3-6. モデル適用範囲の検討

・3-2, 3-3, 3-4章では、構築した予測モデルにより予測されるPredicted Probabilitiesが0.5より大きい場合は「陽性」、0.5より小さい場合は「陰性」、と判断した上で予測精度が計算される。この判断基準を変更することで予測精度を上げることができると予想される。

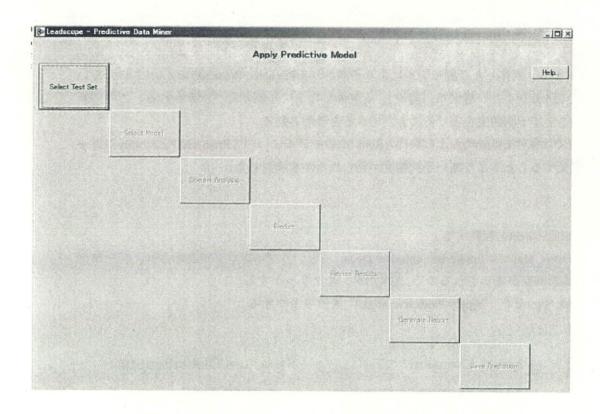
従って、以下では条件4(Run#4)および条件7(Run#16)のモデルについてPredicted Probabilityのカットオフの値を変更することでより高い予測精度が得られるかを検討した。

[手順]

- (1) Leadscope Hosted Clientを起動する。
- (2) 起動中に"Content Modules Enterprise: animal:9 (secur..."というダイアログが表示され、データベース一覧の確認画面が表示されるので、OKボタンをクリックする。
- (3) Leadscope Task Wizardで、"Apply Predictive Model"をクリックする。



(4) 下記のようなフローが表示されるので、モデル構築の際と同様に上から順にパネルをクリック していき各設定をすることでモデルを適用できる。個別的な設定内容は3-6-1以降の各セクショ ンで表にまとめた。



<u>3-6-1. 条件4のモデル(Run#4)のLiver conv2に対する適用範囲の検討:</u>

項目 authorize strong)	選択条件
Select Test Set: Select Test Set Project	Liver_training
Select Model: Select Predictive Model	PredictiveModel(Run #4)
Domain Analysis : How you would like to select the test domain	Manually select structures*

[※]Domain AnalysisでAutoにすると予測範囲外の化合物が出てくるので、Manualを選択した。

以下にModel構築の結果を記載した。

判定のカットオフ値をマニュアル設定にして予測した場合:

項目	選択条件
Select Reference Property (Optional)	Liver_conv2
Qualified Prediction: How you would like to define the property threshold:	Expert Definition: Manually define the probability threshold
	Select the minimum probability needed for a positive call: 0.75
	Select the maximum probability needed for a negative call: 0.25

以下に予測結果を示した。

		条件4(Run#4)			%
		+	-	total	concordance	91.35
Liver_conv2	+	118	27	145	sensitivity	81.38
	-	16	336	352	specificity	95.45
		-		497		
予測可能な化	合物	7の割合(%	6):	61.51		

3-6-2. 条件4のモデル(Run#4)のLGWに対する適用範囲の検討

項目	選択条件
Select Test Set: Select Test Set Project	Liver_training
Select Model: Select Predictive Model	PredictiveModel(Run #4)
Domain Analysis : How you would like to select the test domain	Manually select structures*

※Domain AnalysisでAutoにすると予測範囲外の化合物が出てくるので、Manualを選択した。

以下にModel構築の結果を記載した。

項目	選択条件
Select Reference Property (Optional)	LGW
Qualified Prediction: How you would like to define the property threshold:	Expert Definition: Manually define the probability threshold Select the minimum probability needed for a positive call: 0.75
	Select the maximum probability needed for a negative call: 0.25

以下に予測結果を示した。

		条件4	条件 4 (Run#4)			%	
		+	-	total	concordance	72.03	
1.004	+	122	127	249	sensitivity	49	
LGW	-	12	236	248	specificity	95.16	
				497			
予測可能な	化合物	の割合(%):	61.51			

考察:

条件4(Run#4)のモデルはLiver_conv2から構築しているため、GOT/GPTやWeight_Liverデータも含む LGWに対しては感度が非常に低くなっていると考えられる。LGWの予測には使えないことが分かった。次にLGWから構築した条件11 Run#15のモデルでLGWを予測させたときにどうなるか?を調べた。

3-6-3. 条件7のモデル(Run#15)の適用範囲の検討:

項目	選択条件
Select Test Set: Select Test Set Project	Liver_training
Select Model: Select Predictive Model	PredictiveModel(Run #15)
Domain Analysis : How you would like to select the test domain	Manually select structures*

※Domain AnalysisでAutoにすると予測範囲外の化合物が出てくるので、Manualを選択した。

以下にModel構築の結果を記載した。

判定のカットオフ値をマニュアル設定にして予測した場合:

項目	選択条件
Select Reference Property (Optional)	LGW
Qualified Prediction: How you would like to define the property threshold:	Expert Definition: Manually define the probability threshold Select the minimum probability needed for a positive call: 0.75 Select the maximum probability needed for a negative call: 0.25

予測結果:

		条件7モデル				%
		. +	-	total	concordance	93.56
LGW	+	107	8 115 sensitivity		8 115 sensitiv	93.04
LGVV	-	9	140	140 149 specificity		93.96
				264		
予測可	能な化	化合物の割	合(%):	32.67		

考察:

条件7による予測精度を計算したところ、concordance93%となり条件4による予測精度と比べ若干小さい値となった。また、適用可能な化合物数も条件4のモデルの場合よりも少ないことが分かった。Positiveと判断する基準を0.75以上から0.6以上に下げると、今度はfalse positiveが増える問題が起きることも分かった。

以下では条件4のモデルで、更に詳細に適用範囲を検討することで予測精度がどの程度改善できるのかを検討した。

3-6-4. 条件4 のモデル(Run#4)の適用範囲の検討:

<方法>

Leadscope Task Wizardで "Apply Predictive Model" をクリックした後、以下の条件でモデルを適用した。

項目 2001年1月1日日本日本日本日本日本日本日本日本日本日本日本日本日本日本日本日本日本日本	選択条件
Select Test Set: Select Test Set Project	Liver_training
Select Model: Select Predictive Model	PredictiveModel(Run #4)
Domain Analysis : How you would like to select the test domain	Manually select structures※

※Domain AnalysisでAutoにすると予測範囲外の化合物が出てくるので、Manualを選択した。

項目 選択条件 ()				
Select Reference Property (Optional)	Liver_conv2			
Qualified Prediction: How you would like to define the property threshold:	Expert Definition: Manually define the probability threshold			
,	Select the minimum probability needed for a positive call: (下記表を参照)			
	Select the maximum probability needed for a negative call: (下記表を参照)			

<結果>

以下に各カットオフに対する予測精度を以下の表にまとめた。パターン1~5については、パターン 1の予測精度が一番高いことがわかる。またパターン6~10では、パターン6が一番予測可能な範囲 が広いことが分かった。

パターン	陰性	陽性	予測可能な割合(%)	concordance	sensitivit	specificit
1	0.1以下	0.75 以上	42.45	94.17	96.72	92.76
2	0.7以下	0.75 以上				
			54.83	93.23	89.39	94.86
3	0.25 以下	0.75 以上	61.51	91.35	81.38	95.45
4	0.3 以下	0.75 以上	65.97	89.49	74.68	95.73
5	0.4 以下	0.75 以上	79.95	84.52	58.42	96.4
6	0.25 以下	0.5以上	73.64	87.06	87.08	87.05
7	0.25 以下	0.6 以上	67.7	89.58	85.08	91.8
8	0.25 以下	0.7以上	63.74	91.07	83.13	94.65
3	0.25 以下	0.75 以上	61.51	91.35	81.38	95.45
9	0.25 以下	0.8以上	60.15	91.56	80.15	96
10	0.25 以下	0.9 以上	55.32	92.84	74.53	98.53

4. 総括と今後の作業

本作業では、808化合物に対する28日間反復投与毒性の肝毒性データをトレーニングデータセットとしてQSAR構築ツールLeadscope Prediction Data Minerを用いて肝毒性予測モデルを構築した。

肝毒性に関連するデータ(LOAEL値)を変換(「-Log(数値)+4」)し、連続的な数値データから予測モデルの構築を試みたが十分な精度のモデルを構築することはできなかった。

