

substrate-specific components is generally evoked by substrate modifications²⁻⁴. However, the recognition and subsequent ubiquitination of sex steroid receptors by AhR requires dioxin-type compounds as ligands but does not require the phosphorylation or ligand binding of sex steroid receptors. We have therefore shown that fat-soluble ligands directly control the function of a ubiquitin ligase complex for targeted protein destruction in animals (see Supplementary Fig. S1). In plants, auxin was recently found to control protein destruction through the auxin receptor SCF^{TIR1} (refs 29, 30). However, whereas SCF^{TIR1} is regulated by ligand-dependent substrate recognition by TIR1, CUL4B^{AhR} is primarily regulated by the assembly of a ligand-dependent complex as well as substrate recognition. Considered together, ubiquitin-ligase-based perception mechanisms of fat-soluble ligands may be diverse in different species. It is possible that other nuclear receptors and binding proteins for fat-soluble ligands also serve as key components of ubiquitin ligases to mediate a non-genomic pathway of fat-soluble ligands to regulate target-protein-selective destruction.

METHODS

More detailed descriptions of all materials and methods are supplied in the Supplementary Information.

Biochemical purification and separation of AhR-associated complexes. The nuclear extracts preparation, anti-Flag affinity purification and mass spectrometry were performed as described previously^{15,20}. For purification of the core CUL4B^{AhR} complex, the nuclear extracts were first bound to the GST-CUL4B-N (amino acid residues 1-318) columns before being loaded on anti-Flag columns²⁰.

In vitro ubiquitination assay. The *in vitro* ubiquitination assay was performed as described previously²³. Purified Flag-AhR (0.2 µg) was incubated either with 3MC (10 µM) or vehicle (dimethylsulphoxide) for 30 min at 25 °C, then mixed with Flag-CUL4B/DDB1/Rbx1 complex (0.2 µg), and after further incubation for 30 min at 25 °C the substrate, ER-α (Calbiochem), was added.

Plasmids, antibodies, immunoprecipitation, in vivo ubiquitination, pulse-chasing, ligand responses in mice, and RNA-mediated interference experiments. Detailed methods used in this study can be found in the Supplementary Information.

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- McKenna, N. J. & O'Malley, B. W. Combinatorial control of gene expression by nuclear receptors and coregulators. *Cell* **108**, 465-474 (2002).
- Hershko, A. & Ciechanover, A. The ubiquitin system. *Annu. Rev. Biochem.* **67**, 425-479 (1998).
- Deshaies, R. J. SCF and Cullin/Ring H2-based ubiquitin ligases. *Annu. Rev. Cell Dev. Biol.* **15**, 435-467 (1999).
- Harper, J. W. A phosphorylation-driven ubiquitination switch for cell-cycle control. *Trends Cell Biol.* **12**, 104-107 (2002).
- Poellinger, L. Mechanistic aspects—the dioxin (aryl hydrocarbon) receptor. *Food Addit. Contam.* **17**, 261-266 (2000).
- Hankinson, O. The aryl hydrocarbon receptor complex. *Annu. Rev. Pharmacol. Toxicol.* **35**, 307-340 (1995).
- Swanson, H. I. & Bradfield, C. A. The Ah-receptor: genetics, structure and function. *Pharmacogenetics* **3**, 213-230 (1993).
- Carlson, D. B. & Perdew, G. H. A dynamic role for the Ah receptor in cell signaling? Insights from a diverse group of Ah receptor interacting proteins. *J. Biochem. Mol. Toxicol.* **16**, 317-325 (2002).
- Mimura, J. & Fujii-Kuriyama, Y. Functional role of AhR in the expression of toxic effects by TCDD. *Biochim. Biophys. Acta* **1619**, 263-268 (2003).
- Lin, T. M. *et al.* Effects of aryl hydrocarbon receptor null mutation and in utero and lactational 2,3,7,8-tetrachlorodibenzo-p-dioxin exposure on prostate and seminal vesicle development in C57BL/6 mice. *Toxicol. Sci.* **68**, 479-487 (2002).
- Brunnberg, S. *et al.* The basic helix-loop-helix-PAS protein ARNT functions as a potent coactivator of estrogen receptor-dependent transcription. *Proc. Natl Acad. Sci. USA* **100**, 6517-6522 (2003).
- Matthews, J., Wihlen, B., Thomsen, J. & Gustafsson, J. A. Aryl hydrocarbon receptor-mediated transcription: ligand-dependent recruitment of estrogen receptor α to 2,3,7,8-tetrachlorodibenzo-p-dioxin-responsive promoters. *Mol. Cell. Biol.* **25**, 5317-5328 (2005).
- Beischlag, T. V. & Perdew, G. H. ER α-AHR-ARNT protein-protein interactions mediate estradiol-dependent transrepression of dioxin-inducible gene transcription. *J. Biol. Chem.* **280**, 21607-21611 (2005).
- Baba, T. *et al.* Intrinsic function of the aryl hydrocarbon (dioxin) receptor as a key factor in female reproduction. *Mol. Cell. Biol.* **25**, 10040-10051 (2005).
- Ohtake, F. *et al.* Modulation of oestrogen receptor signalling by association with the activated dioxin receptor. *Nature* **423**, 545-550 (2003).
- Romkes, M., Piskorska-Pliszczynska, J. & Safe, S. Effects of 2,3,7,8-tetrachlorodibenzo-p-dioxin on hepatic and uterine estrogen receptor levels in rats. *Toxicol. Appl. Pharmacol.* **87**, 306-314 (1987).
- Davarinos, N. A. & Pollenz, R. S. Aryl hydrocarbon receptor imported into the nucleus following ligand binding is rapidly degraded via the cytoplasmic proteasome following nuclear export. *J. Biol. Chem.* **274**, 28708-28715 (1999).
- Roberts, B. J. & Whitelaw, M. L. Degradation of the basic helix-loop-helix/Per-ARNT-Sim homology domain dioxin receptor via the ubiquitin/proteasome pathway. *J. Biol. Chem.* **274**, 36351-36356 (1999).
- Maxwell, P. H. *et al.* The tumour suppressor protein VHL targets hypoxia-inducible factors for oxygen-dependent proteolysis. *Nature* **399**, 271-275 (1999).
- Kitagawa, H. *et al.* The chromatin-remodeling complex WINAC targets a nuclear receptor to promoters and is impaired in Williams syndrome. *Cell* **113**, 905-917 (2003).
- Zhong, W., Feng, H., Santiago, F. E. & Kipreos, E. T. CUL-4 ubiquitin ligase maintains genome stability by restraining DNA-replication licensing. *Nature* **423**, 885-889 (2003).
- Higa, L. A. *et al.* CUL4-DDB1 ubiquitin ligase interacts with multiple WD40-repeat proteins and regulates histone methylation. *Nature Cell Biol.* **8**, 1277-1283 (2006).
- Groisman, R. *et al.* The ubiquitin ligase activity in the DDB2 and CSA complexes is differentially regulated by the COP9 signalosome in response to DNA damage. *Cell* **113**, 357-367 (2003).
- Wertz, I. E. *et al.* Human De-etiolated-1 regulates c-Jun by assembling a CUL4A ubiquitin ligase. *Science* **303**, 1371-1374 (2004).
- Jin, J., Arias, E. E., Chen, J., Harper, J. W. & Walter, J. C. A family of diverse Cul4-Ddb1-interacting proteins includes Cdt2, which is required for S phase destruction of the replication factor Cdt1. *Mol. Cell* **23**, 709-721 (2006).
- Angers, S. *et al.* Molecular architecture and assembly of the DDB1-CUL4A ubiquitin ligase machinery. *Nature* **443**, 590-593 (2006).
- He, Y. J., McCall, C. M., Hu, J., Zeng, Y. & Xiong, Y. DDB1 functions as a linker to recruit receptor WD40 proteins to CUL4-ROC1 ubiquitin ligases. *Genes Dev.* **20**, 2949-2954 (2006).
- Valley, C. C. *et al.* Differential regulation of estrogen-inducible proteolysis and transcription by the estrogen receptor alpha N terminus. *Mol. Cell. Biol.* **25**, 5417-5428 (2005).
- Dharmasiri, N., Dharmasiri, S. & Estelle, M. The F-box protein TIR1 is an auxin receptor. *Nature* **435**, 441-445 (2005).
- Kepinski, S. & Leyser, O. The *Arabidopsis* F-box protein TIR1 is an auxin receptor. *Nature* **435**, 446-451 (2005).

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Developmental toxicity of dibutyltin dichloride in cynomolgus monkeys

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Abstract

Dibutyltin dichloride (DBTCl) has been shown to be teratogenic in rats. The present study was conducted to determine the teratogenic potential of DBTCl given to pregnant monkeys during the entire period of organogenesis. Cynomolgus monkeys were dosed once daily by nasogastric intubation with DBTCl at 0, 2.5 or 3.8 mg/kg on days 20–50 of pregnancy, the whole period of organogenesis. The pregnancy outcome was determined on day 100 of pregnancy. In both DBTCl-treated groups, a significant increase in the incidence of pregnant females with soft stool and/or diarrhea, and with yellowish stool was observed. Maternal body weight gain at 3.8 mg/kg and food consumption at 2.5 and 3.8 mg/kg were decreased during the administration period. The survival rate of fetuses at terminal cesarean sectioning was decreased in the DBTCl-treated groups and significantly decreased at 2.5 mg/kg. There were no changes in the developmental parameters of surviving fetuses, including fetal body weight, crown-rump length, tail length, sex ratio, anogenital distance and placental weight, in the DBTCl-treated groups. No external, internal or skeletal malformations were found in the fetuses in any group. Although internal and skeletal variations were found, no difference in the incidence of fetal variation was noted between the control and DBTCl-treated groups. No effect on skeletal ossification was observed in fetuses in the DBTCl-treated groups. The data demonstrate that DBTCl is embryolethal but not teratogenic in cynomolgus monkeys.

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1. Introduction

Organotin compounds are widely used in agriculture and industry. The most important non-pesticidal route of entry of organotin compounds into the environment is through the leaching of organotin-stabilized polyvinyl chloride (PVC) by water [1], and its use in antifouling agents, resulting in the entry of organotin into the aquatic environment [2]. Disubstituted organotin compounds are commercially the most important derivatives, being used as heat and light stabilizers for PVC plastics to prevent degradation of the polymer during melting and the forming of the resin into its final products, as catalysts in the production of polyurethane foams, and as vulcanizing agents for silicone rubbers [3,4]. The identification of dibutyltin (DBT) and tributyltin (TBT) in aquatic marine organisms [5,6] and marine

products [7] has been reported. TBT is degraded spontaneously and biochemically via a debutylation pathway to DBT in the environment [8,9]. Organotin compounds are introduced into foods by the use of pesticides and antifoulants and via the migration of tin from PVC materials [4].

