. and Kanagalingam, K. (1986): Thirteen-week toxicity study of d- $\alpha$ -tocopheryl acetate (Vitamin E) in Fischer 344 rats. *Food Chem. Toxicol.*, **24**, 1043-1050.
- Arterburn, L.M., Boswell, K.D., Koskelo, E., Kassner, S.L., Kelly, C. and Kyle, D.J. (2000): A combined Subchronic (90-day) toxicity and neurotoxicity study of a single-cell source of docosahexaenoic acid triglyceride (DHASCO oil). *Food Chem. Toxicol.*, **38**, 35-49.
- Bär, A., Til, H.P. and Timonen, M. (1995): Subchronic oral toxicity study with regular and enzymatically depolymerized sodium carboxymethylcellulose in rats. *Food Chem. Toxicol.*, **33**, 909-917.
- Barber, E.D. and Topping, D.C. (1995): Subchronic 90-day oral tox-

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- icity study of suplatast tosilate (IPD-1151T) in rats. *J. Toxicol. Sci.*, **17**, Suppl. 2, 11-38.
- Nishiguchi, Y., Adachi, T., Nakazawa, M., Tawaratani, T., Uchimoto, H., Yoshida, M., Ishibashi, S., Kitayama, E., Iwakura, K. and Sumi, N. (1997): A 13-week oral repeated dose toxicity study of (+/-)-4-diethylamino-1,1-dimethylbut-2-yn-1-yl 2-cyclohexyl-2-hydroxy-2-phenylacetate monohydrochloride monohydrate (NS-21), a novel drug for urinary frequency and incontinence, in rats followed by a 5-week recovery test. *J. Toxicol. Sci.*, **22**, Suppl. 1, 27-57.
- Nishiguchi, Y., Ishibashi, S., Tarui, M., Fujimoto, M., Tawaratani, T., Yoshida, M., Uchimoto, H., Adachi, T., Hamasu, Y. and Sumi, N. (1994): 13-week oral toxicity study of lactitol (BS-4) in rats followed by 5-week recovery test. *J. Toxicol. Sci.*, **19**, Suppl. 3, 305-326.
- O'Hagan, S. and Menzel, A. (2003): A Subchronic 90-day oral rat toxicity study and *in vitro* genotoxicity studies with a conjugated linoleic acid product. *Food Chem. Toxicol.*, **41**, 1749-1760.
- Okazaki, K., Imazawa, T., Nakamura, H., Furukawa, F., Nishikawa, A. and Hirose, M. (2002): A repeated 28-day oral dose toxicity study of 17alpha-methyltestosterone in rats, based on the "enhanced OECD test guideline 407" for screening the endocrine-disrupting chemicals. *Arch. Toxicol.*, **75**, 635-642.
- Okazaki, S., Ueno, H., Anabuki, K., Tamura, K., Takeuchi, M., Hasegawa, H. and Kawase, Y. (1993): A 3-month oral toxicity study of MCP-1034 in rats with a recovery period of one month. *Jpn. Pharmacol. Ther.*, **21**, 947-974.
- Omosu, M., Eluard, B., Thien-A, H., Stepniowski, J.P., Vidal, J.M., Bode, G., Roeder, V., Sacre-Salem, B. and Guffroy, M. (2003): Oral repeated-dose toxicity studies and toxicokinetic studies of telithromycin in rats. *Jpn. J. Chemother.*, **51**, 103-114.
- Oshina, T., Sugiura, M., Takasu, K., Kimura, M. and Kudo, S. (1999): Repeated dose toxicity study of 3-amino-4-[[[(7)-4-[[4-(piperidylmethyl)-2-pyridyl]oxy]-2-butenyl]amino]-3-cyclobutene-1,2-dione monohydrochloride (Pibutidine hydrochloride) in rats (1): 13-week oral toxicity study with 4-week recovery test. *Jpn. Pharmacol. Ther.*, **27**, 331-345.
- Poon, R., Chu, L., Lecavalier, P., Valli, V.E., Foster, W., Gupta, S. and Thomas, B. (1998): Effects of antimony on rats following 90-day exposure via drinking water. *Food Chem. Toxicol.*, **36**, 21-35.
- Sakaki, H., Igarashi, S., Ikeda, T., Imamizo, K., Omichi, T., Kadota, M., Kawaguchi, T., Takizawa, T., Tsukamoto, O., Terai, K., Tozuka, K., Hirata, J., Handa, J., Mizuma, H., Murakami, M., Yamada, M. and Yokouchi, H. (2000): Statistical method appropriate for general toxicological studies in rats. *J. Toxicol. Sci.*, **25**, 71-98.
- Sato, S., Kitamura, H., Chino, M., Takci, Y., Hiruma, M. and Nomura, M. (2007): A 13-week oral dose subchronic toxicity study of gardenia yellow containing geniposide in rats. *Food Chem. Toxicol.*, **45**, 1537-1544.
- Shim, H.J., Kim, Y.C., Jang, J.M., Park, K.J., Kim, D.H., Kang, K. K., Ahn, B.O., Kwon, J.W., Kim, W.B. and Lee, M.G. (2003): Subacute toxicities and toxicokinetics of DA-8159, a new phosphodiesterase type V inhibitor, after single and 4-week repeated oral administration in rats. *Biopharm. Drug Dispos.*, **24**, 409-418.
- Shimpo, K., Takeuchi, M., Okazaki, S., Kiguchi, M., Hashimoto, Y., Aoki, Y., Kuwata, M. and Yamashita, K. (1990): Three-month oral subacute toxicity study of mozezofac (N-22) in rats. *J. Toxicol. Sci.*, **15**, Suppl. 2, 11-41.
- Shiraishi, K., Miyata, K., Houshuyama, S., Imatanaka, N., Umamo, T., Minobe, Y. and Yamasaki, K. (2006): Subacute oral toxicity study of diethylphthalate based on the draft protocol for "enhanced OECD test guideline no. 407". *Arch. Toxicol.*, **80**, 10-16.
- Suzuki, S., Ogawa, Y., Kamata, E., Naïto, K., Umemura, T., Saito, M., Kaneko, T. and Kurokawa, U. (1994): A 28-day repeated dose toxicity test of Phoxim in Wistar rats. *Eisei Shikenjo Houkoku*, **112**, 40-47.
- Suzuki, T., Koike, Y., Kitagaki, T. and Yamoto, T. (1997): A 90-day oral repeated dose toxicity study of azechidipine in rats followed by a 28-day recovery period. *Jpn. Pharmacol. Ther.*, **25**, 977-1004.
- Takagi, A., Momma, J., Aida, Y., Takada, K., Suzuki, S., Naitoh, K., Tobe, M., Hasegawa, R. and Kurokawa, Y. (1992): Toxicity studies of a synthetic antioxidant, 2,2'-methylenebis(4-ethyl-6-tert-butylphenol) in rats. I. Acute and subchronic toxicity. *J. Toxicol. Sci.*, **17**, 135-153.
- Takagi, A., Takada, K., Sai, K., Ochiai, T., Matsumoto, K., Sekita, K., Momma, J., Aida, Y., Saitoh, M., Naitoh, K., Furuya, T., Hasegawa, R. and Kurokawa, Y. (1994): Acute, Subchronic and chronic toxicity studies of a synthetic antioxidant, 2,2'-methylenebis(4-methyl-6-tert-butylphenol) in rats. *J. Toxicol. Sci.*, **19**, 77-88.
- Takahashi, N., Kadota, T., Kawano, S., Ohta, K., Ishikawa, K., Kuroyanagi, K., Hamajima, Y., Ohta, S., Kai, S. and Kohmura, H. (1986): Toxicity studies of VP16-213(II)--Oral one-month subacute toxicity in rats. *J. Toxicol. Sci.*, **11**, Suppl. 1, 17-49.
- Takeuchi, M., Ichionhe, M., Okazaki, S., Sato, M., Kiguchi, M., Shimpo, K., Sudo, J. and Tanabe, T. (1985): Study on toxicity of halopredone acetate. (II). Subacute toxicity study in rats. *J. Toxicol. Sci.*, **10**, Suppl. 1, 11-39.
- Tamura, J., Sato, N., Ezaki, H., Miyamoto, H., Oda, S., Hirai, K., Takado, H., Matsumoto, M. and Shirai, T. (1983): Acute toxicity of ranitidine and its metabolite in mice, rats and rabbits, and subacute oral toxicity of ranitidine in rats. *J. Toxicol. Sci.*, **8**, Suppl. 1, 1-24.
- Thomas, W.C., McGrath, L.F., Baarson, K.A., Auletta, C.S., Daly, I.W. and McConnell, R.F. (1991): Subchronic oral toxicity of cellulose acetate in rats. *Food Chem. Toxicol.*, **29**, 453-458.
- Topping, D.C., Bernard, L.G., O'Donoghue, J.L. and English, J.C. (2007): Hydroquinone: acute and subchronic toxicity studies with emphasis on neurobehavioral and nephrotoxic effects. *Food Chem. Toxicol.*, **45**, 70-78.
- Webb, D.R., Wood, F.B., Bertram, T.A. and Fortier, N.E. (1993): A 91-day feeding study in rats with caprenin. *Food Chem. Toxicol.*, **31**, 935-946.
- Williams, D.A. (1971): A test for differences between treatment means when several dose levels are compared with a zero dose control. *Biometrics.*, **27**, 103-117.
- Williams, D.A. (1972): The comparison of several dose levels with zero dose control. *Biometrics.*, **28**, 519-531.
- Yamazaki, K., Aiso, S., Matsumoto, M., Arito, H., Nagano, K., Yamamoto, S. and Matsushima, T. (2005): Thirteen-week oral toxicity study of 1,4-dichloro-2-nitrobenzene in rats and mice. *Ind. Health*, **43**, 597-610.
- Yi, Y., Yoon, H.J., Kim, B.O., Shim, M., Kim, S.O., Hwang, S.J. and Seo, M.H. (2007): A mixed polymeric micellar formulation of itraconazole: Characteristics, toxicity and pharmacokinetics. *J. Control Release*, **117**, 59-67.