連続データをバイナリ化 (1か0に変換) し、このバイナリ値をトレーニングデータセットとした場合には、上記808化合物に対して感度の高いモデル (条件7 Run#15)、特異性の高いモデル (条件4 Run#4) などを構築することができた。肝毒性に関連するデータとしては、病理組織学的データ以外にGOT/GPT、肝重量変化をトレーニングデータセットとして加える場合と加えない場合を検討した。条件4と条件7の各モデルを相補的に組み合わせることで高い予測精度を得ることができた。また。エデルが子別の際によりまするPolicial Picture の関係ないできた。

また、モデルが予測の際に出力するProbabilitiesの閾値を設定することにより適用範囲を制限し、 更に予測精度を高めることが可能なことが分かった。

今後の作業としては以下の方向性が考えられる。

(1) External Validation

別途評価用データセットを用意して、本作業で構築した予測モデルの予測精度を評価する。データソースとしてはLeadscope Toxicity Databaseが候補となる。既に報告書「2006年度 ClassPharmer 4.0による28日間反復投与毒性の解析」で肝毒性化合物のデータセット(Ratに対して一週間以上の経口反復投与により肝臓に病理組織学的異常が認められるデータ)を用意して ClassPharmerで見出されたToxicophoreのヒット率を算出している。このデータセットが利用可能と考えられる。

- (2) トレーニングデータセットの拡充とモデルの再構築
 - Leadscope Toxicity Databaseから類似している構造群をトレーニングデータセットとして抽出し、 改めて予測モデルを構築する。その後、Leadscope Toxicity DatabaseからValidation用データセットを抽出し、予測精度の評価を行う。
- (3) Derek for Windowsの肝毒性alert構築プロジェクトで得られているToxicophoreのScaffoldへの追加と予測モデルの再構築

Lhasa社による2006年度の肝毒性alert構築プロジェクトのfeasibility studyで得られたalertおよび 2007年度に構築された肝毒性alertの構造式をLeadscopeのScaffoldとして登録した上で改めて予測モデルの構築を行う。

Ⅲ. 研究成果の刊行に関する一覧表

雑誌

発表者氏名	論文タイトル名	発表誌名	巻号	ページ	出版年
Matsumoto M	Two-generation reproductive toxicity study of the flame retardant hexabromocyclododecane	Reprod Toxicol	in press		
	in rats				
Ema M, Fujii S, Matsumoto M, Hirata-Koizumi M,	Two-generation reproductive toxicity study of the rubber accelerator	Reprod Toxicol	25	21-38	2008
Hirose A, Kamata E	N,N-dicyclohexyl-2-benzo thiazolesulfenamide in rats.				
Yabe K, Matsumoto	Evaluation of reproductive and developmental toxicity of the rubber accelerator N,N-dicyclohexyl-2-benzo thiazole sulfenamide in rats	10000	49	149-155	2007
Hirata-Koizumi M,	Repeated dose and reproductive toxicity of the ultraviolet absorber 2-(3',5'-di- <i>tert</i> -butyl-2'-hy droxyphenyl) -5-chlorobenzotriazole in rats.		in press		
Hirata-Koizumi M,	developmental	Reprod Toxicol	25	89-99	2008
Hirata-Koizumi M, Matsuyama T, Imai		Drug Chem Toxicol	31	115-126	2008

Hirata-Koizumi M, Matsuyama T, Imai		Drug Chem	31	275-287	2008
T, Hirose A,	of	Toxicol			
Kamata E, Ema M.	2-(2'-hydroxy-3',5'-di- <i>tert</i>				
Kanata D, Dila W.	-butylphenyl)				
	benzotriazole in	17			
	preweaning rats				
Hirata-Koizumi M,		Reprod Toxicol	25	231-238	2008
Noda A, Hirose A,				251 250	2000
Kamata E, Ema M.	screening test of				
	tetrahydrofurfuryl alcohol				V.
	in rats				
Hirata-Koizumi M,	A 52-week repeated dose	Drug Chem	31	81-96	2008
Ogata H, Imai T,	toxicity study of	Toxicol			
Hirose A, Kamata	ultraviolet absorber				
E, Ema M	2-(2'-hydroxy-3',5'-di-tert-				
	butylphenyl) benzotriazole				
	in rats.				
Hirata-Koizumi M,	A 28-day repeated dose	Drug Chem	30	327-341	2007
Watari N, Mukai D,	toxicity study of	Toxicol			
Imai T, Hirose A,	ultraviolet absorber				
Kamata E, Ema M	2-(2'-hydroxy-3',5'-di- <i>tert</i>				
	-butylphenyl)benzotriazole				
	in rats				
	Non-homologous	DNA Repair.	6	781-788	2007
	end-joining for repairing				
	I-SceI-induced DNA				28
	double strand breaks in				
	human cells				
Hayashi M	Concitivity of 1	Mutat D	(EA	225 240	2007
	•	Mutat Res	654	235-240	2007
	erythrocyte micronucleus				
MacGregor JT	assay: Dependence on number of cells scored and				
wacoregor J1	inter-animal variability				
Kohayashi K Pillai	Evaluation of statistical	I Toxical Sai	33	97-104	2008
	tools used in short-term	J TOXICOI SCI	33	7/-104	2008
Abe T, Kamata E,	Section 1 to the section of the sect				
Hayashi M	administration toxicity		-		
114 543111 111	studies with rodent				
	studies with fourth				

Luan Y, Suzuki T,		Mutat Res	619	113-123	2007
	treatment predominantly				
	cuases large deletions, but				
	not GC>AT transversion				
	in human cells				
Koizumi T, Saito					
M, Matsufuji H,					
Yamagata K,					
Yamaguchi T,					
Hayashi M, Honma					
M					
Matsumoto M,	Combined repeated dose	Environ	23	169-183	2008
Furuhashi T,	and	Toxicol			
Poncipe C, Ema M	reproductive/development				
	al toxicity screening test of				
	the nitrophenolic herbicide				
	dinoseb,				
	2-sec-butyl-4,6-dinitrophe				
-	nol, in rats.				
Matsumoto M,	Potential adverse effects of	Regulat Toxicol	50	37-49	2008
Hirata-Koizumi M,	phthalic acid esters on	Pharmacol	8.		
Ema M	human health: a review of				
	recent studies on				
	reproduction				
Thybaud V,	Relevance and follow-up	Mutat Res	633	67-79	2007
Aardema M,	of positive results in in				
Casciano D,	vitro genetic toxicity				
Dellarco V, Embry	assays: an ILSI-HESI				
MR, Gollapudi BB,	initiative				
Hayashi M,					
Holsapple MP,					
Jacobson-Kram D,					
Kasper P,					
MacGregor JT,					
Rees R					
江馬 眞	有機スズ化合物の生殖	国立医薬品食	125	35-50	2007
	発生毒性	品衛生研究所			
		報告			

高橋美加、松本真	OECD 化学物質対策の動	化学生物総合	3	43-55	2007
理子、川原和三、	向 (第12報) -第20回	管理学会雑誌	-1-		
菅野誠一郎、菅谷	、第 21 回 OECD 高生産量				100
芳雄、 <u>広瀬明彦</u> 、	化学物質初期評価会議				
鎌田栄一、江馬	(2005年パリ、ワシント				11-50
眞	ンDC)				
高橋美加、松本真	OECD化学物質対策の動	国立医薬品食	125	101-106	2007
理子、川原和三、	向 (第 13 報)-第 22 回	品衛生研究所			
菅野誠一郎、菅谷	OECD高生産量化学物質	報告			
芳雄、 <u>広瀬明彦</u> 、	初期評価会議 (2006 年				
鎌田栄一、江馬	パリ)				
眞					
松本真理子、大井	OECD 高生産量化学物質	化学生物総合	3	56-65	2007
恒宏、宮地繁樹、	点検プログラム―第 23	管理学会誌			
菅谷芳雄、江馬	回初期評価会議概要				
眞					
松本真理子、山本	OECD 高生産量化学物質	化学生物総合	3	180-189	2007
展裕、宮地繁樹、	点検プログラム―第 24	管理学会誌		31 1	
菅谷芳雄、江馬	回初期評価会議概要				
眞	¥		- 1		

200736016A (2/2)

IV. 研究成果の刊行物・別刷

ARTICLE IN PRESS



Available online at www.sciencedirect.com



Reproductive Toxicology xxx (2008) xxx-xxx



Two-generation reproductive toxicity study of the flame retardant hexabromocyclododecane in rats