We previously demonstrated that tributyltin chloride (TBTCl) during early pregnancy caused early embryonic loss [10–12], and TBTCl on days 10–12 and on days 13–15, but not on days 7–9 of pregnancy, produced fetal malformations in rats [13]. The predominant malformation induced by TBTCl was cleft palate [13,14]. It has been reported that TBT is metabolized to DBT and MBT, and DBT was metabolized to monobutyltin (MBT) [15–17]. DBT is also reported to have toxic effects on reproduction and development in rats [18]. The oral administration of dibutyltin dichloride (DBTCl) during early pregnancy caused early embryonic loss in rats [19–21]. The oral administration of DBTCl to rats throughout the period of organogenesis resulted in a significant increase in the incidence of fetuses with malformations [22], and rat embryos were highly susceptible to the

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teratogenic effects of DBTCl when it was administered on day 7 and 8 of pregnancy [23]. Dibutyltin diacetate (DBTA) [24–28], dibutyltin maleate, dibutyltin oxide, and dibutyltin dilaurate [26] were teratogenic in rats when administered orally. Developmental toxicity studies on butyltins suggest that the teratogenicity of DBT is different from that of tetrabutyltin (TeBT), TBT and MBT in its mode of action because the period of susceptibility to teratogenicity and the types of malformations induced by DBT are different from those induced by TeBT, TBT and MBT [29,30]. DBTCl had dysmorphic effects in rat embryos in a whole embryo culture system [31,32]. DBT was detected in rat maternal blood at 100 ng/g and embryos at 720 ng/g at 24 h after gavage of DBTA at 22 mg/kg on day 8 of pregnancy [27]. The dysmorphic concentrations of DBTCl in cultured embryos were within the range of levels detected in maternal blood after the administration of a teratogenic dose of DBT. These findings suggest that DBT itself is a causative agent in DBT teratogenesis, which may be due to direct interference with embryos.

As described above, the teratogenic effects of organotin compounds, including DBT, were extensively investigated in rodents [18]. No reports on the assessment of the teratogenicity of DBT in any other species are available. It appears that conclusive evidence in support of the teratogenicity of DBT is still lacking,

because the teratogenicity of DBT only has been reported in a single animal species. Studies in non-rodents would be of great value in estimating the teratogenicity of DBT in humans. The present study was conducted to determine the teratogenic potential of DBTCl given to pregnant cynomolgus monkeys during the entire period of organogenesis.

2. Materials and methods

2.1. Animals

Cynomolgus monkeys (*Macaca fascicularis*) were used in this study. The monkeys were obtained from Guangxi Primate Center of China (Guangxi, China) through Guangdong Scientific Instruments and Materials Import/Export Co. (Guangzhou, China). The monkeys were quarantined for 4 weeks, and confirmed to be free from tuberculosis, *Salmonella* and *Shigera*. The animals were maintained in an air-conditioned room at 23.0–29.0 °C, with a relative humidity of 45–58%, under a controlled 12/12 light/dark cycle, with a ventilation rate of 15 air changes/hour, and were housed individually, except during the mating period. The monkeys were fed 108 g/day of diet (Teklad global 25% protein primate diet; Harlan Sprague-Dawley Inc., Madison, USA) and tap water ad libitum from automatic lixit devices. Healthy male and female monkeys were selected for use. Only females showing 25–32 days menstrual cycles were used in the experiment. Each female monkey was paired with a male of proven fertility for three consecutive days between days 11–15 of the menstrual cycle. The visual confirmation of copulation and/or the presence of sperm in the vagina were considered evidence of successful mating. When copulation was confirmed, the

Table 1
Maternal findings in monkeys given DBTCl on days 20–50 of pregnancy

	Dose (mg/kg)		
	0 (control)	2.5	3.8
Number of pregnant females	12	12	10
Number of females showing toxicological signs			
Death	0	0	0
Soft stool/diarrhea	1	12*	10*
Yellowish stool	0	8*	8*
Vomiting	0	3	3
Initial body weight	3.53 ± 0.59	3.49 ± 0.43	3.79 ± 0.36
Body weight gain during pregnancy (g) ^a			
Days 0–20	76 ± 114	42 ± 160	73 ± 142
Days 20–51	57 ± 237	−242 ± 423	−556 ± 526*
Days 51–100	710 ± 162	755 ± 174	848 ± 263
Food consumption during pregnancy (g/day) ^a			
Days 20–21	99 ± 18	93 ± 23	76 ± 33
Days 23–24	91 ± 27	71 ± 31	55 ± 31*
Days 27–28	77 ± 28	47 ± 19*	37 ± 34*
Days 30–31	63 ± 32	33 ± 15*	22 ± 10*
Days 34–35	88 ± 25	53 ± 42	23 ± 17*
Days 37–38	86 ± 28	53 ± 42*	25 ± 24*
Days 41–42	87 ± 27	59 ± 59	36 ± 29*
Days 44–45	95 ± 22	62 ± 40	41 ± 31*
Days 48–49	98 ± 18	70 ± 48	59 ± 44
Days 51–52	94 ± 20	97 ± 24	71 ± 39
Days 55–56	102 ± 12	107 ± 2	100 ± 20
Days 58–59	106 ± 7	108 ± 0	104 ± 10
Days 62–63	106 ± 7	108 ± 0	106 ± 5
Days 80–81	108 ± 0	108 ± 0	108 ± 0
Days 90–91	106 ± 7	108 ± 0	108 ± 0
Days 99–100	108 ± 0	108 ± 0	108 ± 0

^a Values are given as the mean ± S.D.

* Significantly different from the control, $p < 0.05$.

median day of the mating period was regarded as day 0 of pregnancy. Pregnancy was confirmed on day 18 or 19 of pregnancy by ultrasound (SSD-4000, Aloka Co., Mitaka, Japan) under anesthesia induced by intramuscular injection of 5% ketamine hydrochloride (Sigma Chemical Co., St. Louis, USA). Pregnant females, weighing 2.51–4.50 kg on day 0 of pregnancy, were allocated randomly to three groups, each of 10–12 monkeys, and housed individually. Animal experiments were performed at Shin Nippon Biomedical Laboratories, Ltd. (SNBL; Kagoshima, Japan) during 2004–2005 in compliance with the Guideline for Animal Experimentation (1987) [33], and in accordance with the Law Concerning the Protection and Control of Animals (1973) [34] and the Standards Relating to the Care and Management of Experimental Animals (1980) [35]. This study has been approved by the Institutional Animal Care and Use Committee of SNBL and performed in accordance with the ethics criteria contained in the bylaws of the committee of SNBL.

2.2. Dosing

The monkeys were dosed once daily with DBTCI (lot no. GG01, 98% pure, Tokyo Kasei Kogyo Co., Ltd., Tokyo, Japan) at 0, 2.5 or 3.8 mg/kg by nasogastric intubation on days 20–50 of pregnancy, i.e., the entire period of organogenesis [36]. Dosing was terminated in the dams in which embryonic/fetal loss occurred. The dosage levels were determined from the results of previous studies in rats, in which DBTCI administered by gavage at 7.6 or 15.2 mg/kg on days 0–3 and days 4–7 of pregnancy caused significant increases in pre- and/or post-implantation embryonic loss in rats [19–21], and in which DBTCI by gavage at 5, 7.5 or 10.0 mg/kg throughout the period of organogenesis resulted in a significant increase in the incidence of fetuses with malformations [22]. DBTCI was dissolved in olive oil (Wako Pure Chemical Industries, Ltd., Osaka, Japan). The dose volume was adjusted to 0.5 ml/kg of the most recent body weight. The control monkeys received olive oil only.

2.3. Observations

The pregnant monkeys were observed for clinical signs of toxicity twice a day during the administration period and once a day during the non-administration

period. The body weight was recorded on days 0, 20, 27, 34, 41, 51, 60, 70, 80, 90 and 100 of pregnancy. The food consumption was recorded on days 20, 23, 27, 30, 34, 37, 41, 44, 48, 51, 55, 58, 62, 80 and 90 of pregnancy. Embryonic/fetal heart-beat and growth were monitored using ultrasound under anesthesia induced by intramuscular injection of 5% ketamine hydrochloride on days 25, 30, 35, 40, 50, 60, 70, 80, 90 and 99 of pregnancy. In the dams in which embryonic/fetal cardiac arrest was confirmed by ultrasound, necropsy was performed under anesthesia induced by intraperitoneal injection of pentobarbital Na (Tokyo Kasei Kogyo Co., Ltd., Tokyo, Japan). The uterus, including the embryo/fetus and placenta and ovaries, was removed from the maternal body and stored in 10% neutral buffered formalin. Dead or aborted embryos/fetuses were morphologically examined.

Terminal cesarean sectioning was performed on day 100 of pregnancy, under anesthesia induced by intramuscular injection of 5% ketamine hydrochloride (0.1–0.2 ml/kg) and inhalation of isoflurane (0.5–2.0%, Dainippon Pharmaceutical Co. Ltd., Osaka, Japan), and contraction was induced with atropine (0.01 mg/kg, Tanabe Seiyaku Co. Ltd., Osaka, Japan). The fetus and placenta were removed from the dams. The placenta was weighed and stored in 10% neutral buffered formalin. Dams that underwent cesarean sectioning were not necropsied.

Fetal viability was recorded, and the fetuses were anesthetized by intraperitoneal injection of pentobarbital Na and euthanized by submersion in saline for 30–40 min at room temperature. Fetuses were sexed and examined for external anomalies after confirmation of the arrested heart-beat. Fetal and placental weights were recorded. The head width, tail length, crown-rump length, chest circumference, paw and foot length, distance between the eyes, umbilical cord length, volume of amniotic fluid and diameters of the primary and secondary placentae were measured. After the completion of external examinations, fetuses were examined for internal anomalies. The peritoneal cavity was opened and the organs were grossly examined. The brain, thymus, heart, lung, spleen, liver, kidneys, adrenal glands and testes/uterus and ovaries were weighed and stored in 10% neutral buffered formalin. The eyeballs, stomach, small and large intestine, head skin and auricles were stored in 10% neutral buffered formalin. Fetal carcasses were fixed in alcohol, stained with alizarin red S [37] and examined for skeletal anomalies. The number of ossification centers of the vertebral column, and lengths of the ossified parts of the humerus, radius, ulna, femur, tibia and fibula were recorded. Histopathological evaluations were performed on single

Table 2
Reproductive and developmental findings in monkeys given DBTCI on days 20–50 of pregnancy

	Dose (mg/kg)		
	0 (control)	2.5	3.8
Number of pregnant females	12	12	10
Number of females with embryonic/fetal loss	1	8*	4
Number of females with live fetuses until terminal cesarean section	11	4*	6
Number of live fetuses at terminal cesarean section	11	4*	6
Sex ratio of live fetuses (male/female)	6/5	1/3	3/3
Body weight of live fetuses (g)			
Male	133 ± 13	125	112 ± 24
Female	118 ± 12	108 ± 20	118 ± 13
Anogenital distance (cm) ^a			
Male	2.0 ± 0.2	1.9	1.7 ± 0.4
Female	1.0 ± 0.1	1.0 ± 0.2	1.0 ± 0.1
Crown-rump length (cm) ^a			
Males	12.8 ± 0.6	12.4	12.4 ± 0.7
Female	12.6 ± 0.4	12.3 ± 0.5	12.6 ± 0.1
Tail length (cm) ^a			
Male	11.8 ± 1.2	11.8	11.4 ± 0.7
Female	11.9 ± 0.8	11.7 ± 1.7	12.4 ± 0.6
Placental weight (g) ^a	42.4 ± 7.2	38.9 ± 6.2	37.5 ± 9.1
Number of a single placenta	1	1	3

^a Values are given as the mean ± S.D.

* Significantly different from the control, $p < 0.05$.

placentas and accessory spleens after fixation, paraffin embedding, sectioning and staining with hematoxylin and eosin.