63 events aside from mutations are encompassed in these test  
64 systems.

65 **Current Quantitative Structure–Activity Relationship**  
66 **(QSAR) Approaches.** The importance of assessing genotoxicity  
67 coupled with the availability of experimental data has prompted  
68 many *in silico* studies. James and Elisabeth Millers's "electro-  
69 philic theory" introduced a chemical concept to help rationalize  
70 the mode of action of genotoxic carcinogens.<sup>1</sup> This prompted  
71 many evaluations to derive so-called structural alerts (SA), simple  
72 yet effective means of encoding qualitative mechanistic under-  
73 standing for predicting potential mutagenicity/carcinogenicity.  
74 Seminal efforts include SA for carcinogenicity by John Ashby,<sup>2</sup>  
75 who subsequently extended his list with additional SA.<sup>3</sup> Bailey  
76 et al. compiled a set of 33 SAs for regulatory use within the  
77 U.S. Food and Drug Administration (FDA), which was predom-  
78 inantly based on the Ashby alerts.<sup>4</sup> Kazius et al. evaluated  
79 a mutagenicity database comprising 4337 mutagens and non-  
80 mutagens taken from the Toxnet database (<http://toxnet.nlm.nih.gov/>) and derived 29 SAs for mutagenicity with associated  
82 detoxification fragments.<sup>5</sup> Some of these alerts exist in software  
83 platforms to enable routine use; for example, 17 SAs for muta-  
84 genicity are implemented into the OASIS tissue metabolism  
85 simulator (TIMES) software.<sup>6</sup> Benigni et al. combined the pub-  
86 lished information from Ashby, Bailey et al., and Kazius et al.  
87 with additional information from the OncoLogic (U.S. EPA)  
88 software (<http://www.epa.gov/oppt/sf/pubs/oncologic.htm>)<sup>7</sup>  
89 to arrive at a list of 33 SA for carcinogens and mutagens.<sup>8</sup>

90 Current quantitative strategies include (Q)SARs and expert  
91 systems. Two types of (Q)SAR models, local and global, exist to  
92 estimate the mutagenic potential of chemicals. Local (Q)SARs  
93 provide estimated results for closely related (congeneric) chem-  
94 ical structures. Such models are most predictive, but only if the  
95 essential features of the model domains are clearly represented.  
96 Models based on physicochemical descriptors with clear mecha-  
97 nistic meaning are particularly helpful in rationalizing genotoxic  
98 outcome as exemplified by Chung et al.<sup>9</sup> Other local models are  
99 based on mathematical representations of chemical structure,  
100 for example, topological indices, and thus are more difficult to  
101 interpret.<sup>10</sup>