Makoto Ema a,*, Sakiko Fujii b, Mutsuko Hirata-Koizumi a, Mariko Matsumoto a

^a Division of Risk Assessment, Biological Safety Research Center, National Institute of Health Sciences, Tokyo 158-8501, Japan
^b Safety Research Institute for Chemical Compounds Co., Ltd., Sapporo 004-0839, Japan

Received 29 August 2007; received in revised form 14 December 2007; accepted 19 December 2007

Abstract

Male and female rats were fed a diet containing flame retardant hexabromocyclododecane (HBCD) at 0, 150, 1500 or 15,000 ppm throughout the study beginning at the onset of a 10-week pre-mating period and continuing through the mating, gestation and lactation periods for two generations. The mean daily intakes of HBCD during the whole period of administration were 10.2, 101 and 1008 mg/kg bw in F0 males, 14.0, 141 and 1363 mg/kg bw in F0 females, 11.4, 115 and 1142 mg/kg bw in F1 males, and 14.3, 138 and 1363 mg/kg bw in F1 females for 150, 1500 and 15,000 ppm, respectively. The incidence of rats with decreased thyroid follicles size was increased in F0 and F1 males and females at 1500 ppm and higher. Serum TSH levels were increased in F0 and F1 females at 1500 ppm and higher, and serum T4 levels were decreased in F0 males and females at 15,000 ppm. The number of the primordial follicles in the ovary of F1 females was reduced at 1500 ppm and higher. There were increases in the absolute and relative weights of the liver in male adults and male and female weanlings at 1500 ppm and higher, and in female adults at 15,000 ppm, and of the thyroid in male and female adults at 15,000 ppm. Decreased body weight and body weight gain associated with reduced food consumption were found in F1 males and females at 15,000 ppm. Decreases were found in the viability index of F2 pups and the body weight of male F1 and F2 pups and female F2 pups at 15,000 ppm. In F2 pups, there were low incidences of the completion of eye opening in males at 15,000 ppm and in females at 1500 ppm and higher, and of completed mid-air righting in females at 15,000 ppm. The data indicate that the NOAEL of HBCD in this study was 150 ppm (10.2 mg/kg bw/day). The estimated human intake of HBCD is well below the NOAEL in the present study.

© 2007 Elsevier Inc. All rights reserved.

Keywords: Hexabromocyclododecane; Brominated flame retardant; Two-generation reproductive toxicity; Developmental toxicity; Rat

1. Introduction

Although about 80 different brominated organic flame retardants are registered, tetrabromobisphenol A, the polybrominated diphenyl ethers and hexabromocyclododecane (HBCD) account for most of the total volume [1]. HBCD is a nonaromatic, brominated cyclic alkane used as an additive flame retardant. Total market demand for HBCD in 2001 was estimated as 2800 tons in America, 9500 tons in Europe, 3900 tons in Asia and 500 tons in the rest of the world [2]. The commercial product is a mixture of three stereoisomers, alpha, beta and gamma, which are typically present at approximately 6, 8 and 80%, respectively [3]. Its primary application is in extruded (XPS) and expanded

(EPS) polystyrene foam that is used as thermal insulation in the building industry. HBCD is the only suitable flame retardant for these applications. A secondary, although important, application of HBCD is as a flame retardant for upholstery textiles [3,4]. The partition coefficient (Log Kow) value of 5.6 suggests that this chemical is suspected to have high bioaccumulation potential [4]. HBCD has been used for about 20 years, and is detected in practically all environmental media [5]. HBCD was identified in sediment from several places along the River Viskan in Sweden [6] and the River Cinca in Spain [7]. HBCD was detected in fishes, pike (Esox lucius) [6] and barbel (Barbus graellsi) [7], indicating that it is bioavailable and bioaccumulates. The bioconcentration factor of this compound is reported to be 18,100 in fathead minnow (Pimephales promelas) [8]. HBCD was also detected from common whelk (Buccinium undatum), sea star (Asterias rubens), hermit crab (Pagurus bernhardus), gadoid fish species whiting (Merlangius merlangus), cod (Gadus morhua),

0890-6238/\$ – see front matter © 2007 Elsevier Inc. All rights reserved. doi:10.1016/j.reprotox.2007.12.004

Corresponding author. Tel.: +81 3 3700 9878; fax: +81 3 3700 1408.
 E-mail address: ema@nihs.go.jp (M. Ema).

harbor seal (*Phoca vitulina*) and harbor porpoise (*Phocoena phocoena*) from the North Sea [9]. These findings show evidence of HBCD bioaccumulation at the trophic level and biomagnification in the ascending aquatic food chain [9]. As a result of widespread use and the physical and chemical properties, HBCD is now considered to be a ubiquitous contaminant in the environment and humans [5,10]. It could be hypothesized that food intake is the largest single source of human exposure to HBCD [11].

HBCD was detected at ranging from 0.3 to 20 µg/g lipid in 49 samples of the 85 human breast milk samples collected from Norway between 1993 and 2001 [12]. The concentration of HBCD in the Stockholm human milk showed a fluctuating increase over time, and from 1980 the concentration increased from 0.13 pmol/g lipid to 0.60 pmol/g lipid in 2004 [13]. The HBCD concentration of human milks collected in 2002 to 2003 from North America was ranging from 0.3 to 10 µg/g lipid [14]. The presence of such a chemical compound in biological systems has aroused great concern about its toxicological potential. The biological effects produced by chemicals should be studied in laboratory animals to investigate their possible influences on human health, and the results of animal tests of chemical toxicity are relevant to humans [15]. The toxic effects of HBCD are briefly summarized by NRC [4], American Chemical Council [3], de Wit [16], Darnerud [11], Birnbaum and Staskal [17]. However, information on the effects of HBCD is insufficient to assess the overall toxicity of this compound. Following oral administration to male rats, HBCD was rapidly absorbed from the gastrointestinal tract, distributed primarily to the body fat, and eliminated rapidly, primarily in the feces [4]. In a 28day repeated dose toxicity study, no toxic effects were noted in male and female SD rats at any dose of HBCD given by gavage at up to 1000 mg/kg bw/day [18]. In a 90-day repeated dose toxicity study in SD rats given HBCD at 0, 100, 300, or 1000 mg/kg bw/day by gavage, increased weights of the liver and prostate, and y-glutamyltransferase, and decreased weight of the thyroid/parathyroid were found [19]. The author of this study concluded that these changes were probably of limited, if any, toxicological significance, because they were reversible, and not associated with specific target organ damage or diminished function. The dose-related effects of HBCD on the thyroid hormone axis were observed in a recent 28-day repeated dose study (OECD407) enhanced for endocrine and immune parameters using Wistar rats dosed by gavage at 0-200 mg/kg bw/day [20]. After a single dose of HBCD by gavage at 0.9 or 13.5 mg/kg bw by gavage on postnatal day (PND) 10, spontaneous activity and learning and memory in the water maze were altered when tested at the age of 3 months in NMRI mice [21]. As for the developmental toxicity of HBCD, two studies are available. There was no maternal or developmental toxicity in SD rats given HBCD by gavage on days 6-19 of pregnancy at any doses up to 1000 mg/kg bw/day [22]. No maternal or developmental toxicity was noted in Wistar rats given HBCD in diet at up to 1% (equivalent to 600 mg/kg bw/day) on days 0-20 of pregnancy [23]. No reproductive difficulties in dams or postnatal development in offspring were found even at the highest dose.

Although the testing for reproductive toxicity in an animal model is an important part of the overall toxicology, no information is available for the reproductive toxicity of HBCD at the present time; therefore, a two-generation reproductive toxicity study was conducted.

2. Materials and methods

This study was performed in 2005–2006 at the Safety Research Institute for Chemical Compounds Co., Ltd. (Sapporo, Japan) in compliance with the OECD guideline 416 Two-generation Reproduction Toxicity Study [24]. This study was conducted in accordance with the principles for Good Laboratory Practice [25], "Law for the Humane Treatment and Management of Animals" [Law No. 105, October 1, 1973, revised December 22, 1999, Revised Law No. 221; revised June 22, 2005, Revised Law No. 68], "Standards Relating to the Care, Management and Refinement of Laboratory Animals" [Notification No. 88 of the Ministry of the Environment, Japan, April 28, 2006] and "Fundamental Guidelines for Proper Conduct of Animal Experiment and Related Activities in the Testing Facility under the Jurisdiction of the Ministry of Health, Labour and Welfare" [Notification No. 0601005 of the Health Sciences Division, Ministry of Health, Labour and Welfare, Japan, June 1, 2006].