2.4. Analysis of plasma steroids hormone levels

Blood samples were collected from the femoral vein on day 51 of pregnancy, 24 h after the last administration of DBTCI. The plasma was separated and stored at -80°C for the later assay of steroid hormones. Plasma progesterone and 17β -estradiol were measured by Teizo Medical Co. Ltd. (Kawasaki, Japan) using liquid chromatography-electrospray ionization Tandem Mass Spectrometry (LC-MS/MS, Applied Biosystems/MDS SCIEX). The detection limits of plasma progesterone and 17β -estradiol were 10.0 pg/ml and 0.25 pg/ml, respectively. The intra- and inter-assay coefficients of variation for 17β -estradiol were below 6.4 and 8.9%, respectively. The intra- and inter-assay coefficients of variation for progesterone were below 9.0 and 7.9%, respectively.

2.5. Data analysis

The data was analyzed by MUSCOT statistical analysis software (Yukums Co. Ltd., Tokyo, Japan) using the dam or fetus as the experimental unit [38]. Data were analyzed using Bartlett's test [39] for the homogeneity of variance. When the variance was homogeneous, Dunnett's test [40] was performed to compare the mean value in the control group with that in each DBTCI group. When the variance was heterogeneous, the data were rank-converted and a Dunnett-type test [41] was performed to compare the mean value in the control group with that in each DBTCI group. The incidences of maternal and embryonic/fetal deaths and anomalous fetuses were analyzed by Fisher's exact test. The 0.05 level of probability was used as the criterion for significance.

3. Results

Table 1 presents maternal findings in monkeys given DBTCI on days 20–50 of pregnancy. No maternal death occurred in any group. In both DBTCI-treated groups, a significant increase in the incidence of females with soft stool and/or diarrhea, and with

yellowish stool was observed. Soft stool and/or diarrhea were observed in one of the 12 females in the control group and in all females of the DBTCI-treated groups. In both groups treated with DBTCI, yellowish stool was noted in eight females and vomiting was observed in three females. Body weight gain on days 0–20, during the pre-administration period, did not significantly differ among the groups. Body weight gain on days 20–50, during the administration period, was lower in the DBTCI-treated groups, and significantly decreased at 3.8 mg/kg. No significant decrease in body weight gain on days 51–100, during the post-administration period, was found in the DBTCI-treated groups. Food consumption during the administration period was significantly reduced at 2.5 mg/kg and higher. Relatively marked decreases in the body weight gain and food consumption were observed in dams showing abortion or embryonic/fetal death.

The reproductive and developmental findings in monkeys given DBTCI on days 20–50 of pregnancy are shown in Table 2. The incidence of females with embryonic/fetal loss was increased in the DBTCI-treated groups, and a significant difference was noted at 2.5 mg/kg. Embryonic/fetal loss was observed in one of the 12 females in the control group, eight of the 12 females in the 2.5 mg/kg group and four of the 10 females in the 3.8 mg/kg group. Abortion occurred on day 30 of pregnancy in the control group, and on day 35, 44, 46, 49 or 60 of pregnancy at 2.5 mg/kg. Embryonic/fetal death was found on day 35, 40 or 64 of pregnancy at 2.5 mg/kg, and on days 38, 40 or 50 (two embryos) of pregnancy at 3.8 mg/kg. External examinations was performed in five of the eight embryonic/fetal losses at 2.5 mg/kg and four of the four embryonic/fetal losses at 3.8 mg/kg, and no anomalies were detected. Eleven, four and six females in the control, 2.5 and 3.8 mg/kg groups, respectively,

Table 3
Morphological findings in fetuses of monkeys given DBTCI on days 20–50 of pregnancy

	Dose (mg/kg)		
	0 (control)	2.5	3.8
Number of fetuses examined	11	4	6
External examination			
Number of fetuses with malformations	0	0	0
Internal examination			
Number of fetuses with malformations	0	0	0
Number of fetuses with variations	0	0	1
Accessory spleen	0	0	1
Skeletal examination			
Number of fetuses with malformations	0	0	0
Number of fetuses with variations	0	1	1
Short supernumerary rib	0	1	1
Degree of ossification ^a			
Number of ossified centers of vertebral column	53.6 ± 0.8	53.0 ± 1.2	54.2 ± 1.0
Skeletal length (mm) ^a			
Humerus	23.6 ± 0.8	23.3 ± 1.3	23.6 ± 1.2
Radius	23.0 ± 1.0	22.3 ± 1.6	23.1 ± 1.7
Ulna	24.6 ± 1.0	23.9 ± 1.5	24.3 ± 2.2
Femur	22.3 ± 1.2	21.8 ± 1.3	22.7 ± 1.6
Tibia	21.5 ± 1.3	20.5 ± 1.7	21.7 ± 1.4
Fibula	19.8 ± 1.0	19.0 ± 1.8	19.9 ± 1.6

^a Values are given as the mean ± S.D.

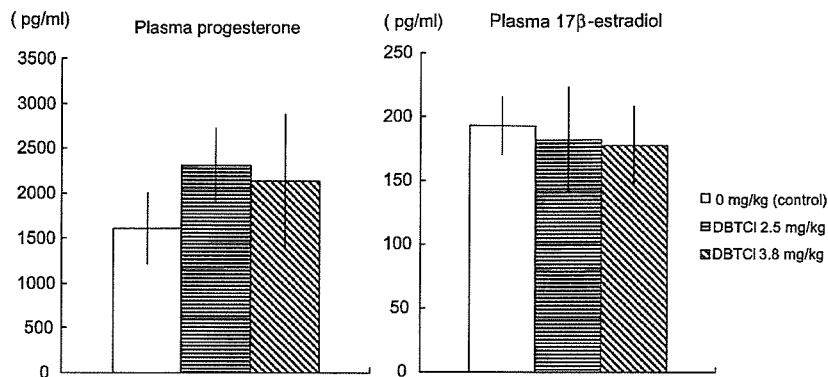


Fig. 1. Plasma progesterone and 17 β -estradiol levels in pregnant monkeys given DBTCl on days 20–50 of pregnancy. Blood samples were collected on day 51 of pregnancy, 24h after the last administration of DBTCl. Values are given as the mean \pm S.E.M. of 5–10 monkeys.

had live fetuses at terminal cesarean sectioning. There were no significant differences between the control and DBTCl-treated groups in parameters of fetal growth, such as body weight, crown-rump length and tail length. No significant differences in the head width, chest circumference, paw and foot length, distance between the eyes, umbilical cord length, volume of amniotic fluid and diameters of the primary and secondary placentae were also noted between the control and DBTCl-treated groups (data not shown). No significant differences between the control and DBTCl-treated groups were found in the sex ratio of live fetuses, anogenital distance or placental weight. A single placenta was observed in one dam in the control group, one dam in the 2.5 mg/kg group and three dams in the 3.8 mg/kg group.

Table 3 shows the morphological changes in fetuses of monkeys given DBTCl on days 20–50 of pregnancy. No external, internal or skeletal malformations were found in fetuses in any group. Although internal and skeletal examinations revealed one fetus with an accessory spleen at 3.8 mg/kg, and one fetus with a short supernumerary rib at both 2.5 and 3.8 mg/kg, no difference in the incidence of fetuses with variation was noted between the control and DBTCl-treated groups. There were no differences between the control and DBTCl-treated groups in the number of ossified centers of the vertebral column or length of the humerus, radius, ulna, femur, tibia or fibula.

Although a significant decrease in the absolute weight of the brain and lung, and increase in the relative weight of the spleen were observed in male fetuses at 3.8 mg/kg, no significant difference in the relative weight of the brain and lung or in absolute weight of the spleen was detected between the control and DBTCl-treated groups. There were no differences in absolute and relative weights of the fetal thymus, heart, lung, liver, kidneys, adrenal glands or testes/uterus and ovaries between the control and DBTCl-treated groups (data not shown). Histopathological examinations revealed no abnormalities in single placenta and accessory spleen, and the histological structures of single placenta and accessory spleen were similar to those of normal placenta and spleen.

Plasma progesterone and 17 β -estradiol levels are shown in Fig. 1. Although higher levels of plasma progesterone were observed in the DBTCl-treated groups, no statistically significant difference was noted between the control and DBTCl-

treated groups. There were no significant differences in the plasma 17 β -estradiol levels between the control and DBTCl-treated groups.

4. Discussion

In previous studies, the teratogenic effects of DBT were investigated in rats. The teratogenicity of DBT should be studied using other animal species to gain a better understanding of the developmental toxicity of butyltins. Non-human primates appear to provide an especially appropriate model for teratogenicity testing because of their high ranking on the evolutionary scale [42]. The close phylogenetic relatedness of old world monkeys to humans appears to render them most desirable as models in teratology studies [43]. The similarities in placentation and embryonic development indicate considerable value in the use of monkeys for investigating the developmental toxicity of chemicals [44]. In the present study, we determined the developmental toxicity, particularly the teratogenicity, of DBTCl in monkeys after administration over the entire period of organogenesis.

The doses of DBTCl set in the present study were expected to induce maternal toxicity, such as decreases in maternal body weight gain and food consumption, and were given to monkeys during organogenesis to characterize the effects of DBTCl on embryonic/fetal development. Toxicological sign, as evidenced by the significant increase in the incidence of pregnant females showing soft stool/diarrhea and yellowish stool, was found at 2.5 and 3.8 mg/kg. A significant decrease in the maternal body weight gain accompanied by significantly reduced food consumption was noted at 3.8 mg/kg. A significant decrease in food consumption was also found at 2.5 mg/kg. These maternal findings indicate that more severe adverse effects on pregnant females were noted at 3.8 mg/kg and DBTCl exerts maternal toxicity at 2.5 mg/kg and higher when administered during the entire period of organogenesis in monkeys.

Embryonic/fetal loss was observed in one dam in the control group and eight dams in the 2.5 mg/kg group and four dams in the 3.8 mg/kg group. The increased incidence of pregnant females with embryonic/fetal loss was observed at 2.5 and 3.8 mg/kg, and a significantly increased incidence of these females was found

at 2.5 mg/kg. Embryonic/fetal loss occurred on days 35–64 of pregnancy at 2.5 mg/kg, and on days 38–50 of pregnancy at 3.8 mg/kg. The embryonic mortality during organogenesis in cynomolgus monkeys of 2.4–18.2% has been reported [45]. Binkerd et al. [46] also noted that post-implantation embryonic loss was 5.4% in vehicle control pregnancies in developmental toxicity studies. Average abortion rate in cynomolgus monkeys was 26.1% in control data from 24 teratogenicity studies, and most of the abortions (66.7%) occurred during organogenesis [47]. In the background control data from 1994 to 2004 of the laboratory that performed this study, the post-implantation embryonic loss was 8.8% (29 of the 330 pregnancies). Because the incidence of embryonic/fetal loss in the DBTCl-treated groups was greater than in the historical control values, it was considered to be due to the administration of DBTCl. The data indicate that DBTCl at 2.5 mg/kg was sufficient to induce embryonic/fetal loss and the latter half of organogenesis was more susceptible for DBTCl-induced embryonic loss in cynomolgus monkeys.