102 Global (Q)SARs aim to provide mutagenicity estimations  
103 for a diverse (noncongeneric) set of chemicals. Such (Q)SARs  
104 may be additionally encoded into expert systems. For example,  
105 TOPKAT empirically makes predictions for a range of different  
106 end points including Ames mutagenicity and rodent carcino-  
107 genicity.<sup>11</sup> Other expert systems such as TIMES attempt to  
108 provide clear mechanistic meaning through the use of SAs,  
109 which address the reactivity toward DNA and/or proteins.<sup>12,13</sup>  
110 TIMES also includes 3D QSARs to underpin some of the avail-  
111 able SAs. All of the aforementioned (Q)SARs have typically  
112 been derived on Ames (*Salmonella* mutagenicity data). TIMES  
113 includes a platform for *in vitro* CA data in addition to that for  
114 Ames.<sup>13</sup> There is a paucity of models for *in vivo* genotoxicity,  
115 but as highlighted in the survey by Benigni et al., there is only  
116 one publically available model for *in vivo* micronucleus.<sup>14</sup> The  
117 scarcity of such models may be due in part to experimental data  
118 being less readily available but also due to the complexity of  
119 how to rationalize and interpret the outputs from the different  
120 test systems.

121 Our own investigation aims to fill in the above *in vitro*–*in*  
122 *vivo* genotoxicity gap by considering both the available test  
123 systems and how they are currently applied to formulate an  
124 approach for modeling *in vivo* genotoxicity. For convenience,  
125 we considered the REACH ITS<sup>15</sup> for mutagenicity since this

described the typical assays used and how their outcomes  
should be interpreted for subsequent decision making. The  
actual experimental test systems are assumed to be reasonably  
familiar and are only briefly described in the next section.

**Experimental Assays and Data for Rodent Mutage-  
nicity and Genotoxicity.** Integrated testing strategies, notably  
those described in the REACH Technical guidance,<sup>15</sup> outline  
the *in vitro* and *in vivo* systems that are most frequently used to  
evaluate the mutagenic potential of chemical substances. The  
*in vitro* systems include the bacterial reverse mutation test (Ames),  
an *in vitro* mammalian cell gene mutation test [such as the  
mouse lymphoma or hypoxanthine–guanine phosphoribosyl-  
transferase (*hprt*) assay], the *in vitro* mammalian chromosome  
aberration (CA) test, and the *in vitro* MNT.<sup>15</sup> The Ames test  
uses amino acid-requiring strains of bacteria to detect (reverse)  
gene mutations (point and frameshift mutations). The *in vitro*  
mouse lymphoma assay (MLA), when correctly performed,  
detects structural chromosome aberrations, aneuploidy, and  
recombination events (e.g., such as gene conversion) that result  
in loss of heterozygosity. The *hprt* test identifies chemicals that  
induce gene mutations in the *hprt* gene of established cell lines.  
The *in vitro* mammalian CA test detects structural chromo-  
some aberrations and increases in polyploidy. The *in vitro* MNT  
has the potential to detect both clastogenic (chromosome aber-  
rations) and aneugenic (chromosome lagging due to dysfunction  
of mitotic apparatus) chemicals.

The scheme under REACH can be summarized as follows.  
As a first tier, three *in vitro* tests are recommended, which  
includes an Ames test, a mouse micronucleus/CA, and a mouse  
lymphoma/HRPT assay. If the results from all three tests are  
negative, then no more testing is merited, and a conclusion of  
nongenotoxicity can be made for the substance under study. If  
one or more tests are positive, then *in vivo* testing may be insti-  
gated. Obviously metabolism, pharmacokinetics, and toxicoki-  
netics factors [absorption, distribution, metabolism, excretion  
(ADME)] are all inherent features in the *in vivo* genotoxicity  
tests, although the genetic end points for the tests address dif-  
ferent genetic mechanisms. The UDS *in vivo* assay is used to  
evaluate the role of DNA repair. The *in vivo* Comet assay is a  
sensitive technique for the detection of DNA strand breaks;  
thus, it can be used for measuring DNA strand breaks in any  
tissue of an animal. Site-specific effects at contact tissues or the  
target tissue where the test compound accumulates or induces  
toxicity can be readily assessed. The specificity of the contact  
tissue under investigation is also feasible for the transgenic  
rodent gene mutation test (TGR), which measures gene muta-  
tions *in vivo*. However, the *in vivo* MNT is probably the most  
widely used test.<sup>16</sup> When performed appropriately, it detects  
both clastogenicity and aneugenicity.<sup>17</sup> The frequency of micro-  
nucleated polychromatic erythrocytes is traditionally determined  
from bone marrow samples, but with the emerging automated  
scoring methods, the emphasis is moving to assessing the induc-  
tion of micronuclei in immature erythrocytes in peripheral blood  
samples.<sup>18</sup>

Most of the established *in vitro* mutagenicity tests, which are  
used for regulatory purposes, exhibit relatively high sensitivity  
for detection of genotoxic carcinogens.<sup>19</sup> However, particularly  
those based on cultured mammalian cells are thought to pro-  
duce a remarkably high occurrence of irrelevant positive results  
(i.e., exhibit low specificity), when compared with rodent carci-  
nogenicity.<sup>19,20</sup> To increase the specificity of predictions, regu-  
lators tend to interpret *in vitro* positive results in an *in vivo*  
perspective, that is, *in vivo* confirmation of *in vitro* mutagens.