2.1. Chemical and dosing

Hexabromocyclododecane (HBCD; 1,2,5,6,9,10-hexabromocyclododecane; CAS No. 3194-55-6) was obtained from Wildlife International, Ltd. (Easton, MD). The test substance was a composite of HBCD commercial products from Albemarle Corporation (Baton Rouge, LA), Great Lakes Chemical Corporation (West Lafayette, IN) and Ameribrom Inc. (New York, NY), and Wildlife International, Ltd. prepared the composite. The preparation of HBCD was a mixture of three enantiomers. HBCD- α , HBCD- β and HBCD- γ , and their respective proportions in the used batch were 8.5, 7.9 and 83.7%. The HBCD (test substance number # 7086) used in this study was 99.7% pure, and was kept in a sealed container under cool (2–7 °C) and dark conditions. The purity and stability of the chemical were verified by analysis using liquid chromatography before and after the study.

Rats were given dietary HBCD at a concentration of 0 (control), 150, 1500 or 15,000 ppm. The dosage levels were determined based on the results of a previous 90-day oral repeated dose toxicity study [19] in male and female Crl:CD(SD)IGS BR rats given HBCD at 0, 100, 300 or 1000 mg/kg bw/day for 90 days. The author concluded that all test article-related changes, even at 1000 mg/kg bw/day, were reversible, not associated with specific target organ damage or diminished function (data not shown).

Dosed diet preparations were formulated by mixing HBCD into an appropriate amount of a powdered basal diet (CRF-1, Oriental Yeast Co., Ltd., Tokyo, Japan) for each dietary concentration. The control rats were fed a basal diet only. Analysis showed that the HBCD was homogeneous in the diet and stable for at least 21 days at room temperature, and was administered at the desired feed concentrations throughout the study.

2.2. Animals and housing conditions

Crl:CD(SD) rats were used throughout this study. Rats of this strain were chosen because they are the most commonly used in reproductive and developmental toxicity studies, and historical control data are available. Male and female rats at 4 weeks of age were purchased from Tsukuba Breeding Center, Charles River Laboratories Japan, Inc. (Yokohama, Japan). The males and females were acclimated to the laboratory for 7 days prior to the start of the experiment. Male and female rats found to be in good health were selected for use. One hundred and ninety-two rats were randomly assigned 24/sex/group as F0 animals, and all animals were assigned a unique number and ear tattooed prior to the start of the experiment. Animals were housed individually in suspended aluminum/stainless steel cages, except during the acclimation, mating and nursing periods. From day 17 of pregnancy to the day of weaning, individual dams and litters were reared using wood chips as bedding (White Flake, Charles River Laboratories Japan, Inc.).

M. Ema et al. / Reproductive Toxicology xxx (2008) xxx-xxx

on PND 18 [26]. All F1 offspring selected as F1 parents were observed daily

Animals were reared on a basal diet or diet containing HBCD and filtered tap water ad libitum and maintained in an air-conditioned room at 22 ± 3 °C, with humidity of $50 \pm 20\%$, a 12-h light/dark (20:00–08:00) cycle and ventilation at 10-15 times/h.

2.3. Experimental design

Twenty-four F0 rats (5-week-old males and females)/sex/group were fed a diet containing HBCD at 0, 150, 1500 or 15,000 ppm for 10 weeks prior to the mating period. Administration of HBCD was continued throughout the mating, gestation and lactation periods. Twenty-four male and 24 female F1 weanlings (1 male and 1 female in each litter) in each group were selected as F1 parents on PNDs 21-25 to equalize the body weights among groups. The day on which F1 parental animals were selected was designated as 0 week of dosing for the F1 generation. The administration of HBCD in the diet was not suspended during PNDs 21-25. F1 selected rats were administered HBCD in the diet of their respective formulations in the same manner as described for F0 rats. Administration of HBCD in the diet was continued throughout the mating, gestation and lactation periods. On PND 26, unselected F1 weanlings and all F2 weanlings were necropsied.

2.4. Mating procedures

Each female was mated with a single male of the same dosage group until copulation occurred or the mating period had elapsed. The mating periods for F0 and F1 animals were 3 weeks. During the mating period, daily vaginal smears were examined for the presence of sperm. The presence of sperm in the vaginal smear and/or a vaginal plug was considered as evidence of successful mating. The day of successful mating was designated as day 0 of pregnancy. F0 females that did not mate during the 3-week mating period were cohabited with another male from the same group who had been proven to copulate. For F1 matings, cohabitation of siblings was avoided.

2.5. Parental data

All adult rats were observed twice a day for clinical signs of toxicity, and body weights and food consumption were recorded weekly. For females exhibiting evidence of successful mating, body weight and food consumption of dams were recorded on days 0, 7, 14 and 20 of pregnancy and days 0, 4, 7, 14 and 21 of lactation. Daily vaginal lavage samples of each F0 and F1 female were evaluated for estrous cyclicity throughout the 2-week pre-cohabitation period and during cohabitation until evidence of copulation was detected. Females having repeated 4-6 day estrous cycles were judged to have normal estrous cycles. After weaning their pups, parental female rats were necropsied at the proestrous stage of the estrous cycle. For each female, the number of uterine implantation sites was recorded.

2.6. Litter data

Once insemination was confirmed, female rats were checked at least three times daily on days 21-25 of pregnancy to determine the time of delivery. The females were allowed to deliver spontaneously and nurse their pups until PND 21 (the day of weaning). The day on which parturition was completed by 13:00 was designated as PND 0. Total litter size and the numbers of live and dead pups were recorded, and live pups were counted, sexed, examined grossly, and individually weighed on PNDs 0, 4, 7, 14 and 21. On PND 4, litters were randomly adjusted to eight pups comprising of four males and four females. No adjustment was made for litters of fewer than eight pups. Pups were assigned a unique number and limb tattooed on PND 4.

2.7. Developmental landmarks

All F1 and F2 pups were observed for pinna unfolding on PND 3, incisor eruption on PND 11, and eye opening on PND 14. One male and one female F1 and F2 pup selected from each dam were evaluated for the surface righting reflex on PND 5, negative geotaxis reflex on PND 8, and mid-air righting reflex for male preputial separation beginning on PND 35 or female vaginal opening beginning on PND 25. Body weight of the respective F1 rats was recorded on the day of preputial separation or vaginal opening. The anogenital distance (AGD) was measured using calipers on PND 4 in all F1 and F2 pups, and the normalized value of AGD to body weight, AGD per cube root of body weight ratio, was calculated [27].

2.8. Behavioral tests

Spontaneous locomotor activity was measured with a multi-channel activity monitoring system (Supermex; Muromachi Kikai Co., Ltd., Tokyo, Japan) in 10 male and 10 female F1 rats selected from each group at 4 weeks of age. Rats were placed individually in transparent polycarbonate cages (27.6 W × 44.5 D × 20.4 H cm, CL-0108-1, CLEA Japan Inc., Tokyo, Japan), which were placed under an infrared sensor that detects thermal radiation from animals. Spontaneous motor activity was determined for 10 min intervals and for a total of 60 min.

A test in a water-filled multiple T-maze was conducted in 10 male and 10 female F1 rats selected from each group at 6 weeks of age. The apparatus was similar to that described by Biel [28]. The water temperature of the maze was kept 21-22 °C. As a preliminary swimming ability test, each rat was allowed to swim three times in a straight channel on the day before the maze trial, and then tested in the maze with three trials per day for the next three consecutive days. The elapsed time between entry into the water at the starting point and touching the goal ramp and number of errors were recorded. To prevent the exhaustion of the rats, no animal was allowed to remain in the water for more than 3 min in

2.9. Termination/necropsy adults

Parental rats were necropsied: males after the parturition of paired females, females after weaning of their pups. The proestrous stage of the estrous cycle was characterized by examination of the vaginal smears of female rats on the day of necropsy. A complete necropsy was performed on all rats found dead and those killed at the scheduled sacrifice. Live rats were euthanized by exsanguination under ether anesthesia. The external surfaces of the rats were examined. The abdomen and thoracic cavities were opened, and a gross internal examination was performed. Weights of the brain, pituitary, thyroid, thymus, liver, kidney, spleen, adrenal, testis, epididymis, seminal vesicle (with coagulating glands and their fluids), ventral prostate, uterus and ovary were recorded. Weights of the thyroid and seminal vesicle were measured after fixation. Major organs were stored in 10% neutral-buffered formalin. The testis and epididymis were fixed with Bouin's solution and preserved in 70% ethanol.