We previously reported that DBTCl during early pregnancy caused pre- and post-implantation embryonic loss in pregnant rats [19,20] and that DBTCl suppressed uterine decidualization and reduced the levels of serum progesterone in pseudopregnant rats at doses that induced implantation failure [48]. We also showed that the suppression of uterine decidualization was reversed by administration of progesterone in pseudopregnant rats [48], and that progesterone protected against DBTCl-induced implantation failure [21]. Based on these findings, we hypothesized that the decline in serum progesterone levels was a primary factor for the implantation failure due to DBTCl in rats. However, no significant changes in plasma progesterone levels were noted in monkeys after the administration of DBTCl during organogenesis. The peripheral serum progesterone levels during the first 8 days of pseudopregnancy were essentially similar to those found in pregnant rats, and the serum progesterone levels rose steadily to a peak on day 4 and remained at a plateau of approximately 70 ng/ml until day 8 of pseudopregnancy [49]. In cynomolgus monkeys, plasma progesterone levels had distinct two peaks, one about 15 days postbreeding and another at about days 23–25, the progesterone decline which followed the second peak reached minimal levels (1–2 ng/ml) by about day 45 of pregnancy, and progesterone levels increased gradually throughout the rest of pregnancy with average levels of approximately 4 ng/ml [50]. In our previous study [48], rat blood samples were obtained on day 4 or 9 of pseudopregnancy. At these stages, progesterone levels could be steadily rising or remained at a plateau in pseudopregnant rats. In the present study, blood samples were collected from pregnant monkeys that were carrying their offspring and had not suffered from miscarriage on day 51 of pregnancy. At this stage, progesterone levels could be remained at a nadir in pregnant cynomolgus monkeys. The discrepancy in the effect of DBTCl on serum progesterone levels between rats and monkeys may be explained by the differences in the status and stage of pregnancy. Further studies are required to characterize more precisely the relationship between embryonic loss and maternal progesterone levels in monkeys given DBTCl.

Decreases in the absolute weights of the brain and lung, and an increase in the relative weight of the spleen, which were observed in male fetuses at 3.8 mg/kg, were not thought to be due to the toxic effects of DBTCl on fetal development, because these changes were not found in female fetuses and differences were not detected in the relative weight of the brain and lung or the absolute weight of the spleen in male fetuses. Any adverse effects on the parameters of fetal growth were also not detected in the surviving fetuses of dams given DBTCl. These findings indicate that DBTCl is not toxic to fetal growth at up to 3.8 mg/kg when administered over the entire period of organogenesis. Placental examinations revealed single placentas in all groups. In the background control data of the laboratory that performed the present study, the incidence of single placenta over a period of 10 years was 0–66.7% (mean = 13.0%, 26 of the 213 pregnancies). Histopathological examinations of single placenta revealed no changes, and the histological structure of single placenta was similar to that of normal placenta. These findings indicate that the single placenta observed in the present study was of no toxicological significance.

In the morphological examinations of the fetuses of exposed dams, a few fetuses with morphological changes were found in the DBTCl-treated groups. An accessory spleen was observed in one fetus at 3.8 mg/kg, and a short supernumerary rib was found in one fetus at both 2.5 and 3.8 mg/kg. In the background control data of the laboratory that performed the present study, the accessory spleen over the last 10 years was not observed. Leemans et al. [51] noted that the exact frequency of accessory spleen is not known, but is estimated to be between 10 and 30% in humans, and the immunohistological structure of the accessory spleen was similar to that of the normal spleens. In the present study, histopathological examinations of the accessory spleen revealed no changes, and the histological structure of accessory spleen was similar to that of the normal spleen. The accessory spleen observed in the present study contained only a minute amount of accessory tissue, and it was not considered to be a malformation. Short supernumerary rib is classified as skeletal variation [52], and the incidence of this change in the historical control data of the laboratory that performed the present study was 13.3% (31 of the 240 fetuses). DBTCl caused no skeletal retardation, as evidenced by no significant changes in the number of ossified centers of the vertebral column or the length of the humerus, radius, ulna, femur, tibia or fibula. Chahoud et al. [53] noted that variations are unlikely to adversely affect survival or health, and might result from a delay in growth or morphogenesis; the fetuses otherwise following a normal pattern of development. Furthermore, morphological examinations of aborted or dead embryos/fetuses in the DBTCl-treated groups revealed no anomalies. Considered collectively, these findings suggest that the morphological changes observed in the fetuses in the present study do not indicate a teratogenic response, and that DBTCl possesses no teratogenic potential in cynomolgus monkeys.

In conclusion, the administration of DBTCl to pregnant cynomolgus monkeys throughout organogenesis had an adverse effect on embryonic/fetal survival, but had no adverse effects on fetal morphological development, even at a maternal toxic

dose level. The data from the present study indicate that DBTCl shows embryonic/fetal lethality in monkeys.

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References

- Quevauviller P, Bruchet A, Donard OFX. Leaching of organotin compounds from poly (vinyl chloride) (PVC) materials. *Appl Organomet Chem* 1991;5:125–9.
- Maguire RJ. Aquatic environmental aspects of non-pesticidal organotin compounds. *Water Poll Res J Canada* 1991;26:243–360.
- Piver WT. Organotin compounds: industrial applications and biological investigation. *Environ Health Perspect* 1973;4:61–79.
- WHO. Environmental health criteria 15. Tin and organotin compounds: a preliminary review. Geneva: World Health Organization; 1980.
- Sasaki K, Ishizaka T, Suzuki T, Saito Y. Determination of tri-*n*-butyltin and di-*n*-butyltin compounds in fish by gas chromatography with flame photometric detection. *J Assoc Off Anal Chem* 1988;71:360–6.
- Lau MM. Tributyltin antifoulings: a threat to the Hong Kong marine environment. *Arch Environ Contam Toxicol* 1991;20:299–304.
- Suzuki T, Matsuda R, Saito Y. Molecular species of tri-*n*-butyltin compounds in marine products. *J Agric Food Chem* 1992;40:1437–43.
- Seligman PF, Valkirs AO, Stang PM, Lee RF. Evidence for rapid degradation of tributyltin in a marina. *Marine Pollut Bull* 1988;19:531–4.
- Stewart C, de Mora SJ. A review of the degradation of tri (n-butyl) tin in the marine environment. *Environ Technol* 1990;11:565–70.
- Harazono A, Ema M, Ogawa Y. Pre-implantation embryonic loss induced by tributyltin chloride in rats. *Toxicol Lett* 1996;89:185–90.
- Harazono A, Ema M, Ogawa Y. Evaluation of early embryonic loss induced by tributyltin chloride in rats: phase- and dose-dependent antifertility effects. *Arch Environ Contam Toxicol* 1998;34:94–9.
- Harazono A, Ema M, Kawashima K. Evaluation of malnutrition as a cause of tributyltin-induced pregnancy failure in rats. *Bull Environ Contam Toxicol* 1998;61:224–30.
- Ema M, Kurosaka R, Amano H, Ogawa Y. Further evaluation of the developmental toxicity of tributyltin chloride in rats. *Toxicology* 1995;96:195–201.
- Ema M, Harazono A, Miyawaki E, Ogawa Y. Effect of the day of administration on the developmental toxicity of tributyltin chloride in rats. *Arch Environ Contam Toxicol* 1997;33:90–6.
- Fish RH, Kimmel EC, Casida JE. Bioorganotin chemistry: reactions of tributyltin derivatives with a cytochrome P-450 dependent monooxygenase enzyme system. *J Organomet Chem* 1976;118:41–54.
- Kimmel EC, Fish RH, Casida JE. Bioorganotin chemistry. Metabolism of organotin compounds in microsomal monooxygenase system and in mammals. *J Agric Food Chem* 1977;25:1–9.
- Iwai H, Wada O, Arakawa Y. Determination of tri-, di-, and monobutyltin and inorganic tin in biological materials and some aspects of their metabolism in rats. *J Anal Toxicol* 1981;5:300–6.
- Ema M, Hirose A. Reproductive and developmental toxicity of organotin compounds. In: Golub MS, editor. *Metals, fertility, and reproductive toxicity*. New York: CRC Press (Taylor & Francis Group); 2006. p. 23–64.
- Ema M, Harazono A. Adverse effects of dibutyltin dichloride on initiation and maintenance of rat pregnancy. *Reprod Toxicol* 2000;14:451–6.
- Ema M, Harazono A. Developmental and reproductive toxicity of tributyltin and its metabolite, dibutyltin, in rats. *Congenit Anom (Kyoto)* 2000;40:108–20.
- Ema M, Harazono A, Hirose A, Kamata E. Protective effects of progesterone on implantation failure induced by dibutyltin dichloride in rats. *Toxicol Lett* 2003;143:233–8.
- Ema M, Itami T, Kawasaki H. Teratogenicity of di-*n*-butyltin dichloride in rats. *Toxicol Lett* 1991;58:347–56.
- Ema M, Itami T, Kawasaki H. Susceptible period for the teratogenicity of di-*n*-butyltin dichloride in rats. *Toxicology* 1992;73:81–92.
- Noda T, Yamano T, Shimizu M, Saitoh M, Nakamura T, Yamada A, et al. Comparative teratogenicity of di-*n*-butyltin diacetate with *n*-butyltin trichloride in rats. *Arch Environ Contam Toxicol* 1992;23:216–22.
- Noda T, Nakamura T, Simizu M, Yamano T, Morita S. Critical gestational day of teratogenesis by di-*n*-butyltin diacetate in rats. *Bull Environ Contam Toxicol* 1992;49:715–22.
- Noda T, Morita S, Baba A. Teratogenic effects of various di-*n*-butyltins with different anions and butyl(3-hydroxybutyl)tin dilaurate in rats. *Toxicology* 1993;85:149–60.
- Noda T, Morita S, Baba A. Enhanced teratogenic activity of di-*n*-butyltin diacetate by carbon tetrachloride pretreatment in rats. *Food Chem Toxicol* 1994;32:321–7.
- Noda T, Yamano T, Shimizu M. Effects of maternal age on teratogenicity of di-*n*-butyltin diacetate in rats. *Toxicology* 2001;167:181–9.
- Ema M, Kurosaka R, Amano H, Ogawa Y. Comparative developmental toxicity of butyltin trichloride, dibutyltin dichloride and tributyltin chloride in rats. *J Appl Toxicol* 1995;15:297–302.
- Ema M, Kurosaka R, Amano H, Ogawa Y. Comparative developmental toxicity of di-, tri-, and tetrabutyltin compounds after administration during late organogenesis in rats. *J Appl Toxicol* 1996;16:71–6.
- Ema M, Iwase T, Iwase Y, Ogawa Y. Dysmorphogenic effects of di-*n*-butyltin dichloride in cultured rat embryos. *Toxicol In Vitro* 1995;9:703–9.
- Ema M, Iwase T, Iwase Y, Ohya N, Ogawa Y. Change of embryotoxic susceptibility to di-*n*-butyltin dichloride in cultured rat embryos. *Arch Toxicol* 1996;70:724–8.
- Guideline for Animal Experimentation Issued by Japanese Association for Laboratory Animal Science (1987).
- Law Concerning the Protection and Control of Animals (LAW No. 105, October 1, 1973).
- Standards Relating to the Care and Management of Experimental Animals (Notification No. 6, March 27, 1980 of the Prime Minister's Office).
- Hendrickx AG, Cukierski MA. Reproductive and developmental toxicology in nonhuman primates. In: Graham CE, editor. *Preclinical Safety of biotechnology products intended for human use. Proceeding of a Satellite Symposium to the IV International Congress of Toxicology*. 1986. p. 78–88.
- Dawson AB. A note on the staining of the skeleton of cleared specimens with alizarin red S. *Stain Technol* 1926;1:123–5.
- Staples RE, Haseman JK. Commentary: selection of appropriate experimental units in teratology. *Teratology* 1974;9:259–60.
- Snedecor GW, Cochran WG. *Statistical Methods*. 7th ed. Ames: Iowa State University Press; 1980.
- Dunnett CW. A multiple comparison procedure for comparing several treatments with control. *J Am Statist Assoc* 1996;50:1096–121.
- Miller Jr RG. *Simultaneous Statistical Inference*. 2nd ed. Berlin: Springer-Verlag; 1981.
- Hendrickx AG, Binkerd PE. Primate teratology: selection of species and future use. In: *advances in the study of birth defects, teratological testing*, Vol. 2. Baltimore: University Park Press; 1979. pp. 1–23.
- Schardein JL. *Hormones and hormonal antagonists*. In: *chemically induced birth defects, revised and expanded*. 3rd ed. New York: Marcel Dekker Inc.; 2000. pp. 281–357.
- Poggel HA, Günzel P. Necessity of using nonhuman primates in assessing prenatal toxicity. View of a scientist from the industry. In: Neubert D, Merker H-J, Hendrickx AG, editors. *Non-human primates- developmental biology and toxicology*. Wien: Ueberreuter Wissenschaft; 1988. p. 585–97.
- Hendrickx AG, Binkerd PE. Fetal deaths in nonhuman primates. In: Porter IH, Hook EB, editors. *Embryonic and fetal death*. New York: Academic Press; 1980. p. 45–69.
- Binkerd PE, Tarantal AF, Hendrickx GH. Embryonic/fetal loss and spontaneous malformations in nonhuman primates. In: Neubert D, Merker H-J, Hendrickx AG, editors. *Non-human primates- developmental biology and toxicology*. Wien: Ueberreuter Wissenschaft; 1988. p. 115–28.