189 In addition, *in vivo* tests can also be utilized to identify chemicals producing *in vivo* only positive results (i.e., chemicals for which mutagenicity is not or poorly detected *in vitro*). Only a very limited number of chemicals have been found to be genotoxic *in vivo* and not in the standard *in vitro* tests. Most of these are pharmaceuticals such as atovaquone (95233-18-4), which is designed to affect pathways of cellular regulation, including cell cycle regulation. One of the most preferred *in vivo* assays, complementing genotoxicity test batteries, is the *in vivo* bone marrow MNT. The preference of this assay is attributed to both its wide mutagenicity range assessment (clastogenicity and aneugenicity) and its remarkably high specificity in concordance with the genotoxic carcinogenicity model, although it shows low sensitivity.<sup>14,21</sup> Therefore, it may be appropriate to include a second *in vivo* test if a positive *in vitro* result has not been adequately confirmed by the *in vivo* bone marrow MNT test. The UDS test is one complement to the bone marrow MNT since it is a surrogate *in vivo* gene mutation assay measuring DNA excision repair of induced DNA damage. The utility of the Comet and the TGR assays to detect genotoxic damage in specific tissues, specifically DNA strand breaks and gene mutations has also been recognized.<sup>15</sup> Thus, an evaluation of *in vivo* genotoxicity potential could involve integrating outcomes from MNT and either UDS, Comet, and TGR tests depending on the outcomes that have been observed *in vitro*. UDS, Comet, and TGR can also be undertaken to address *in vivo* liver genotoxicity. Such tissue-specific assays are useful in *in vivo* follow-up tests especially since the liver is an organ of high metabolic capacity and therefore is frequently subjected to significant toxic overload.

**Aims of the Study.** Bearing in mind the way in which these different assays are integrated together, our goal was to investigate the *in vitro* and *in vivo* relationship, the so-called *in vitro*–*in vivo* “gap” to inform the development of mechanistic (Q)SAR model(s). A large body of data covering *in vitro* mutagenicity, *in vivo* (liver) genotoxicity, and *in vivo* bone marrow MNT test results was collected for the same set of substances. The scope of the investigation can be summarized in the following three questions: (a) To what extent are *in vitro* mutagenic chemicals *in vivo* (liver) genotoxic, that is, what *in vivo* detoxification pathways exist? (b) To what extent are *in vivo* (liver) genotoxic chemicals *in vivo* bone marrow MNT positive? (c) Are there *in vitro* nonmutagenic chemicals that are *in vivo* liver or bone marrow genotoxic; that is, what *in vivo* bioactivation pathways exist? These questions were structured into a workflow (Figure 1) and enabled a stepwise evaluation of the *in vitro*–*in vivo* gap.

## MATERIALS AND METHODS

**Compilation of Data Set.** Our training set comprised 557 chemicals (“557 list”) with *in vivo* MNT data (Appendix I of the Supporting Information lists the substances and their overall calls). In *in vitro* mutagenicity and *in vivo* (liver) data were collected for the same set of substances to the extent possible. This helped maximize the overlap between chemicals with various genotoxicity effects and the *in vivo* MNT data set. Documented *in vitro* mutagenicity data from multiple literature sources were identified for 397 noncongeneric chemicals within the training set (Appendix II of the Supporting Information). Positive calls were categorized by the digit 1, negative calls by 0, and N/A signified “no data available”, based on the literature searches that were performed. Our *in vitro* data comprised that from the Ames assay, the CA assay, and the MLA, since these are the typical assays considered under REACH. Out of necessity and as typically the case for modeling efforts, reported study results were accepted as

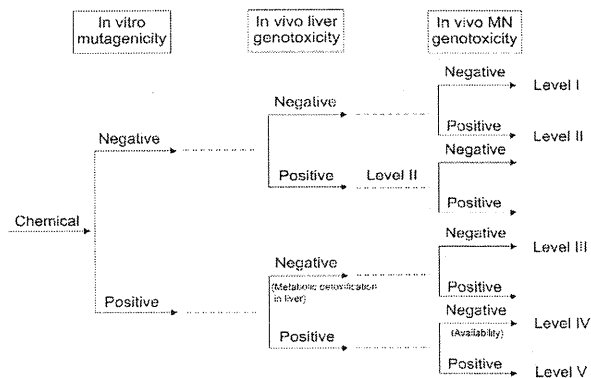


Figure 1. Workflow outlining the *in vivo*–*in vitro* gap.

reported, although an extensive effort was made in expert judgment and evaluation of the data quality and correctness of the calls.

Ames results with the rat liver S9 metabolic activation system were available for 283 noncongeneric chemicals. Of these chemicals, 109 (38%) were associated with positive calls and 174 (62%) with negative calls. Documented *in vitro* CA test data were identified for 296 chemicals, of which 186 (63%) were positive and 110 (37%) were considered negative. Data from 194 chemicals had been assessed in *in vitro* MLA. The majority of the chemicals tested positive (148 chemicals, i.e., 76%) and 46 chemicals (24%) tested negative. For the 397 *in vitro* mutagenicity data, these comprised 267 positive calls (68%) and 124 negative calls (32%), and six calls were inconclusive. These substances were ethylene dichloride (107-06-2), sulfan blue (129-17-9), thiabendazole (148-79-8), methyl parathion (298-00-0), dibutyl nitrosamine (924-16-3), C.I. direct black 38 (1937-37-7). In these six cases, only Ames and *in vitro* CA test outcomes were available with positive calls in Ames and negative calls in *in vitro* CA tests.

Results from *in vivo* Comet, UDS, and TGR assays were also collected to help evaluate *in vivo* liver genotoxic potential. Data were available for 185 diverse chemicals, which are listed in Appendix III of the Supporting Information. The Comet assay provided liver genotoxicity assignments for 127 (69%) of the 185 chemicals. Of the 127 chemicals, 78 (61%) were positive, and 49 (39%) were negative. The TGR comprised rodent liver genotoxicity data for 34 (18%) of the 185 chemicals; 27 (80%) of these were reported as positive, and 7 (20%) were negative. The *in vivo* UDS assay was associated with the least amount of liver genotoxicity data, only 24 (13%) of the 185 chemicals had overall calls, and five of them were observed to be positive in this assay (21%), and 19 were (79%) negative in this assay. Overall, of the 185 substances with liver assignments, 109 were associated with positive calls (59%) and 76 with negative calls (41%). The “557 list” included almost equal numbers of positive (267 chemicals, i.e., 48%) and negative (290 chemicals, i.e., 52%) MNT assignments performed in either bone marrow or peripheral blood. Figure 2 summarizes the distribution of assignments in each of the test systems.

The evaluation of this investigation was often hampered by conflicting *in vivo* MNT data available in the public domain. The compromised quality of these MNT data was attributed to the fact that many chemicals had been evaluated in the early 1980s; when species (rat vs mouse) and gender (male vs female) differences may not always have been considered, etc. To date, the validity of the *in vivo* MNT data has only been verified for chemicals where the *in vitro* mutagenicity outcome appeared to be negative, relative to the *in vivo* case (in either liver or bone marrow), where the genotoxicity result was positive. Expert judgment was relied upon to consider whether there were factors resulting in inconsistent *in vitro* results as compared with the *in vivo* situation, for example, rodent species differences, nonphysiological culture conditions, etc.