Histopathological evaluation of F0 and F1 adults was performed on the tissues specified below after fixation, paraffin embedding, and sectioning and staining with hematoxylin and eosin: the pituitary, liver, thymus, kidney, spleen, adrenal, bone marrow, mesenteric lymph node, Peyer's patches, testis, epididymis, seminal vesicle, coagulating gland, ventral prostate, ovary, uterus, vagina and mammary gland of all males and females in the control and highest dose (15,000 ppm) groups and of females with abnormal estrous cycles, males and females without evidence of copulation or insemination and females with abnormal delivery or totally dead pups in all groups. Any organs or tissues of F0 and F1 adults showing gross alterations were evaluated histopathologically. The thyroid in all rats in all groups was examined histopathologically. In ten F1 females of each group, the number of primordial follicles was counted [29]. The right ovary was fixed in 10% neutral-buffered formalin and then dehydrated and embedded in paraffin in a longitudinal orientation by routine procedures. Sections were cut serially at 5 µm and every 20th section was serially mounted on a slide and stained with hematoxylin and eosin. About 40 sections per ovary were used to determine the primordial follicles.

2.10. Termination/necropsy pups

Following the adjustment of litter size on PND 4, culled pups were euthanized by inhalation of carbon dioxide and subjected to a gross external and internal necropsy. No tissues from these pups were collected.

ARTICLE IN PRESS

M. Ema et al. / Reproductive Toxicology xxx (2008) xxx-xxx

The weanlings not selected to become parents were euthanized and necropsied as described for the adults. Organ weights of one male and one female F1 and F2 weanling selected from each dam were measured as described above for adults. The weights of the pituitary, thyroid and seminal vesicle were not determined. All pups found dead before weaning were also necropsied.

In all male and female F1 and F2 weanlings whose organs were collected, histopathological evaluations of the liver, in the control and 15,000 ppm groups, and thyroid, in all groups, were performed after fixation, paraffin embedding, and sectioning and staining with hematoxylin and eosin.

2.11. Hematological and blood biochemical parameters

On the day of the scheduled sacrifice, blood samples were collected from the abdominal aorta of adult rats under ether anesthesia.

Hematological examinations were performed for 10 males and 10 females of F0 and F1 rats randomly selected from each group. Blood samples were analyzed for the following hematological parameters, using 2K-EDTA as an anticoagulant: white blood cell (WBC) count and differential leukocyte count.

Blood biochemical evaluations were performed in 10 males and 10 females of F0 and F1 rats randomly selected from each group. Serum samples obtained from centrifuged whole blood were analyzed for biochemistry parameters such as total protein, albumin and globulin.

2.12. Serum hormone levels

On the day of the scheduled sacrifice, blood samples were collected from the abdominal aorta of adult rats. Eight males and eight proestrous females of F0 and F1 generations from each group were selected randomly for blood collection. Hormone levels were determined by Panapharm Laboratories Co., Ltd. (Uto, Japan). Serum levels of testosterone, 5α-dihydrotestosterone (DHT), luteinizing hormone (LH) and follicle stimulating hormone (FSH), thyroxine (T4), triiodothyronine (T3) and thyroid stimulating hormone (TSH) in males, and estradiol, progesterone, LH, FSH, T3, T4 and TSH in females were measured with a radioimmunoassay kit. Double antibody kits were used for measurement of testosterone, estradiol, progesterone, T3 and T4 concentration (Diagnostic Products Corp., Los Angeles, CA) and DHT concentration (Diagnostic Systems Laboratories Inc., Webster TX). Serum concentrations of LH, FSH and TSH were measured using (rat LH)[¹²⁵I], (rat FSH)[¹²⁵I] and (rat TSH)[¹²⁵I]assay systems (Amersham Biosciences Ltd., Little Chalfont, Buckinghamshire, UK), respectively.

2.13. Sperm parameters

Sperm parameters were determined for all F0 and F1 male adults on the day of the scheduled sacrifice. The right testis was used to count testicular homogenization-resistant spermatid heads. The right cauda epididymis was weighed and used for sperm analysis. Sperm motility was analyzed using a computer-assisted cell motion analyzer (TOX IVOS, Hamilton Thorne Biosciences, Beverly, MA). The percentage of motile sperm and progressively motile sperm, and the swimming speed and pattern were determined. After recording sperm motion, the cauda epididymal fluid was diluted and the sperm were enumerated using a hemacytometer under a light microscope. Sperm count per gram of epididymal tissue was obtained by dividing the total count by the gram weight of the cauda epididymis. Sperm were stained with eosin and mounted on a slide glass. Two hundred sperm in each sample were examined under a light microscope, and the percentage of morphologically abnormal sperm was calculated.

2.14. Statistical analysis

Statistical analysis was performed according to the methods of Gad [30]. Data on offspring before weaning were statistically analyzed using the litter as the experimental unit.

Body weight, body weight gain, food consumption, length of estrous cycle, pre-coital interval, gestation length, numbers of implantations and pups delivered, delivery index, sperm parameters, hematological and blood biochemical parameters, hormone levels, organ weight, organ/body weight ratio (relative

organ weight), number of primordial follicles, reflex response time, age and body weight at sexual maturation, parameters of behavioral tests, AGD, AGD/cube root of body weight ratio, and viability of pups were analyzed for statistical significance using the following method. Bartlett's test of homogeneity of variance was used to determine if the groups had equivalent variances. If the variances were equivalent, the groups were compared by one-way analysis of variance (ANOVA). If significant differences were found, Dunnett's multiple comparison test was performed. If the groups did not have equivalent variances, the Kruskal–Wallis test was used to assess the overall effects. Whenever significant differences were noted, pairwise comparisons were made by the Mann–Whitney U test.

The incidence of pups with changes in clinical and gross internal observations, and completion rate of developmental landmarks and reflexes were analyzed by the Wilcoxon rank sum test.

The incidence of parent animals with changes in clinical, gross internal and histopathological findings, the incidence of weanlings with changes in histopathological findings, the incidence of females with normal estrous cycles, the copulation index, fertility index, gestation index, neonatal sex ratio and completion rate of the reflex response test were analyzed by Fisher's exact test.

The 0.05 level of probability was significant. The probability was designated as the cut-off for statistical significance.

3. Results

3.1. Clinical observations, body weight and food consumption during the pre-mating, mating, gestation and lactation periods (F0 and F1)

One F0 male at 15,000 ppm was euthanized at 13 weeks of dosing because of a moribund condition resulting from accidental injury in the home cage. One F1 male at 1500 ppm was dead from accidental injury in the home cage. One F0 male at 15,000 ppm and one F1 male at 1500 ppm died without any apparent clinical signs of toxicity at 5 and 7 weeks of dosing, respectively. In F0 females at 15,000 ppm, one was euthanized during the pre-mating period because of a moribund condition, and one died on day 22 of pregnancy due to dystocia. No significant difference was seen between control and HBCD-treated groups in the incidence of clinical signs of toxicity in either male or female F0 and F1 rats during the pre-mating, mating, gestation, or lactation period (data not shown).

Fig. 1 shows the body weights of F0 males and females during dosing. In F0 males, the mean body weight and/or body weight gain were significantly higher than those of controls almost throughout the dosing period at 1500 ppm and in the first 5 weeks of dosing at 15,000 ppm. In F0 females, the mean body weight gain was significantly increased on days 0–4 of lactation at 150 ppm and during weeks 0–3 of dosing at 15,000 ppm compared to controls, and the mean body weight was significantly increased on week 2 of dosing at 15,000 ppm. The body weight gain was significantly decreased on days 0–14 of pregnancy at 15,000 ppm compared to controls.