- [47] Korte R, Vogel F, Osterburg I, Bell DA. Prenatal waste and spontaneous malformations in Macaques. In: Neubert D, Merker H-J, Hendrickx AG, editors. *Non-human primates- developmental biology and toxicology*. Wien: Ueberreuter Wissenschaft; 1988. p. 141–50.
- [48] Harazono A, Ema M. Suppression of decidual cell response induced by dibutyltin dichloride in pseudopregnant rats: as a cause of early embryonic loss. *Reprod Toxicol* 2003;17:393–9.
- [49] Pepe GL, Rothchild I. A comparative study of serum progesterone levels in pregnancy and in various types of pseudopregnancy in the rat. *Endocrinology* 1974;95:275–9.
- [50] Stabenfeldt GH, Hendrickx AG. Progesterone studies in the *Macaca fascicularis*. *Endocrinology* 1973;92:1296–300.
- [51] Leemans R, Harms G, Timens W. The utility of the accessory spleen: a spare part after accidental splenectomy. In: *the human spleen after trauma: saving techniques and autotransplantation*. Leeuwarden: Grafisch Bedrijf Hellinga bv; 1999. pp. 103–114.
- [52] Solecki R, Bürger H, Buschmann J, Clark R, Duvergere M, Fialkowski O, et al. Harmonisation of rat fetal skeletal terminology and classification. Report of the third workshop on the terminology in developmental toxicology Berlin, 14–16 September 2000. *Reprod Toxicol* 2001;15:713–21.
- [53] Chahoud I, Buschmann J, Clark R, Druga A, Falke H, Faqi A, et al. Classification terms in developmental toxicology: need for harmonization. *Reprod Toxicol* 1999;13:77–82.

Early Pregnancy Failure Induced by Dibutyltin Dichloride in Mice

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ABSTRACT: In this study, we examined the adverse effects of dibutyltin on initiation and maintenance of pregnancy after maternal administration during early pregnancy in mice. Following successful mating, female ICR mice were given dibutyltin dichloride (DBTCl) at 0, 7.6, 15.2, or 30.4 mg/kg bw/day by gastric intubation on days 0–3 or days 4–7 of pregnancy. Female mice were sacrificed on day 18 of pregnancy, and the pregnancy outcome was determined. After administration of DBTCl on days 0–3, the rate of non-pregnant females and the incidence of preimplantation embryonic loss were significantly increased at 30.4 mg/kg bw/day. The incidences of postimplantation embryonic loss in females given DBTCl on days 0–3 at 15.2 mg/kg and higher and on days 4–7 at 7.6 mg/kg bw/day and higher were increased. No increase in the incidence of fetuses with external malformations was observed after the administration of DBTCl on days 0–3 or days 4–7. A decline in the serum progesterone levels was detected in mice given DBTCl at 30.4 mg/kg bw/day on days 0–3 or days 4–7 of pregnancy. The data show that DBTCl adversely affects the initiation and maintenance of pregnancy when administered during early pregnancy in mice and suggest that the decline in serum progesterone levels is responsible for pregnancy failure. © 2007 Wiley Periodicals, Inc. *Environ Toxicol* 22: 44–52, 2007.

Keywords: dibutyltin dichloride; organotin; pregnancy failure; early embryonic loss; progesterone

INTRODUCTION

Organotin compounds are chemicals widely used in agriculture and industry. Disubstituted organotin compounds are commercially the most important derivatives, being used as heat and light stabilizers for polyvinyl chloride (PVC) plastics to prevent degradation of the polymer during the melting and forming of the resin into its final products, as catalysts in the production of polyurethane foams, and as vulcanizing agents for silicone rubbers (Piver, 1973; WHO, 1980). Wide-spread use of organotin compounds has caused increasing amounts to be released into environment.

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The most important route of entry of organotin compounds as nonpesticides into the environment is through the leaching of organotin-stabilized PVC by water (Quevauviller et al., 1991), and its use in antifouling agents resulting in the entry of organotin into the aquatic environment (Maguire, 1991). The identification of dibutyltin (DBT) and tributyltin (TBT) in aquatic marine organisms (Sasaki et al., 1988; Lau, 1991) and marine products (Suzuki et al., 1992) has been reported. TBT is degraded spontaneously and biochemically via a debutylation pathway to DBT in the environment (Seligman et al., 1988; Stewart and de Mora, 1990). Food chain bioaccumulation of butyltin in oysters (Waldock and Thain, 1983), mud crabs (Evans and Laughlin, 1984), marine mussels (Laughlin et al., 1986), Chinook salmon (Short and Thrower, 1986), and dolphin, tuna, and shark (Kannan et al., 1996) has been reported. These findings indicate that butyltins accumulate in the

food chain and are bioconcentrated, and that humans can be exposed to butyltins via food.

Organotins possess toxic effects on reproduction and development in experimental animals (Ema and Hirose, 2006). We previously reported that dibutyltin dichloride (DBTCl) by gavage throughout the period of organogenesis resulted in a significant increase in the incidence of fetal malformations in rats (Ema et al., 1991) and that rat embryos were highly susceptible to the teratogenic effects of DBTCl when administered on day 7 and day 8 of pregnancy (Ema et al., 1992). Tetrabutyltin (TeBT) is metabolized to TBT, DBT, and monobutyltin (MBT) derivatives (Fish et al., 1976; Kimmel et al., 1977). The TBT compound is metabolized to DBT and MBT derivatives and DBT is metabolized to MBT derivatives (Iwai et al., 1981). The developmental toxicity studies on butyltins suggest that the teratogenicity of DBT is different from those of TeBT, TBT, and MBT in its mode of action, because the susceptible period for teratogenicity and types of malformations induced by DBT are different from those induced by TeBT, TBT, and MBT (Ema et al., 1995, 1996). Tributyltin chloride (TBTCl) (Harazono et al., 1996, 1998ab) and DBTCl (Ema and Harazono, 2000ab) during early pregnancy produced pregnancy failure in rats. In rats, the predominant adverse effects on reproduction and development of TBTCl and DBTCl on days 0–3 of pregnancy were a decrease in the pregnancy rate and an increase in the incidence of preimplantation embryonic loss, and TBTCl and DBTCl on days 4–7 of pregnancy mainly caused postimplantation embryonic loss (Harazono et al., 1998b; Ema and Harazono, 2000ab). The doses of DBTCl that caused early embryonic loss were lower than those of TBTCl (Ema and Harazono, 2000b). Thus, the possibility exists that DBTCl and/or metabolites participate in the induction of early embryonic loss due to TBTCl.

The reproductive and developmental effects of organotin compounds, including DBT, were extensively investigated in rats (Ema and Hirose, 2006). We are unaware of any studies in which the adverse effects of DBT on initiation and maintenance of pregnancy have been assessed in mice. Studies in mice would be of great value in evaluating the reproductive and developmental toxicity of DBT. The present study was therefore conducted to determine the adverse effects on the initiation and maintenance of pregnancy of maternal exposure to DBTCl during early pregnancy in mice.

MATERIALS AND METHODS

Animal Husbandry and Maintenance

Male and female Crlj:CD1(ICR) mice at 8 weeks of age were purchased from Atsugi Breeding Center, Charles River Japan, (Yokohama, Japan). The mice were acclimat-

ized to the laboratory for 11 days prior to the start of the experiment. Male and female mice found to be in good health were selected for use. Female mice were caged with male mice and checked the following morning for signs of successful mating by examining vaginal plugs. The day when vaginal plugs were detected was considered to be day 0 of pregnancy. Successfully mated females were distributed into eight groups of 12 mice each and housed individually. Animals were reared on a γ -irradiated basal diet (CRF-1; Oriental Yeast, Tokyo, Japan) and filtered tap water *ad libitum*, and maintained in an air-conditioned room at $22^{\circ}\text{C} \pm 3^{\circ}\text{C}$, with a relative humidity of $50\% \pm 20\%$, under a controlled 12 h light/dark cycle, and ventilation with 10–15 air changes/hour. This study was performed in 2005 at the Safety Research Institute for Chemical Compounds. (Sapporo, Japan) in compliance with the “Law for the Humane Treatment and Management of Animals” (Ministry of the Environment, Japan, 1973), “Standards Relating to the Care and Management, etc. of Experimental Animals” (Prime Minister’s Office, Japan, 1980) and “Guidance for Animal Care and Use of the Safety Research Institute for Chemical Compounds, Co.”

Chemicals and Dosing

DBTCl was purchased from Tokyo Kasei Kogyo (Tokyo, Japan). The DBTCl used in this study was 99.5% pure, and it was kept in a dark and cool place. DBTCl was dissolved in olive oil (Wako Pure Chemical Industries, Osaka, Japan). The female mice were dosed once daily by gastric intubation with DBTCl at a dose of 7.6, 15.2, or 30.4 mg/kg bw (25, 50 or 100 $\mu\text{mol/kg}$ bw) on days 0–3 of pregnancy or on days 4–7 of pregnancy. The dosage levels were determined based on the results of our previous studies, in which increases in the incidence of pre- and postimplantation embryonic loss were caused in female rats gavaged with DBTCl at 7.6 mg/kg bw/day and higher on days 0–3 and days 4–7 of pregnancy, respectively (Ema and Harazono, 2000ab) and our dose-finding study in which no adverse effects on embryonic survival at 15.2 mg/kg bw/day and lower, increased embryonic loss at 30.4 mg/kg bw/day, and one death and three pregnancy failure in four females at 60.8 mg/kg bw/day were found in mice gavaged with DBTCl on days 0–3 of pregnancy. The volume of each dose was adjusted to 5 mL/kg of body weight based on the daily body weight. The control mice received olive oil only on days 0–3 or days 4–7 of pregnancy. All DBTCl solutions were prepared fresh daily.