To illustrate the structural diversity of the training set, the 557 list was profiled against the set of DNA and protein binding alerts available within the OECD Toolbox v2.1. The distribution chart is



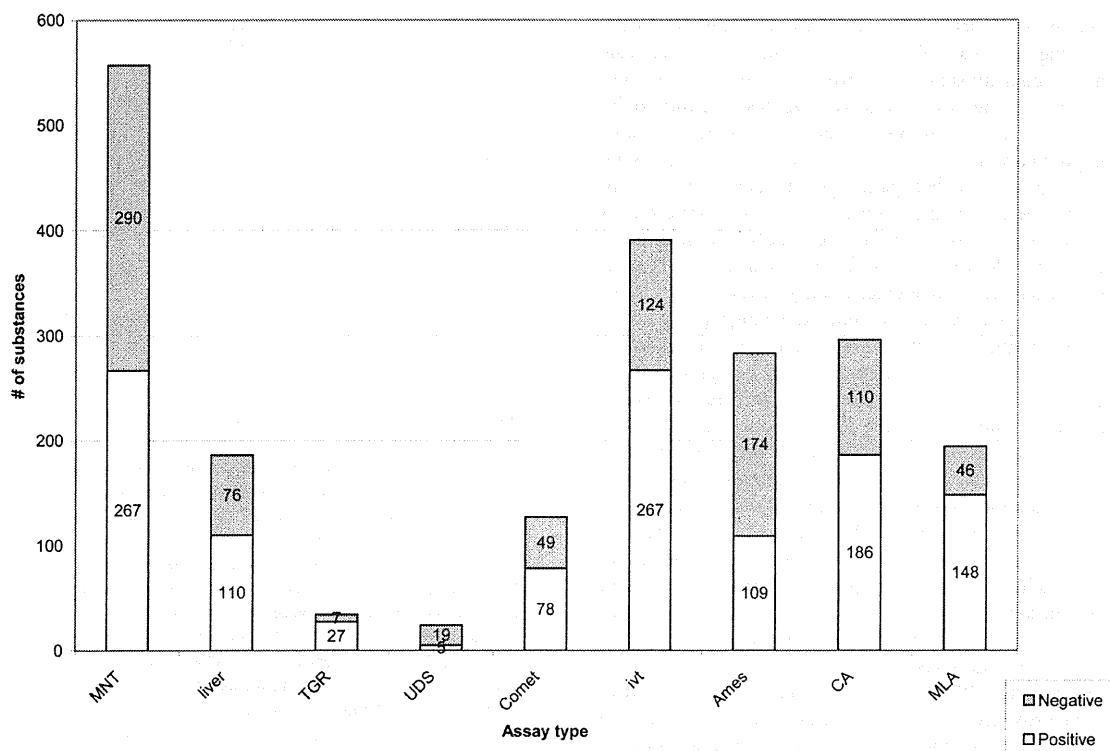


Figure 2. Distribution of the overall calls for each of the test assays under study.

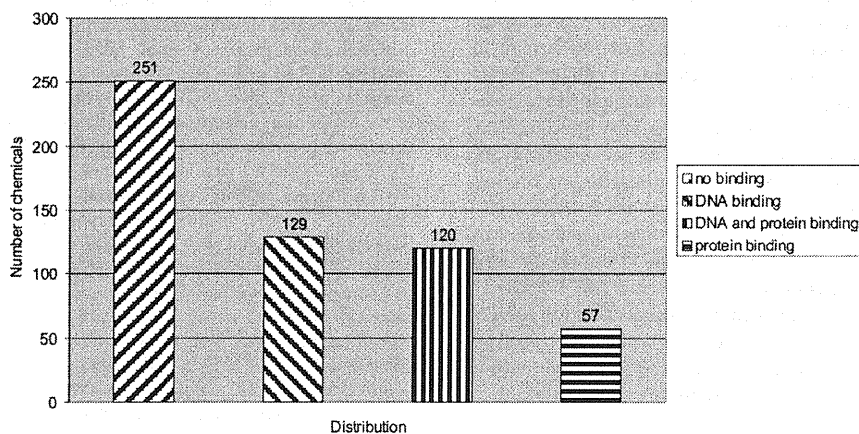


Figure 3. Distribution of training set chemicals across DNA and protein binding alerts.

303 shown in Figure 3. The results reveal that 251 (45%) of the 557 chem-  
 304 icals possess no DNA and/or protein binding alerts. One hundred  
 305 twenty-nine of the remaining 306 (55%) chemicals have one or more  
 306 DNA binding alerts, 57 chemicals have a protein binding alert, and 120  
 307 chemicals have both DNA and protein binding alerts. This distribution  
 308 shows a broad spread of chemical mechanisms as depicted by the SAS  
 309 triggered.


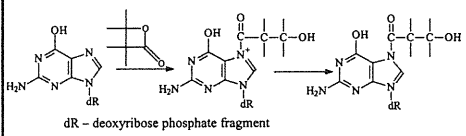

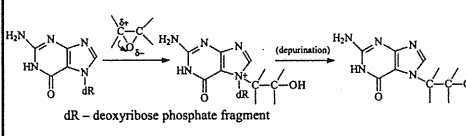
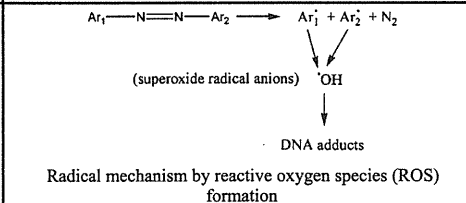
310 Our modeling approach sought to use the existing TIMES for-  
 311 malism and refine the components that had been originally developed  
 312 to estimate Ames and in vitro CA. Here, we provide a brief overview of  
 313 these components.