Fig. 2 presents the body weights of F1 males and females during dosing. Significant decreases compared to controls were observed in the body weight during weeks 3–6 of dosing and body weight gain during the first 6 weeks of dosing in F1 males at 15,000 ppm. Compared with control group, a significantly lowered mean body weight was observed during weeks 3 and 6–10 of dosing, the whole period of gestation and days 0–14

Please cite this article in press as: Ema M, et al., Two-generation reproductive toxicity study of the flame retardant hexabromocyclododecane in rats, Reprod Toxicol (2008), doi:10.1016/j.reprotox.2007.12.004

4

M. Ema et al. / Reproductive Toxicology xxx (2008) xxx-xxx

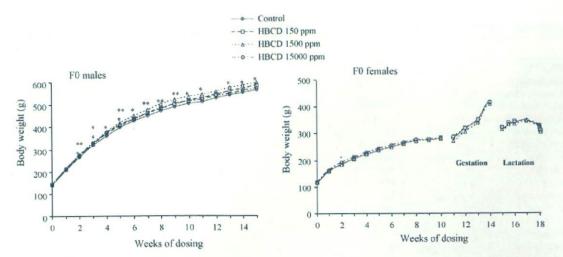


Fig. 1. Body weights of F0 male and female rats. (*) Significantly different from the control, P < 0.05. (**) Significantly different from the control, P < 0.01.

of lactation, and a significantly reduced mean body weight gain was observed during weeks 0–10 of dosing at 15,000 ppm in F1 females.

Food consumption was generally paralleled to the body weights/body weight gains during most of the study (data not shown).

The mean daily intakes of HBCD were 12.5, 125 and 1238 mg/kg bw during the pre-mating period, 9.6, 96 and 941 mg/kg bw during the gestation period, and 23.4, 240 and 2200 mg/kg bw during the lactation period in F0 females for 150, 1500 and 15,000 ppm, respectively. The mean daily intakes of HBCD were 14.0, 138 and 1365 mg/kg bw during the pre-mating period, 9.7, 100 and 995 mg/kg bw during the gestation period, and 19.6, 179 and 1724 mg/kg bw during the lactation period in F1 females for 150, 1500 and 15,000 ppm, respectively. The mean daily intakes of HBCD during the whole period were 10.2, 101 and 1008 mg/kg bw in F0 males, 14.0, 141 and 1363 mg/kg bw in F0 females,

11.4, 115 and 1142 mg/kg bw in F1 males, and 14.3, 138 and 1363 mg/kg bw in F1 females for 150, 1500 and 15,000 ppm, respectively.

3.2. Reproductive effects (F0 parents/F1 offspring and F1 parents/F2 offspring)

Table 1

presents the reproductive and developmental parameters for F0 parent/F1 offspring. HBCD produced no significant deviations in estrous cycles, although a few control and HBCD-treated rats had extended estrus or diestrus. Copulation was not observed in two males and two females at 1500 ppm and two males and one female at 15,000 ppm. Two females each at 150 and 1500 ppm did not become pregnant and three females at 15,000 ppm neither. One pregnant female each at 150 and 15,000 ppm did not deliver live pups. There were significantly longer gestation length and lower sex ratio of live pups at 1500 ppm compared

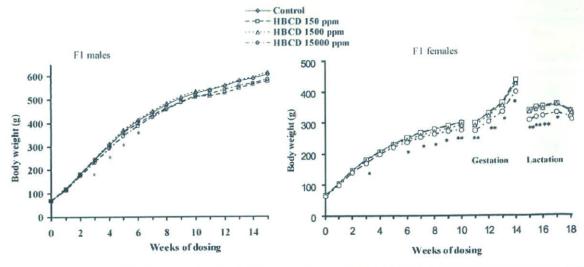


Fig. 2. Body weights of F1 male and female rats. (*) Significantly different from the control, P < 0.05. (**) Significantly different from the control, P < 0.01.

6

ARTICLE IN PRESS

M. Ema et al. / Reproductive Toxicology xxx (2008) xxx-xxx

Table 1 Reproductive and developmental findings in F0 parents/F1 offspring and F1 parents/F2 offspring

IBCD (ppm)	0 (control)	150	1500	15,000
O parents/F1 offspring				
No. of rats (male/female)	24/24	24/24	24/24	23/23
Females with normal estrous cycles (%)a	91.7	95.8	87.5	87.0
Copulation index (male/female) (%)b	100/100	100/100	91.7/91.7	91.3/95.7
Fertility index (male/female) (%) ^c	100/100	91.7/91.7	90.9/90.9	85.7/86.4
No. of pregnant females	24	22	20	19
Pre-coital interval (days) ^d	3.4 ± 3.9	3.1 ± 3.3	2.7 ± 1.4	3.5 ± 4.3
	14.2±2.1			
No. of implantations ^d Gestation index (%) ^c		13.7 ± 3.3	14.5 ± 1.4	14.5 ± 2.7
	100	95.5	100	94.7
Delivery index (%)f	92.0	89.3	90.7	93.6
Gestation length (days)d	22.1±0.3	22.3 ± 0.5	22.6 ± 0.5**	22.2 ± 0.4
No. of pups delivered ^d	13.0 ± 2.3	13.3 ± 1.7	13.3 ± 2.6	13.5 ± 2.8
No. of litters	24	21	20	18
Sex ratio of F1 pupsg	0.524	0.471	0.426*	0.572
No. of litters totally lost	0	0	0	1
Viability index during lactation (%)h,i,j				
Day 0	99.6	97.5	98.8	99.2
Day 4	95.6	98.7	98.7	95.8
Day 21	93.2	99.4	98.1	93.8
Male pup weight during lactation (g) ^d				
Day 0	6.8 ± 0.5	6.9 ± 0.6	7.2 ± 0.7	6.8 ± 0.6
Day 4	10.2 ± 1.7	10.7 ± 1.8	10.8 ± 1.6	9.5 ± 1.8
	16.4 ± 3.1			
Day 7		17.5 ± 2.4	16.9 ± 2.2	$15.6 \pm 2.0 (17)^k$
Day 14	$36.1 \pm 4.8 (23)^k$	36.3 ± 3.6	36.1 ± 3.9	$33.5 \pm 2.6 (17)^k$ $55.4 \pm 4.0 (17)^k$
Day 21	$61.1 \pm 7.1 (23)^k$	62.3 ± 6.5	61.9 ± 6.5	55.4 ± 4.0 (17)**
Female pup weight during lactation (g) ^d			Andrew Street,	
Day 0	$6.3 \pm 0.5 (23)^{k}$	6.6 ± 0.7	$6.8 \pm 0.6^{\circ}$	6.5 ± 0.7
Day 4	$9.6 \pm 1.4 (23)^{k}$	10.3 ± 1.8	10.4 ± 1.5	9.2 ± 1.6
Day 7	$15.4 \pm 2.8 (23)^{k}$	17.0 ± 2.5	16.9 ± 2.3	$15.1 \pm 1.6 (17)^{k}$
Day 14	$33.5 \pm 5.3 (23)^{k}$	35.5 ± 3.6	35.7 ± 3.6	$32.6 \pm 3.0 (17)^{k}$
Day 21	$56.5 \pm 8.0 (23)^{k}$	59.9 ± 6.4	60.5 ± 5.9	$53.2 \pm 4.7 (17)^k$
1 parents/F2 offspring				
No. of rats (male/female)	24/24	24/24	23/24	24/24
Females with normal estrous cycles (%) ^a	95.8	91.7	91.7	91.7
Copulation index (male/female) (%) ^b	100/100	100/100	100/100	100/100
Fertility index (male/female) (%) ^c	95.8/95.8	95.8/95.8	87.0/87.5	87.5/87.5
No. of pregnant females	23	23	21	21
Pre-coital interval (days)d	2.6 ± 1.6	3.4 ± 4.1	3.3 ± 3.7	2.3 ± 1.3
No. of implantations ^d	14.3 ± 2.5	14.7 ± 3.4	14.0 ± 3.2	14.3 ± 2.8
Gestation index (%) ^c	100	100	95.2	100
Delivery index (%) ^f	91.4	94.8	88.1	92.6
Gestation length (days) ^d	22.5 ± 0.5	22.4 ± 0.6	22.4 ± 0.5	22.4 ± 0.5
No. of pups delivered ^d	13.2 ± 3.4	13.9 ± 3.3	13.4 ± 2.4	13.1 ± 2.4
No. of litters	23	23	20	21
Sex ratio of F2 pupsg	0.523	0.492	0.517	0.486
No. of litters totally lost	1	1	0	8**
Viability index during lactation (%)h,i,j				
Day 0	98.6	97.7	96.0	97.8
Day 4	86.9	87.3	92.1	68.4*
Day 21	85.0 (22) ^k	89.6 (22) ^k	71.3	49.7 (20)k,**
Male pup weight during lactation (g) ^d	6.8 ± 0.8	67+07(20)k	71+06	66+06
Day 0		$6.7 \pm 0.7 (22)^k$	7.1 ± 0.6	6.6 ± 0.6
Day 4	$9.1 \pm 2.3 (22)^k$	$9.3 \pm 1.3 (22)^k$	9.0 ± 1.8	$8.0 \pm 1.3 (19)^k$
Day 7	$14.7 \pm 3.9 (22)^k$	$15.4 \pm 2.8 (22)^k$	$14.3 \pm 3.6 (19)^k$	$11.5 \pm 2.9 (17)^k$
Day 14	$31.4 \pm 8.0 (22)^k$	$33.8 \pm 5.0 (22)^{k}$	$31.0 \pm 7.2 (18)^k$	$24.2 \pm 6.6 (14)^{k}$
Day 21	$53.0 \pm 12.6 (22)^k$	$56.2 \pm 6.7 (22)^k$	$54.1 \pm 10.1 (18)^k$	$42.6 \pm 8.3 (13)^k$
Female pup weight during lactation (g)d				
I - I - O	6.5 ± 0.8	6.3 ± 0.6	6.7 ± 0.6	6.2 ± 0.6