Observations

All mice were observed for clinical signs of toxicity twice a day during the administration period and daily during the nonadministration period. Females showing a moribund condition were euthanized under ether anesthesia. Maternal

TABLE I. Maternal findings in mice given DBTCI by gastric intubation on days 0–3 of pregnancy

DBTCI (mg/kg)	0 (control)	7.6	15.2	30.4
No. of females successfully mated	12	12	12	12
No. of females showing clinical signs				
Dead	0	1	0	0
Moribund condition (euthanized)	0	1	1	1
Vaginal discharge	0	1	0	0
Jaundice	0	2	7*	10*
Decreased locomotor activity	0	2	1	1
Hypothermia	0	1	1	1
Soil of perigenital fur	0	0	1	0
Initial body weight (g) ^a	27.4 ± 2.0	27.2 ± 2.1	27.2 ± 2.4	27.2 ± 2.1
Body weight gain (g) ^a				
Days 0–4	1.7 ± 1.1	0.6 ± 1.2	1.2 ± 1.6	0.3 ± 0.9*
Days 4–8	2.9 ± 1.5	2.5 ± 2.6	2.1 ± 2.0	1.6 ± 1.5
Days 8–18	20.1 ± 9.1	21.3 ± 12.4	13.6 ± 12.2	8.6 ± 12.2
Adjusted weight gain ^b	8.9 ± 3.4	9.9 ± 3.8	7.9 ± 4.8	5.3 ± 5.0
Food consumption (g) ^a				
Days 0–4	18.2 ± 1.8	15.0 ± 1.9*	16.7 ± 3.2	14.8 ± 2.3*
Days 4–8	22.9 ± 4.9	22.0 ± 2.7	21.7 ± 3.5	20.9 ± 3.5
Days 8–18	71.7 ± 10.1	71.0 ± 12.5	64.6 ± 13.3	57.8 ± 13.4*

^aValues are given as mean ± SD.

^bAdjusted weight gain refers to body weight gain excluding the uterus.

*Significantly different from the control, $P < 0.05$.

body weight was recorded daily, and food consumption was recorded on days 0, 4, 8, 12, and 18 of pregnancy. The females were euthanized by exsanguination under ether anesthesia on day 18 of pregnancy. The uterus was weighed and the number of corpora lutea was recorded. The numbers of implantations, live and dead fetuses, and of resorptions were counted. The uteri were placed in 10% ammonium sulfide for confirmation of the dam's pregnancy status (Salewski, 1964). The live fetuses removed from the uterus were sexed, weighed, and inspected for external malformations and malformations within the oral cavity. The placental weight was also measured.

Analysis of Serum Steroids Hormone Levels

Blood samples were collected from the abdominal aorta under ether anesthesia on day 4 or day 8 of pregnancy, 24 h after the last administration of DBTCI at 0 or 30.4 mg/kg bw/day on days 0–3 or days 4–7 of pregnancy. The serum was separated and stored at -80°C for later assay of steroid hormones. Serum progesterone and 17β -estradiol were measured by Teizo Medical (Kawasaki, Japan) using the liquid chromatography-electrospray ionization Tandem Mass Spectrometry (LC-MS/MS, Applied Biosystems/MDS SCIEX). The detection limits of serum progesterone and 17β -estradiol were 10.0 and 0.25 pg/mL, respectively. The intra- and interassay coefficients of variation for 17β -estradiol were below 6.4% and 8.9%, respectively. The intra- and interassay

coefficients of variation for progesterone were below 9.0% and 7.9%, respectively.

Statistical Analysis

The statistical analysis of fetuses was carried out using the litter as the experimental unit. Maternal body weight, body weight gain, adjusted weight gain, food consumption, numbers of corpora lutea, implantations, embryonic/fetal loss and live fetuses, fetal weight, and placental weight were analyzed for statistical significance as follows. Bartlett's test of homogeneity of variance was used to determine if the groups had equivalent variances at the 5% level of significance. If the variances were equivalent, the groups were compared by one-way analysis of variance. If significant differences were found, Dunnett's multiple comparison test was performed. If the groups were not equivalent, the Kruskal-Wallis test was used to assess the overall effects. Whenever significant differences were noted, pair-wise comparisons were made using the Mann-Whitney U test. The incidences of pre- and postimplantation embryonic loss and fetuses with external malformations were analyzed using Wilcoxon's rank sum test. The incidence of clinical signs in dams, pregnancy, nonpregnancy, and litters with fetal malformations, and the sex ratio of live fetuses were analyzed using Fisher's exact test. The levels of serum progesterone and 17β -estradiol were analyzed by Student's t -test. The 0.05 level of probability was used as the criterion for significance.

TABLE II. Reproductive and developmental findings in mice given DBTCl by gastric intubation on days 0-3 of pregnancy

DBTCl (mg/kg)	0 (control)	7.6	15.2	30.4
No. of females successfully mated	12	12	12	12
No. of nonpregnant females	1	3	4	7*
No. of pregnant females	11	9	8	5*
No. of implantations per female ^{a,b}	9.5 ± 5.1	9.8 ± 7.1	8.3 ± 7.0	5.4 ± 6.7
Pre-implantation loss per female (%) ^{a,b}	9.7	29.7 ^c	34.0	58.3*
No. of pregnant females surviving until scheduled sacrifice	11	8	7	4
No. of litters totally resorbed	0	0	1	1
No. of corpora lutea per litter ^{a,d}	10.5 ± 4.3	13.1 ± 4.9	12.4 ± 4.4	13.3 ± 1.3
No. of implantations per litter ^{a,d}	10.4 ± 4.3	12.6 ± 4.9	12.3 ± 4.4	13.3 ± 1.3
Pre-implantation loss per litter (%) ^{d,e}	1.5	3.3	1.1	0
No. of post-implantation loss per litter ^{a,d}	1.0 ± 1.0	1.1 ± 1.5	4.1 ± 3.2	4.0 ± 5.4
Post-implantation loss per litter (%) ^{d,f}	10.1	14.1	41.3*	32.2
No. of live fetuses per litter ^{a,d}	9.4 ± 4.2	11.5 ± 5.3	8.1 ± 5.0	9.3 ± 6.2
Sex ratio of live fetuses (male / female)	50/53	47/45	30/27	21/16
Body weight of live fetuses (g) ^a				
Male	1.54 ± 0.19	1.30 ± 0.12*	1.14 ± 0.22*	1.12 ± 0.10*
Female	1.42 ± 0.15	1.28 ± 0.20	1.08 ± 0.26*	1.01 ± 0.11*
External examinations of fetuses				
No. of fetuses (litters) examined	103 (11)	92 (8)	57 (6)	37 (3)
No. of fetuses (litters) with anomalies	1 (1)	0	1 (1)	0
Cleft palate	1	0	1	0
Kinked tail	0	0	1	0
Placental weight (mg) ^a	125 ± 56	116 ± 15	120 ± 17	119 ± 16

^a Values are given as mean ± SD.

^b Values obtained from females successfully mated.

^c Value obtained from 11 females, because corpora lutea were indistinguishable in one female.

^d Values obtained from pregnant females surviving until scheduled sacrifice.

^e [(No. of corpora lutea—no. of implantations)/no. of corpora lutea] × 100.

^f [(No. of resorptions and dead fetuses/no. of implantations) × 100.

* Significantly different from the control, $P < 0.05$.

RESULTS

Administration of DBTCl on Days 0-3 of Pregnancy

Table I shows the maternal findings in mice given DBTCl on days 0-3 of pregnancy. One death was observed at 7.6 mg/kg bw/day, and one female each showed a moribund condition at 7.6, 15.2, and 30.4 mg/kg bw/day, and was euthanized. The female mice in the DBTCl-treated groups showed vagina discharge, jaundice, decreased locomotor activity, hypothermia and/or soiled perigenital fur, and the incidence of females showing jaundice was significantly increased at 15.2 mg/kg bw/day and higher. A significantly decreased body weight gain on days 0-4 was noted at 30.4 mg/kg bw/day. Food consumption on days 0-4, days 4-8, and days 8-18 in the DBTCl-treated groups were reduced, and significantly decreased food consumptions on days 0-4 at 7.6 and 30.4 mg/kg bw/day and on days 8-18 at 30.4 mg/kg bw/day were observed.

The reproductive and developmental findings in mice given DBTCl on days 0-3 of pregnancy are shown in

Table II. The total absence of any implantation site, i.e., nonpregnancy, was found in one, three, four, and seven of the 12 females in the control, 7.6, 15.2, and 30.4 mg/kg bw/day groups, respectively. In the successfully mated females, the pregnancy rate was significantly decreased, and the incidence of preimplantation embryonic loss per females was significantly increased at 30.4 mg/kg bw/day. In the pregnant females that survived until the scheduled sacrifice, the number of corpora lutea per litter, implantations per litter, live fetuses per litter, the incidence of litters totally resorbed and of preimplantation loss per litter, and the sex ratio of live fetuses were not significantly different between the control and DBTCl-treated groups. The incidence of postimplantation loss per litter was increased in the DBTCl-treated groups, and a significant increase was observed at 15.2 mg/kg bw/day. A significantly lower fetal weight was found in males at 7.6 mg/kg bw/day and in both sexes at 15.2 and 30.4 mg/kg bw/day. One fetus with cleft palate in the control group and one fetus with a cleft palate and kinked tail in the 15.2 mg/kg bw/day group were observed. The placental weight in the DBTCl-treated

TABLE III. Maternal findings in mice given DBTCl by gastric intubation on days 4–7 of pregnancy

DBTCl (mg/kg)	0 (control)	7.6	15.2	30.4
No. of females successfully mated	12	12	12	12
No. of females showing clinical signs				
Dead	0	0	1	0
Moribund condition (euthanized)	0	0	0	1
Vaginal discharge	0	0	4	4
Jaundice	0	0	2	6*
Decreased locomotor activity	0	0	0	1
Hypothermia	0	0	0	1
Initial body weight (g) ^a	28.1 ± 1.8	28.1 ± 1.8	28.1 ± 1.8	28.2 ± 1.7
Body weight gain (g) ^a				
Days 0–4	1.6 ± 1.0	1.9 ± 0.8	1.2 ± 1.2	1.6 ± 0.9
Days 4–8	3.1 ± 1.1	1.9 ± 1.6	0.5 ± 1.8*	−0.3 ± 2.1*
Days 8–18	24.9 ± 9.1	14.9 ± 8.9*	2.9 ± 6.3*	2.4 ± 2.4*
Adjusted weight gain ^b	8.3 ± 3.5	8.1 ± 4.3	3.2 ± 5.3*	3.8 ± 3.2*
Food consumption (g) ^a				
Days 0–4	18.5 ± 1.9	18.9 ± 2.4	18.4 ± 2.7	18.8 ± 1.3
Days 4–8	21.8 ± 1.9	19.2 ± 2.6	16.4 ± 3.3*	15.6 ± 3.5*
Days 8–18	74.5 ± 12.1	67.7 ± 9.9	55.2 ± 12.6*	57.2 ± 6.2*

^aValues are given as mean ± SD.

^bAdjusted weight gain refers to body weight gain excluding the uterus.