314 **Modeling Reactivity to DNA and Proteins.** According to the  
 315 working hypothesis, interaction of chemicals with DNA and/or with  
 316 specific proteins (such as histone, topoisomerase, spindle protein  
 317 tubulus, and DNA repair enzymes) encompasses a diversity of genotoxic

318 events, which can damage mammalian cells. For example, the forma-  
 319 tion of micronuclei arises as a result of the covalent interaction be-  
 320 tween chemicals with DNA and/or specific proteins. Accordingly, a  
 321 reactivity component for an in vivo model, which predicts genotoxic  
 322 effects such as formation of micronuclei or liver damages, should be  
 323 based on the assessment of the potential of that chemical to interact  
 324 with DNA and/or proteins.

325 TIMES models predicting the outcomes in Ames and the CA test  
 326 have previously been published.<sup>12,13</sup> It has been established that the  
 327 Ames test primarily accounts for the direct interaction of chemicals  
 328 with DNA, whereas the in vitro CA test assesses both DNA and pro-  
 329 tein (e.g., histone, topoisomerase, spindle protein tubulus, and DNA  
 330 repair enzymes) binding. This implies that Ames mutagenic chemicals  
 331 should be CA positive, but the converse is not necessarily true.  
 332 A recent comparative analysis of in vitro mutagenic data for a large

Table 1. Alerting Groups and Descriptors Used in COREPA Models for Estimating Their Reactivity Associated with Supporting Mechanistic Information<sup>a</sup>

#	Alerting group	Chemical class	Descriptors in the COREPA model*	Interaction mechanism	Reference
1		Lactones	-	 dR - deoxyribose phosphate fragment Ring opening S <sub>N</sub> 2 reaction	(77)
2		Epoxides	MW E <sub>HOMO</sub>	 dR - deoxyribose phosphate fragment Ring opening S <sub>N</sub> 2 reaction	(78)
3	—N=N—	Azo compounds	log K <sub>OW</sub> Van der Waals surface	 Radical mechanism by reactive oxygen species (ROS) formation	(78)

<sup>a</sup>\*E<sub>HOMO</sub>, the energy of the highest occupied molecular orbital (eV); MW, molecular weight (Da); log K<sub>OW</sub>, octanol–water partitioning coefficient (mol L<sub>0</sub><sup>-1</sup> mol<sup>-1</sup> L<sub>w</sub>); and van der Waals surface area (Å<sup>2</sup>).

333 number of chemicals confirmed this assumption. Eighty percent of  
334 chemicals that elicited bacterial mutagenicity (based on Ames test  
335 results) also induced CA, whereas only 60% of chemicals that induced  
336 CA were found to be active in the Ames test.<sup>22,23</sup> To distinguish  
337 these two mechanisms, the reactivity component of the newly derived  
338 models for MNT and liver genotoxicity was structured into two parts.  
339 The first part accounted for the interaction of chemicals with DNA.  
340 More than 60 alerting groups (being considered as a part of a future  
341 publication) were used to simulate covalent interaction with DNA.  
342 The use of each alert had been justified by the mechanistic interpretation  
343 of that interaction. Some alerts were additionally underpinned  
344 by mechanistically based COMmon REactivity PAttern (COREPA)  
345 3D QSAR models.<sup>24,25</sup> Examples of these DNA binding alerts are presented  
346 in Table 1. The SAs are described together with physico-  
347 chemical property/molecular parameter exclusion/inclusion rules.  
348 Supporting reaction mechanism information is also provided.

349 As seen from Table 1, the SAs can be categorized into two types:  
350 (1) those eliciting mutagenicity without the need for modulating  
351 factors (#1 in Table 1) and (2) those for which specific molecular  
352 parameter(s) define the degree of activation (#2 and #3 in Table 1).

353 The second part of the reactivity component accounts for the interaction  
354 of chemicals with specific proteins. More than 50 SAs were  
355 proposed that were associated with protein interaction (<http://www.oasis-lmc.org/>).  
356 Examples of protein binding alerts associated with  
357 parameters for reactivity and their supporting reaction mechanism  
358 information are presented in Table 2. These are characterized  
359 similarly—either requiring modulating factors (#1, #2, and #3 in  
360 Table 2) or not (#4 in Table 2).

361 Most of the DNA binding alerts are also able to bind proteins. An  
362 example to demonstrate the mechanism by which a DNA binding alert  
363 interacts with proteins is presented for quinones in Figure 4.


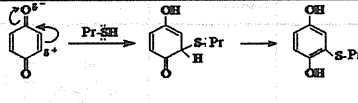
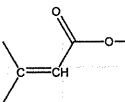
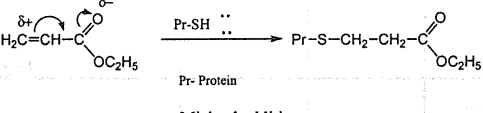

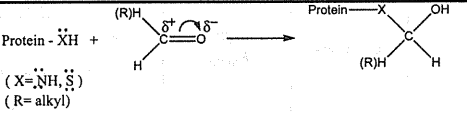
364 Quinones are well-known mutagens, and they are included in  
365 the list of DNA-causing alerts. Topoisomerases are enzymes that  
366 participate in all stages of replication, functional activity, and structural  
367 maintenance of DNA. The inhibition of these enzymes by quinones is

368 considered to elicit CA26. This is an example of how the same alert  
369 can elicit different outcomes depending on the interaction target. The  
370 structure of the reactivity component used in the in vivo genotoxicity  
371 models is provided in Figure 5.

372 A new chemical is first submitted to the reactivity component that  
373 encompasses the alerts associated with DNA interactions. A positive  
374 prediction for mutagenicity is assigned if the requirements for interaction  
375 with DNA are met, indicating that the ultimate mutagenic effect  
376 is due to this interaction mechanism. Regardless of whether the chemical  
377 meets the requirements for direct interaction with DNA, it is then  
378 forwarded to the second part of the reactivity component, which investigates  
379 the ability of the chemical to interact with proteins. This is to  
380 flag those cases where mutagenicity may arise by both mechanisms  
381 (direct interaction with DNA and interaction with protein) simultaneously.  
382 If the chemical passes through both parts of the reactivity component  
383 without being flagged for activity, a prediction of “unable to produce  
384 mutagenicity” is noted.