Table 1 (Continued)

HBCD (ppm)	0 (control)	150	1500	15.000	
пвер (ррш)	o (condor)	o (condo)			
Day 4	$8.9 \pm 2.3 (22)^{k}$	$8.5 \pm 1.3 (22)^k$	8.8 ± 1.8	$7.3 \pm 1.3 (20)^{k,**}$	
Day 7	$14.3 \pm 3.5 (21)^{k}$	$14.2 \pm 2.8 (22)^{k}$	13.5 ± 3.9	$10.7 \pm 2.6 (17)^{k,*}$	
Day 14	$31.2 \pm 6.5 (21)^k$	$31.3 \pm 5.1 (22)^k$	29.3 ± 7.3	$23.9 \pm 5.9 (13)^{k,*}$	
Day 21	$52.0 \pm 10.0 (21)^{k}$	$52.8 \pm 6.6 (22)^{k}$	51.2 ± 10.8	$41.6 \pm 8.4 (13)^{k,*}$	

a Incidence of females with normal estrous cycles (%)=(no, of females with normal estrous cycles/no. of females examined) × 100.

b Copulation index (%) = (no. of animals with successful copulation/no. of animals paired) × 100.

c Fertility index (%) = (no. of animals that impregnated a female or were pregnant/no. of animals with successful copulation) x 100.

d Values are given as the mean ± S.D.

^c Gestation index (%) = (no. of females that delivered live pups/no. of pregnant females) × 100.

f Delivery index (%) = (no. of pups delivered/no. of implantations) × 100.

g Sex ratio = total no. of male pups/total no. of pups.

- h Viability index on postnatal day 0 (%) = (no. of live pups on postnatal day 0/no. of pups delivered) × 100.
- ¹ Viability index on postnatal day 4 (%) = (no. of live pups on postnatal day 4/no. of live pups on postnatal day 0) × 100.
- j Viability index on postnatal day 21 (%) = (no. of live pups on postnatal day 21/no. of live pups on postnatal day 4 after cull) x 100.
- k Data were obtained from the numbers of litters in parentheses because females that had no male and/or female pups and/or experienced total male and/or female pup loss during lactation were excluded.
- Significantly different from the control, P < 0.05.</p>
- ** Significantly different from the control, P < 0.01.

to controls. One dam experienced total litter loss by day 5 of lactation at 15,000 ppm; however, there were no significant differences in the copulation index, fertility index, gestation index, pre-coital interval, number of implantations, delivery index, number of F1 pups delivered, or viability of F1 pups during lactation between the control and HBCD-treated groups. Mean body weight of female F1 pups on PND 0 was significantly higher at 1500 ppm, and that of male F1 pups on PND 21 was significantly lowered at 15,000 ppm, compared to controls.

Table 1 also shows the reproductive and developmental parameters for F1 parent/F2 offspring. In F1 females, there were extended diestrus vaginal smears in a few control and HBCDtreated rats, but no significant effect of HBCD was found on the incidence of females with normal estrous cycles. All pairs in all groups copulated. One female each in the control and 150 ppm groups, and three females each at 1500 and 15,000 ppm were not impregnated. One pregnant female did not deliver live pups at 1500 ppm. One dam experienced total litter loss by day 4 of lactation in the control group and by day 2 of lactation at 150 ppm. At 15,000 ppm, eight dams experienced total litter loss by days 4, 5, 7, 9, 11, 13 or 18 of lactation, and a significantly increased incidence of dams with total litter loss was noted. No clear clinical signs of toxicity were noted in these dams with total litter loss. No significant changes were observed in the copulation index, fertility index, gestation index, pre-coital interval, gestation length, number of implantations, delivery index, number of F2 pups delivered or the sex ratio of F2 pups. A significantly decreased viability index was noted in F2 pups on PNDs 4 and 21 at 15,000 ppm. Mean body weights were significantly lowered compared to controls in male F2 pups on PNDs 7, 14 and 21 and in female F2 pups on PNDs 4, 7, 14 and 21 at 15,000 ppm.

3.3. Developmental landmarks (F1 and F2)

Table 2 presents physical development of F1 and F2 pups. There was no significant difference in the incidence of male and

female F1 and F2 pups that displayed pinna unfolding, or incisor eruption between the control and HBCD-treated groups. The incidence of male and female F1 pups showing completion of eye opening was increased compared to controls at 1500 ppm. In F2 pups, the incidence of pups showing eye opening was lowered compared to controls in males at 15,000 ppm and in females at 1500 and 15,000 ppm. The AGD and AGD per cube root of body weight ratio were not significantly different between control and HBCD-treated groups in male and female F1 and F2 pups.

Table 3 shows reflex ontogeny in F1 and F2 pups. All male and female F1 pups in all groups completed the surface righting reflex, negative geotaxis reflex and mid-air righting reflex. No significant changes were observed in reflex response time, except for faster response in the surface righting in males at 15,000 ppm, in F1 pups of both sexes in HBCD-treated groups. In F2 pups, a few pups failed to complete the reflex response in HBCD-treated groups, and a significantly low incidence of females completed mid-air righting was noted at 15,000 ppm; however, there was no significant difference in the incidence of male and female pups with completed response in other reflexes and in the reflex response time between control and HBCD-treated groups.

Table 4 presents data on sexual development in F1 rats. No significant differences between control and HBCD-treated groups were noted in the age at preputial separation in males or vaginal opening in females, or body weight at the age of preputial separation or vaginal opening.

3.4. Behavioral effects (F1)

Spontaneous locomotor activity for 10 min intervals and for a total of 60 min was not significantly different between control and HBCD-treated groups in male and females F1 rats (data not shown).

On the first day of the T-maze test, the pre-test swimming trials in the straight channel revealed that all male and female F1 rats in each group could swim satisfactorily, and no sigM. Ema et al. / Reproductive Toxicology xxx (2008) xxx-xxx