*Significantly different from the control, $P < 0.05$.

groups was not significantly different from that in the control group.

Administration of DBTCl on Days 4–7 of Pregnancy

Table III shows the maternal findings in mice given DBTCl on days 4–7 of pregnancy. One death was observed at 15.2 mg/kg bw/day, and one female that showed a moribund condition at 30.4 mg/kg bw/day was euthanized. The female mice in the DBTCl-treated groups showed vaginal discharge, jaundice, decreased locomotor activity, and/or hypothermia, and the incidence of females with jaundice was significantly increased at 30.4 mg/kg bw/day. The body weight gain on days 4–8 and adjusted weight gain, which indicates the net weight gain of female mice, at 15.2 mg/kg bw/day and higher, and on days 8–18 at 7.6 mg/kg bw/day and higher were significantly decreased. Food consumption on days 4–8 and days 8–18 was significantly lowered at 15.2 mg/kg bw/day and higher.

The reproductive and developmental findings in mice given DBTCl on days 4–7 of pregnancy are presented in Table IV. Although nonpregnancy was found in one, two, and one of the 12 females in the control, 7.6, 15.2, and 30.4 mg/kg bw/day groups, respectively, no significant decrease in the pregnancy rate was noted in the DBTCl-treated groups. In the successfully mated females, the number of implantations per female was significantly decreased at 15.2 mg/kg bw/day. In the pregnant females that survived until the scheduled sacrifice, totally resorbed litters were found in 2 of the 11 females at 7.6 mg/kg bw/day, 8 of the 9 females at 15.2 mg/kg bw/day,

and 10 of the 10 females at 30.4 mg/kg bw/day. At 30.4 mg/kg bw/day, no live fetuses were obtained. The numbers of corpora lutea per litter, implantations per litter, and preimplantation loss per litter, and the sex ratio of live fetuses in the DBTCl-treated groups were not significantly different from those in the control group. A significant increase in the number and incidence of postimplantation loss per litter, and a decrease in the number of live fetuses were found in the DBTCl-treated groups. The weights of male and female fetuses were significantly lowered at 7.6 mg/kg bw/day. One fetus with omphalocele, and one fetus with exencephaly and open eyelids were observed at 7.6 mg/kg bw/day. The placental weight was not significantly different between the control and the DBTCl-treated groups.

Serum Progesterone and 17 β -Estradiol Levels

The serum progesterone and 17 β -estradiol levels are shown in Figure 1. A significant reduction in the serum progesterone levels was noted in female mice given DBTCl on days 0–3 or days 4–7 of pregnancy. Although higher levels of serum 17 β -estradiol were observed after the administration of DBTCl on days 4–7 of pregnancy, no statistically significant difference in 17 β -estradiol levels were detected between the control and DBTCl-treated groups.

DISCUSSION

The present study was designed to evaluate the adverse effects of DBTCl on the initiation and maintenance of

TABLE IV. Reproductive and developmental findings in mice given DBTCl by gastric intubation on days 4-7 of pregnancy

DBTCl (mg/kg)	0 (control)	7.6	15.2	30.4
No. of females successfully mated	12	12	12	12
No. of nonpregnant females	1	1	2	1
No. of pregnant females	11	11	10	11
No. of implantations per female ^{a,b}	12.6 ± 4.4	13.2 ± 4.6	7.5 ± 5.7*	11.1 ± 5.4
Pre-implantation loss per female (%) ^{a,b}	8.9	8.9	24.7	18.3 ^c
No. of pregnant females surviving until scheduled sacrifice	11	11	9	10
No. of litters totally resorbed	0	2	8*	10*
No. of corpora lutea per litter ^{a,d}	13.8 ± 2.1	14.5 ± 2.3	10.6 ± 5.2	13.9 ± 2.8
No. of implantations per litter ^{a,d}	13.7 ± 2.1	14.4 ± 2.2	9.4 ± 5.1	12.7 ± 4.1
Pre-implantation loss per litter (%) ^{d,e}	0.6	0.6	10.7	10.2
No. of postimplantation loss per litter ^{a,d}	0.6 ± 1.0	7.2 ± 6.1*	8.7 ± 4.8*	12.7 ± 4.1*
Post-implantation loss per litter (%) ^{d,f}	4.3	48.3*	94.4*	100*
No. of live fetuses per litter ^{a,d}	13.1 ± 2.0	7.2 ± 5.6*	0.8 ± 2.3*	0
Sex ratio of live fetuses (male/female)	82/62	50/29	4/3	
Body weight of live fetuses (g) ^a				
Male	1.45 ± 0.10	1.23 ± 0.10*	1.27	
Female	1.39 ± 0.10	1.18 ± 0.14*	1.18	
External examinations of fetuses				
No. of fetuses (litters) examined	144 (11)	79 (9)	7 (1)	
No. of fetuses (litters) with anomalies	0	2 (2)	0	
Omphalocele	0	1	0	
Exencephaly and open eyelids	0	1	0	
Placental weight (mg) ^a	102 ± 10	99 ± 12	114	

^aValues are given as mean ± SD.^bValues obtained from females successfully mated.^cValue obtained from 11 females, because corpora lutea were indistinguishable in one female.^dValues obtained from pregnant females surviving until scheduled sacrifice.^e[(No. of corpora lutea—no. of implantations)/no. of corpora lutea] × 100.^f(No. of resorptions and dead fetuses/no. of implantations) × 100.*Significantly different from the control, *P* < 0.05.

pregnancy following maternal exposure during early pregnancy in mice. The most striking finding in the present study is pregnancy failure, decrease in the pregnancy rate, and litters totally resorbed, in females given DBTCl during early pregnancy.

Death and/or moribund condition were observed after the administration of DBTCl at 7.6 mg/kg bw/day and higher on days 0-3 of pregnancy and at 15.2 mg/kg bw/day and higher on days 4-7 of pregnancy, and significant increased incidence of females showing clinical signs of toxicity were found after the administration of DBTCl at 15.2 mg/kg bw/day and higher on days 0-3 of pregnancy and at 30.4 mg/kg bw/day on days 4-7 of pregnancy. These findings indicate that more severe general toxicity was induced by DBTCl on days 0-3 of pregnancy than that on days 4-7 of pregnancy. However, adverse effects on body weight gain were detected after the administration of DBTCl at 30.4 mg/kg bw/day on days 0-3 of pregnancy and at 7.6 mg/kg bw/day and higher on days 4-7 of pregnancy. Although the recovery of body weight gain was observed after the administration of DBTCl on days 0-3 of

pregnancy, recovery by the end of the study was not found in females given DBTCl at 7.6 mg/kg bw/day and higher after the administration on days 4-7 of pregnancy. Following the administration on days 4-7 of pregnancy, a significantly lower adjusted weight gain was also noted in females given DBTCl at 15.2 mg/kg/day and higher. These findings indicate that more severe adverse effects on body weight gain were induced by DBTCl on days 4-7 of pregnancy than that on days 0-3 of pregnancy. More severe effects of DBTCl on body weight gain following the administration on days 4-7 may be attributable to the significant decrease in the number of live fetuses.

The earlier administration period, days 0-3 of pregnancy, corresponds to the period before implantation, and the later administration period, days 4-7 of pregnancy, corresponds to the period when implantation is in progress and the period shortly after implantation in mice (Rugh, 1968). We expected that DBTCl insult on days 0-3 of pregnancy might result in preimplantation loss of embryos; i.e., the absence or decrease of implantation sites, and DBTCl insult on days 4-7 of pregnancy might result in postimplantation loss of embryos; i.e.,

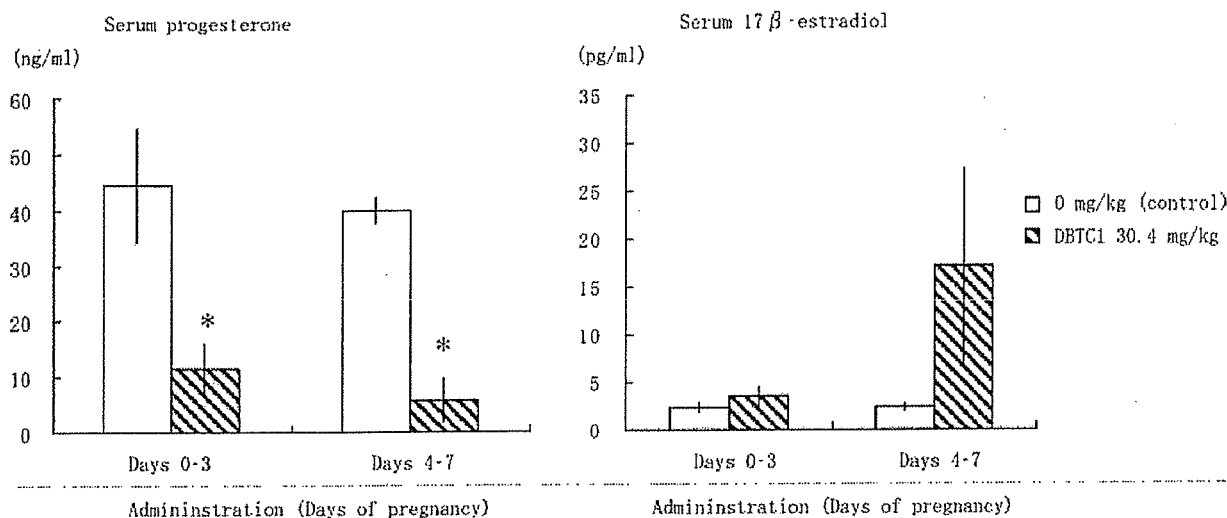


Fig. 1. Serum progesterone and 17β -estradiol levels in female mice given DBTCl on days 0–3 or days 4–7 of pregnancy. Blood samples were collected on day 4 or day 8 of pregnancy, 24 h after the last administration of DBTCl. Values are given as the mean \pm SEM of seven or eight mice. *Significantly different from the control group, $P < 0.05$.

the resorption of implantation sites. The most striking adverse effects of DBTCl on reproduction and development were a decrease in pregnancy rate, complete implantation failure, when DBTCl was given to mice on days 0–3 of pregnancy. The findings of an increased incidence of preimplantation embryonic loss in successfully mated females, and an increased incidence of postimplantation embryonic loss and low fetal weight in pregnant females survived until scheduled sacrifice after the administration of DBTCl on days 0–3 of pregnancy may suggest that DBTCl adversely affects preimplantation embryos and also the later survival and growth of embryos/fetuses when administered during the preimplantation period. On the other hand, the predominant adverse effects of DBTCl on reproduction and development were postimplantation loss, complete litter loss, when DBTCl was given to mice on days 4–7 of pregnancy. The findings of an increase in the incidence of postimplantation embryonic loss and a decrease in the fetal weight after administration of DBTCl on days 4–7 of pregnancy may suggest that DBTCl has effects on the later survival and growth of embryos/fetuses when administered during the peri-implantation period. Considered collectively, these findings indicate that the manifestation of adverse effects of DBTCl on reproduction and development varies with the stages of pregnancy at the time of maternal exposure.