385 **Conformational Analysis by Genetic Algorithm.** To derive 3D  
386 QSARs, the flexibility of chemicals needs to be taken into account  
387 since this will give rise to the formation of many different conformers,  
388 and their reactivity profiles would accordingly differ. Common practice  
389 is to calculate molecular parameters for the lowest energy conformation,  
390 even though this necessarily may not be the form that drives the  
391 response and therefore not the most relevant one to study.<sup>27</sup> Given a  
392 systematic conformational analysis search would be computationally  
393 intensive (since the number of conformers would increase exponentially  
394 with the number of degrees of freedom), LMC derived a procedure  
395 to address the issue of conformation space using a genetic algorithm,  
396 which minimizes 3D similarity among generated conformers.<sup>28</sup>  
397 This made addressing the conformation space practical, even for large  
398 and very flexible chemicals. A procedure was also developed to saturate  
399 the conformation space, that is, to ensure consistency in the reproducibility  
400 of generated conformers and their distribution in the structural space.<sup>28</sup>  
401 This allowed the conformational space of chemicals to be populated  
402 with an optimal number of conformers.

Table 2. Alerting Groups for Protein Binding, Parameters for Reactivity, and Supporting Interaction Mechanisms<sup>a</sup>

#	Alerting group	Chemical class	Descriptors* in the model.	Interaction mechanism	Reference
1		Quinones	MW		(79)
3		Acrylates	log K <sub>OW</sub>	 Michael addition	(80)
4		Aldehydes		 Schiff base formation	(81)

<sup>a</sup>MW, molecular weight (Da); log K<sub>OW</sub>, octanol-water partitioning coefficient (mol L<sub>0</sub><sup>-1</sup> mol<sup>-1</sup>L<sub>w</sub>).

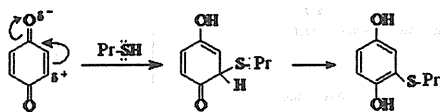


Figure 4. Interaction mechanism of quinones with proteins (Pr).

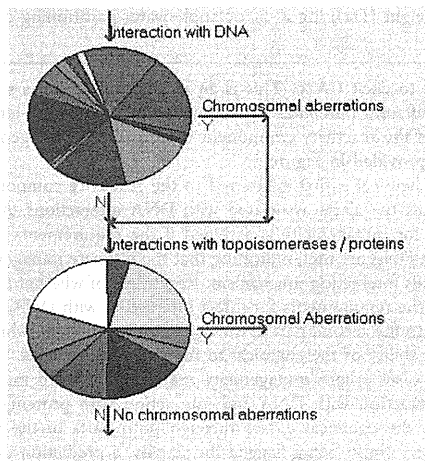


Figure 5. Structure of the reactivity component of the in vivo genotoxicity models.

403 **TIMES.** The TIMES platform comprises SA, 3D QSARs, and a  
404 metabolism simulator. This simulator comprises a list of hierarchically  
405 ordered transformations and a substructure matching engine for their  
406 implementation. The modeling is based on a probabilistic approach<sup>29</sup>  
407 whereby a hierarchy of transformations is defined by the probabilities  
408 of transformations determined in such a way as to reproduce a data-  
409 base of documented metabolic transformations or data for their rate  
410 of disappearance. The transformation probabilities are related to the  
411 feasibility of occurrence of various metabolic reactions. It is assumed  
412 that the transformations are independent and performed sequentially.  
413 Each molecular transformation consists of parent submolecular frag-  
414 ments, transformation products, and inhibiting masks. The latter play  
415 the role of reaction inhibitors. If a functional group assigned as a mask  
416 is attached to the target fragment, the execution of the transformation  
417 on the parent chemical is prevented. The presence of groups that

can promote or inhibit metabolic reactions significantly increases the 418  
number of principal transformations. Currently, 343 principal transfor- 419  
mations are used to model rat liver metabolism in vitro. The simulator 420  
starts by matching the parent molecule with the reaction fragment 421  
associated with the transformation having highest probability of occur- 422  
rence. When a match is identified, the molecule is metabolized, and 423  
transformation products are treated as parent molecules for the 424  
next degradation step. The procedure is repeated for the newly formed 425  
chemicals until the product of probabilities of consecutively performed 426  
transformations reaches a user-defined threshold. The mathematical 427  
formalism defining the amount of metabolite, formation, and meta- 428  
bolism probabilities is described elsewhere.<sup>6,29-31</sup> The intent with 429  
this study was to refine the existing structure-activity and structure- 430  
metabolism rules within TIMES to account for the differences 431  
observed between the in vitro and the in vivo results. Where a realistic 432  
and feasible hypothesis could be generated and substantiated with 433  
data, these would inform the refinement of existing rules or intro- 434  
duction of new transformation rules. 435

## RESULTS AND DISCUSSION

436  
437  
**Workflow for Genotoxicity at Different Levels of** 438  
**Biological Organization.** While the full set of data comprised 439  
557 chemicals, a set of data where results from all assays were 440  
available were required to develop the mechanistic (Q)SAR 441  
models. Overall, calls for in vitro, liver genotoxicity, and in vivo 442  
MNT were available for 162 chemicals. Table 3 shows the list 443  
of 162 chemicals. A hierarchical workflow (Figure 6) outlines 444  
the results.

The first tier of in vitro tests comprises 162 chemicals that 445  
were either positive or negative in Ames, CA, and MLA. Four 446  
chemicals were assigned as inconclusive since Ames and CA 447  
data were found to be conflicting. All four were Ames positive 448  
but CA negative. The four chemicals were ethylene dichloride 449  
(107-06-2), thiabendazole (148-79-8), dibutyl nitrosamine (924- 450  
16-3), and C.I. direct black 38 (1937-37-7). These were excluded 451  
from further study. Thirty-two (20%) of the 158 chemicals re- 452  
maining were found to be in vitro negative, and 126 (80%) were 453  
found to elicit in vitro positive responses. Substances were cate- 454  
gorized as negative if two or more results were negative and posi- 455  
tive if they were positive in at least one of the three tests. 456

The 32 (20%) nonmutagenic chemicals in vitro were inves- 457  
tigated in both liver and MNT in vivo tests. Thirty of the 32 in 458  
vitro nonmutagenic chemicals were confirmed negative in vivo 459

Table 3. List of the 162 Chemicals and Their Summary Calls Both in Vitro and in Vivo Test Systems