Table 2 Physical development in F1 and F2 pups

HBCD (ppm)	0 (control)	150	1500	15,000
F1 pups	THE THE TAX TO SEE		THE RESERVE OF THE PERSON OF T	
No. of litters examined	24	21	20	18
Pinna unfolding (%)a,b				
Male	86.0 ± 26.5	92.5 ± 16.5	93.6 ± 15.7	81.3 ± 27.9
Female	85.8 ± 29.5 (23)°	94.7 ± 14.7	97.3 ± 7.5	86.4 ± 23.8
Incisor eruption (%)a,b				
Male	$91.6 \pm 17.6 (23)^{\circ}$	96.4 ± 12.0	92.1 ± 17.0	$89.7 \pm 19.9 (17)^{c}$
Female	$94.9 \pm 11.4 (23)^{c}$	95.2 ± 10.1	92.5 ± 20.0	$92.2 \pm 15.4 (17)^{\circ}$
Eye opening (%)a,b				
Male	48.2 ± 41.5 (23)°	56.7 ± 37.9	$77.1 \pm 36.3^{\circ}$	$45.8 \pm 34.6 (17)^{c}$
Female	49.3 ± 37.8 (23)°	66.7 ± 41.3	82.9 ± 33.5°*	$54.9 \pm 41.4 (17)^{c}$
AGD ^a				
Male pup AGD (mm)	5.37 ± 0.41	5.44 ± 0.36	5.38 ± 0.32	5.20 ± 0.51
Male pup AGD/(bw ^{1/3})	2.49 ± 0.11	2.48 ± 0.10	2.44 ± 0.12	2.46 ± 0.14
Female pup AGD (mm)	$2.60 \pm 0.23 (23)^{\circ}$	2.67 ± 0.16	2.62 ± 0.18	2.57 ± 0.23
Female pup AGD/(bw ^{1/3})	$1.22 \pm 0.09 (23)^{c}$	1.23 ± 0.06	1.20 ± 0.06	1.23 ± 0.06
F2 pups				
No. of litters examined	23	22	20	21
Pinna unfolding (%)a,b				
Male	$79.9 \pm 36.4 (22)^{c}$	90.5 ± 22.8	82.1 ± 29.8	$70.1 \pm 39.2 (20)^{c}$
Female	73.6 ± 39.6	90.6 ± 22.8	81.5 ± 31.1	66.8 ± 40.9
Incisor eruption (%)a,b				
Male	86.4 ± 25.3 (22)°	92.8 ± 19.6	$97.2 \pm 11.8 (18)^{\circ}$	$86.3 \pm 27.7 (14)^{c}$
Female	$85.7 \pm 26.9 (21)^{\circ}$	90.9 ± 26.2	97.5 ± 11.2	$90.0 \pm 28.0 (15)^{c}$
Eye opening (%)a,b				
Male	$72.7 \pm 40.0 (22)^{\circ}$	62.5 ± 40.6	$47.2 \pm 44.8 (18)^{c}$	$33.9 \pm 34.7 (14)^{c,*}$
Female	$82.9 \pm 26.8 (21)^{c}$	72.7 ± 37.7	53.8 ± 40.3°	$48.1 \pm 42.0 (13)^{c,*}$
AGD ^a				
Male pup AGD (mm)	$5.12 \pm 0.54 (22)^{c}$	5.12 ± 0.41	5.04 ± 0.42	$4.84 \pm 0.39 (19)^{\circ}$
Male pup AGD/(bw ^{1/3})	$2.46 \pm 0.12 (22)^{c}$	2.44 ± 0.13	2.43 ± 0.08	$2.42 \pm 0.12 (19)^{\circ}$
Female pup AGD (mm)	$2.69 \pm 0.30 (22)^{c}$	2.71 ± 0.24	2.71 ± 0.29	2.54 ± 0.21 (20) ^c
Female pup AGD/(bw ^{1/3})	$1.30 \pm 0.07 (22)^{c}$	1.33 ± 0.09	1.32 ± 0.09	$1.32 \pm 0.06 (20)^{c}$

^a Values are given as the mean \pm S.D.

nificant changes were observed in the elapsed time to traverse the straight channel. In males, there were a significantly shorter elapsed time at 1500 and 15,000 ppm and fewer number of errors at 15,000 ppm on day 3 of the T-maze. In females, there was no significant difference in the elapsed time or number of errors of the T-maze between control and HBCD-treated groups (data not shown).

3.5. Necropsy and histopathology (F0, F1 and F2)

No compound-related gross lesions or microscopic alterations were observed in reproductive organs in male and female F0 and F1 adults showing reproductive difficulties, in male and female F0 and F1 adults of the highest dose group and in dead animals before scheduled sacrifice. There were no compoundrelated gross lesions or remarkable microscopic alterations in other tissues and organs, except for the thyroid, in male and female F0 and F1 adults.

Table 5 presents the histopathological findings in the thyroid of male and female F0 and F1 adults. Decreased size of follicles in the thyroid was found in F0 and F1 adults at 1500 ppm and higher, and in F1 females at 150 ppm as well. A significant increased incidence of rats with decreased follicle size was noted in F0 males (25%) and females (21%) and F1 females (21%) at 1500 ppm and F0 males (87%) and females (48%) and F1 males (46%) and females (54%) at 15,000 ppm, compared to controls (0%). Background incidence of decreased follicle size in the laboratory performed current study was 0% in a total of 56 males and 56 females in 6 studies (5–12/sex/study) form 1998 to 2004. Hypertrophy of the follicular cells in the thyroid was

b Incidence of animals that displayed pinna unfolding, incisor eruption or eye opening (%).

^c Data were obtained from the numbers of litters in parentheses because females that had no male and/or female pups and/or experienced total male and/or female pup loss during lactation were excluded.

^{*} Significantly different from the control, P < 0.05.

^{**} Significantly different from the control, P < 0.01.

Table 3 Reflex ontogeny in F1 and F2 pups

HBCD (ppm)	0 (control)	150	1500	15,000
F1 pups No. of pups examined (male/female)	24/23	21/21	20/20	17/17
Surface righting reflex completion rate (%) Male/female	100/100	100/100	100/100	100/100
Surface righting reflex response time (s) ^a Male Female	2.3 ± 1.1 3.1 ± 1.8	2.0 ± 0.6 2.4 ± 1.5	1.8 ± 0.5 2.9 ± 2.6	$1.6 \pm 0.3^{**}$ 2.6 ± 2.6
Negative geotaxis reflex completion rate (%) Male/female	100/100	100/100	100/100	100/100
Negative geotaxis reflex response time (s) ^a Male Female	17.7 ± 7.1 13.9 ± 6.2	$16.8 \pm 8.0 \\ 11.5 \pm 6.2$	15.2 ± 7.8 12.7 ± 6.3	$19.4 \pm 5.9 \\ 17.0 \pm 6.9$
Mid-air righting reflex completion rate (%) Male/female	100 (23)b/100	100/100	100/100	100/100
F2 pups No. of pups examined (male/female)	22/22	22/22	19/20	19/18
Surface righting reflex completion rate (%) Male/female	100/100	100/100	100/100	100/88.9
Surface righting reflex response time (s) ^a Male Female	2.1 ± 1.7 2.3 ± 0.9	2.0 ± 1.5 2.4 ± 1.7	2.8 ± 2.5 2.1 ± 0.9	2.2 ± 2.3 $3.7 \pm 3.7 (16)^{b}$
Negative geotaxis reflex completion rate (%) Male/female	100/100 (21) ^b	95.5/100	100/100	81.3 (16) ^b /88.2 (17) ^b
Negative geotaxis reflex response time (s) ^a Male Female	17.3 ± 8.6 $12.4 \pm 5.3 (21)^{b}$	$14.7 \pm 6.8 (21)^{b}$ 12.0 ± 5.2	15.2 ± 6.4 16.7 ± 6.4	$14.1 \pm 6.7 (13)^{b}$ $14.6 \pm 6.6 (15)^{b}$
Mid-air righting reflex completion rate (%) Male/female	100/100 (21) ^b	100/100	94.4 (18)b/90.0	100 (13) ^b /76.9 (13) ^b ·

Surface righting reflex on postnatal day 5 (three trials), negative geotaxis reflex on postnatal day 8 (one trial) and mid-air righting reflex on postnatal day 18 (three trials) were examined. Completion rate (%) = (no. of animals showing all positive responses of the trials/no. of animals examined) × 100.

a Values are given as the mean ± S.D.

b Data were obtained from the numbers of pups in parentheses.

* Significantly different from the control, P < 0.05.

** Significantly different from the control, P < 0.01.

also observed in F0 males at 1500 ppm and higher, and in F0 females at 1500 ppm.

Fig. 3 shows the number of the primordial follicles in the ovary of F1 females. The number of primordial follicles (mean ± S.D.) was significantly decreased at 1500

 (197.9 ± 76.9) and $15,000\,\mathrm{ppm}$ (203.4 ± 79.5) , but not at $150\,\mathrm{ppm}$ (294.2 ± 66.3) , compared to controls (316.3 ± 119.5) . The range of the background control data in the laboratory performed current study was 189.5-353.4 (mean = 295.6) in 4 studies using 10 females per study in 2005-2006.

Table 4
Sexual development in F1 males and females

HBCD (ppm)	0 (control)	150	1500	15,000
F1 rats				
Male preputial separation				
No. of males examined	24	24	24	24
Age (days) ^a	42.8 ± 1.7	41.7 ± 1.8	42.8 ± 2.2	43.7 ± 1.5
Body weight (g) ^a	225.6 ± 17.1	219.6 ± 20.0	235.0 ± 20.8	226.5 ± 16.2
Female vaginal opening				
No. of females examined	24	24	24	24
Age (days)a	30.9 ± 2.0	30.3 ± 2.6	30.1 ± 1.8	30.8 ± 2.2
Body weight (g) ^a	106.0 ± 13.8	102.9 ± 13.8	106.0 ± 10.6	100.7 ± 13.0

a Values are given as the mean ± S.D.