CAS	name	ivt	liver	MNT	CAS	name	ivt	liver	MNT
50-06-6	phenobarbital	1	1	1	97-56-3	<i>o</i> -aminoazotoluene	1	1	0
50-32-8	benzo( <i>a</i> )pyrene	1	1	1	99-56-9	1,2-diamino-4-nitrobenzene	1	0	0
50-55-5	reserpine	0	0	0	100-41-4	ethylbenzene	1	0	0
51-03-6	piperonyl butoxide	1	0	0	100-42-5	styrene	1	1	0
51-79-6	urethane	1	1	1	100-51-6	benzyl alcohol	1	0	0
52-24-4	thio-TEPA	1	1	1	100-75-4	1-nitrosopiperidine	1	1	0
56-04-2	methylthiouracil	0	0	0	101-14-4	4,4'-methylenebis(2-chlorobenzenamine)	1	1	1
56-23-5	carbon tetrachloride	0	0	0	101-77-9	4,4'-methylenebis(aniline)	1	1	1
56-57-5	4-nitroquinoline 1-oxide	1	1	1	103-33-3	aminoazobenzene	1	1	1
56-75-7	chloramphenicol	0	0	0	103-90-2	acetaminophen	1	1	1
57-14-7	dimazine	1	1	1	104-55-2	cinnamaldehyde	1	0	0
57-22-7	vincristine	1	0	1	105-11-3	<i>p</i> -quinone dioxime	1	0	0
57-30-7	phenobarbital, sodium	0	0	0	105-60-2	hexahydro-2 <i>h</i> -azepin-2-one	0	0	0
57-50-1	sucrose	0	0	0	106-46-7	1,4-dichlorobenzene	0	1	1
57-57-8	propiolactone	1	1	0	106-93-4	ethylene dibromide	1	1	0
57-97-6	7,12-dimethylbenz(A)anthracene	1	1	1	106-99-0	butadiene	1	0	0
58-08-2	caffeine	1	0	0	107-06-2	ethylene dichloride	no conclusion	1	0
58-89-9	lindane	0	0	0	107-13-1	acrylonitrile	1	0	0
59-05-2	methotrexate	1	1	1	108-88-3	toluene	0	0	0
59-89	<i>N</i> -nitrosomorpholine	1	1	1	108-95-2	phenol	1	1	0
60-09-2-3	<i>p</i> -aminoazobenzene	1	1	1	110-00-9	furan	1	1	0
60-11-7	4-dimethylaminoazobenzene	1	1	1	110-44-1	sorbic acid	0	0	0
60-35-5	acetamide	0	0	0	110-86-1	pyridine	0	0	0
60-57-1	dieldrin	1	1	1	117-39-5	quercetin	1	0	0
62-44-2	acetophenetidin	1	0	1	117-81-7	bis(2-ethylhexyl)phthalate	0	0	0
62-53-3	aniline	1	1	1	118-96-7	2,4,6-trinitrotoluene	1	0	0
62-55-5	thioacetamide	1	0	1	119-53-9	benzoin	1	0	0
64-86-8	colchicine	1	0	1	119-93-7	tolidine	1	1	1
66-27-3	methyl methanesulfonate	1	1	1	120-47-8	ethylparaben	1	0	0
67-20-9	nitrofurantion	1	1	0	120-71-8	<i>p</i> -cresidine	1	0	0
67-66-3	chloroform	1	0	0	121-79-9	propyl gallate	1	0	0
67-68-5	dimethyl sulfoxide	0	0	0	123-91-1	1,4-dioxane	0	0	0
68-12-2	dimethylformamide	0	0	0	124-48-1	chlorodibromomethane	1	1	0
70-25-7	<i>N</i> -methyl- <i>N'</i> -nitro- <i>N'</i> -nitrosoguanidine	1	1	1	126-72-7	tris(2,3-dibromopropyl) Phosphate	1	1	1
71-43-2	benzene	1	1	1	128-37-0	butylated hydroxytoluene	1	0	0
75-07-0	acetaldehyde	1	1	1	128-44-9	saccharin, sodium	0	0	0
75-09-2	methylene chloride	1	1	0	134-32-7	1-naphthylamine	1	1	1
75-25-2	bromoform	1	0	0	136-40-3	phenazopyridine hydrochloride [USAN]	1	1	1
75-56-9	propylene oxide	1	0	1	139-13-9	triglycollamic acid	1	1	0
79-06-1	acrylamide	1	1	1	140-11-4	benzyl acetate	0	0	0
79-34-5	1,1,2,2-tetrachloroethane	1	1	1	140-88-5	ethyl acrylate	1	1	0
81-07-2	saccharin	0	0	0	142-04-1	aniline HCl	1	1	1
84-16-2	hexestrol	1	0	0	147-94-4	cytosine arabinoside	1	0	1
89-65-6	erythorbic acid	0	0	0	148-79-8	thiabendazole	no conclusion	1	1
90-43-7	2-phenylphenol	1	1	0	148-82-3	melphalan	1	1	1
91-20-3	naphthalene	1	0	0	301-04-2	lead acetate	1	0	0
91-59-8	2-naphthalenamine	1	1	1	305-03-3	chlorambucil	1	1	1
91-64-5	coumarin	1	0	0	309-00-2	aldrin	1	0	0
91-94-1	3,3'-dichlorobenzidine	1	1	1	366-70-1	procarbazine hydrochloride	1	1	1
92-52-4	biphenyl	1	1	0	427-51-0	cyproterone acetate	0	1	0
92-67-1	4-biphenylamine	1	1	1	446-86-6	azathioprine	1	1	1
92-87-5	benzidine	1	1	1	492-80-8	auramine	1	1	0
95-50-1	1,2-dichlorobenzene	1	0	0	501-30-4	kojic acid	1	0	0
95-53-4	<i>o</i> -toluidine	1	1	0	532-32-1	sodium benzoate	1	0	0
95-80-7	2,4-diaminotoluene	1	1	0	542-75-6	1,3-dichloropropene [BSI:ISO]	1	1	0
95-83-0	4-chloro-1,2-diaminobenzene	1	1	1	602-87-9	5-nitroacenaphthene	1	1	1
96-09-3	styrene oxide	1	1	0	604-75-1	oxazepam	1	1	0
96-12-8	1,2-dibromo-3-chloropropane	1	1	1	609-20-1	2,6-dichloro- <i>para</i> -phenylenediamine	1	1	1
96-45-7	ethylenethiourea	1	1	0					
97-53-0	eugenol	1	0	0					