The corpora lutea are essential up to the end of pregnancy in mice (Deansely, 1966). The embryo transport process in mice is triggered by progesterone and requires progesterone activity for its maintenance (Kendle and Lee, 1980). In mice, 24 h of progesterone priming is not only adequate for implantation, but this priming has a long-term effect on implantation

(Huet-Hudson and Dey, 1990). In our previous studies in rats, increases in the incidences of early embryonic loss were observed after the administration of DBTCl during early pregnancy (Ema and Harazono, 2000ab). The suppression of uterine decidualization and reduced levels of serum progesterone were found in female rats given DBTCl on days 0–3 or days 4–7 of pseudopregnancy (Harazono and Ema, 2003), and lowered reproductive parameters in female rats given DBTCl were recovered by the administration of progesterone (Ema et al., 2003). Based on these findings, we hypothesized that the decline in serum progesterone levels in pregnant animals was a primary mechanism for the implantation failure due to DBTCl in rats. In the present study in mice, a decline in serum progesterone levels was detected after the administration of DBTCl during early pregnancy. These findings are in good agreement with previous findings that DBTCl induced early embryonic loss and decreased serum progesterone levels in pregnant rats. There is a similarity in the effects of DBTCl on progesterone levels in early pregnancy in rats and mice, and these suggest that the decline in the serum progesterone levels is also the factor responsible for the DBTCl-induced pregnancy failure in mice. Early pregnancy failure was also caused by systemic activation of the CD-40 immune costimulatory pathway in mice (Erlebacher et al., 2004). They noted that pregnancy failure resulted from impaired progesterone synthesis by the corpus luteum of the ovary, an endocrine defect in turn associated with ovarian resistance to the gonadotropic effects of prolactin and that pregnancy failure also required the proinflammatory cytokine TNF- α and correlated with the luteal induction of the prolactin receptor signaling inhibitors suppressor of cytokine signaling 1

(Socs1) and Socs3. Our results of the present study may support their argument. To further evaluate the adverse effects of DBTCI during early pregnancy, determination of the gene expression profile in the uterus of mice and rats is currently in progress.

In conclusion, DBTCI adversely affects the initiation and maintenance of pregnancy when administered during early pregnancy in mice, and the present data suggest that the decline in progesterone is the responsible factor for the early pregnancy failure in mice.

REFERENCES

- Deansely R. 1966. The endocrinology of pregnancy and foetal life. In: Parks AS, editor. *Marshall's Physiology of Reproduction*, Vol. 3. Boston: Little Brown. pp 891–1063.
- Ema M, Harazono A. 2000a. Adverse effects of dibutyltin dichloride on initiation and maintenance of rat pregnancy. *Reprod Toxicol* 14:451–456.
- Ema M, Harazono A. 2000b. Developmental and reproductive toxicity of tributyltin and its metabolite, dibutyltin, in rats. *Congenit Anom (Kyoto)* 40:108–120.
- Ema M, Hirose A. 2006. Reproductive and developmental toxicity of organotin compounds. In: Golub MS, editor. *Metals, Fertility, and Reproductive Toxicity*. New York: CRC Press (Taylor & Francis Group). pp 23–64.
- Ema M, Itami T, Kawasaki H. 1991. Teratogenicity of di-*n*-butyltin dichloride in rats. *Toxicol Lett* 58:347–356.
- Ema M, Itami T, Kawasaki H. 1992. Susceptible period for the teratogenicity of di-*n*-butyltin dichloride in rats. *Toxicology* 73: 81–92.
- Ema M, Kurosaka R, Amano H, Ogasawa Y. 1995. Comparative developmental toxicity of butyltin trichloride, dibutyltin dichloride and tributyltin chloride in rats. *J Appl Toxicol* 15:297–302.
- Ema M, Kurosaka R, Amano H, Ogawa Y. 1996. Comparative developmental toxicity of di-, tri-, and tetrabutyltin compounds after administration during late organogenesis in rats. *J Appl Toxicol* 16:71–76.
- Ema M, Harazono A, Hirose A, Kamata E. 2003. Protective effects of progesterone on implantation failure induced by dibutyltin dichloride in rats. *Toxicol Lett* 143:233–238.
- Erlbacher A, Zhang D, Parlow AF, Glimcher LH. 2004. Ovarian insufficiency and early pregnancy loss induced by activation of the innate immune system. *J Clin Invest* 114:39–48.
- Evans DW, Laughlin RB Jr. 1984. Accumulation of bis(tributyltin)oxide by the mud crab, *Rhithropanopeus harrisi*. *Chemosphere* 13:213–219.
- Fish RH, Kimmel EC, Casida JE. 1976. Bioorganotin chemistry: Reactions of tributyltin derivatives with a cytochrome P-450 dependent monooxygenase enzyme system. *J Organomet Chem* 118:41–54.
- Harazono A, Ema M. 2003. Suppression of decidual cell response induced by dibutyltin dichloride in pseudopregnant rats: As a cause of early embryonic loss. *Reprod Toxicol* 17:393–399.
- Harazono A, Ema M, Ogawa Y. 1996. Pre-implantation embryonic loss induced by tributyltin chloride in rats. *Toxicol Lett* 89:185–190.
- Harazono A, Ema M, Kawashima K. 1998a. Evaluation of malnutrition as a cause of tributyltin-induced pregnancy failure in rats. *Bull Environ Contam Toxicol* 61:224–230.
- Harazono A, Ema M, Ogawa Y. 1998b. Evaluation of early embryonic loss induced by tributyltin chloride in rats: Phase- and dose dependent antifertility effects. *Arch Environ Contam Toxicol* 34:94–99.
- Huet-Hudson YM, Dey SK. 1990. Requirement for progesterone priming and its long-term effects on implantation in the mouse. *Proc Soc Exp Biol Med* 193:259–263.
- Iwai H, Wada O, Arakawa Y. 1981. Determination of tri-, di-, and monobutyltin and inorganic tin in biological materials and some aspects of their metabolism in rats. *J Anal Toxicol* 5: 300–306.
- Kannan K, Corsolini S, Focardi S, Tanabe S, Tatsukawa R. 1996. Accumulation pattern of butyltin compounds in dolphin, tuna, and shark collected from Italian coastal waters. *Arch Environ Contam Toxicol* 31:19–23.
- Kendle KE, Lee B. 1980. Investigation of the influence of progesterone on mouse embryo transport by using antiprogestational steroids. *J Reprod Fertil* 58:253–258.
- Kimmel EC, Fish RH, Casida JE. 1977. Bioorganotin chemistry. Metabolism of organotin compounds in microsomal monooxygenase system and in mammals. *J Agric Food Chem* 25:1–9.
- Lau MM. 1991. Tributyltin antifoulings: A threat to the Hong Kong marine environment. *Arch Environ Contam Toxicol* 20: 299–304.
- Laughlin RB Jr, French W, Guard HE. 1986. Accumulation of bis(tributyltin) oxide by the marine mussel *Mytilus edulis*. *Environ Sci Technol* 20:884–890.
- Maguire RJ. 1991. Aquatic environmental aspects of nonpesticidal organotin compounds. *Water Pollut Res J Can* 26:243–260.
- Ministry of the Environment. 1973. Law for the Humane Treatment and Management of Animals, Law No. 105, October 1, 1973, revised June 15, 2005. Japan: Ministry of the Environment.
- Piver WT. 1973. Organotin compounds: Industrial applications and biological investigation. *Environ Health Perspect* 4:61–79.
- Prime Minister's Office. 1980. Standards Relating to the Care and Management, etc. of Experimental Animals, Notification No. 6, March 27, 1980 of the Prime Minister's Office. Japan: Prime Minister's Office.
- Quevauviller P, Bruchet A, Donard OFX. 1991. Leaching of organotin compounds from poly (vinyl chloride) (PVC) materials. *Appl Organomet Chem* 5:125–129.
- Rugh R. 1968. *The Mouse: Its Reproduction and Development*. Minneapolis: Burgess Publishing. 85p.
- Salewski E. 1964. Färbemethode zum makroskopischen nachweis von implantationsstellen am uterus der ratte. *Naunyn-Schm Arch Exp Pathol Pharmacol* 247:367–372.
- Sasaki K, Ishizaka T, Suzuki T, Saito Y. 1988. Determination of tri-*n*-butyltin and di-*n*-butyltin compounds in fish by gas chromatography with flame photometric detection. *J Assoc Off Anal Chem* 71:360–366.

- Seligman PF, Valkirs AO, Stang PM, Lee RF. 1988. Evidence for rapid degradation of tributyltin in a marina. *Mar Pollut Bull* 19:531-534.
- Short JW, Thrower FP. 1986. Accumulation of butyltins in muscle tissue of Chinook salmon reared in sea pens treated with tri-*n*-butyltin. *Mar Pollut Bull* 17:542-545.
- Stewart C, de Mora SJ. 1990. A review of the degradation of tri (*n*-butyl) tin in the marine environment. *Environ Technol* 11:565-570.
- Suzuki T, Matsuda R, Saito Y. 1992. Molecular species of tri-*n*-butyltin compounds in marine products. *J Agric Food Chem* 40:1437-1443.
- Waldock MJ, Thain JE. 1983. Shell thickening in *Crassostea gigas*: Organotin antifouling or sediment induced? *Mar Pollut Bull* 14:411-415.
- WHO. 1980. Environmental Health Criteria 15 Tin and Organotin Compounds: A Preliminary Review. Geneva: World Health Organization.



Identification of amino acid residues in the Ah receptor involved in ligand binding

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Abstract

The Ah receptor (AhR) is a ligand-activated transcription factor. Five amino acids as candidate amino acids necessary for ligand binding within or near the ligand-binding domain were selected based on their evolutionary conservation and their aromatic nature that could interact with xenobiotic ligands. These amino acids were changed to Ala, and the mutated AhRs were subjected to a test of their transactivation activity in HeLa cells. Mutation of Phe318 completely lost its activity whereas other mutations only weakly impaired activity. The Leu-substituted mutant, AhR(Phe318Leu), activated the luciferase activity to the level comparable to wild type in the cells treated with 3-methylcholanthrene (MC) but not at all with β -naphthoflavone (β -NF). Ligand-binding activity of mutants was examined with [³H]MC *in vitro*. AhR(Phe318Ala) could not bind to [³H]MC. [³H]MC bound by AhR(Phe318Leu) was competed with unlabeled MC but not with β -NF. A structural model of the ligand-binding domain was constructed.

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Administration of xenobiotics such as 2,3,7,8-tetrachlorodibenzo-*p*-dioxin (TCDD), 3-methylcholanthrene (MC),

and β -naphthoflavone (β -NF) into experimental animals induces several drug-metabolizing enzymes such as CYP1A1 in the liver. These inducers act as ligands for the Ah receptor (AhR), and subsequently, the ligand-activated AhR activates transcription of genes encoding the enzymes [1]. Numerous environmental pollutants, agricultural chemicals, and drugs are known to serve as ligands for the AhR. Polyhalogenated aromatic hydrocarbons such as TCDD and coplanar polychlorinated biphenyls, polycyclic aromatic hydrocarbons such as 3-MC, benzo[a]pyrene and formylindolo[3,2-*b*]carbazoles, and flavonoids such as β -NF are representative potent ligands [1,2]. The most noticeable characteristic of the ligands is that they are organic molecules with planar aromatic rings. In resting cells, the AhR is associated with Hsp90 in the cytoplasm

Abbreviations: AhR, aryl hydrocarbon receptor; MC, 3-methylcholanthrene; β -NF, β -naphthoflavone; TCDD, 2,3,7,8-tetrachlorodibenzo-*p*-dioxin; YFP, yellow fluorescent protein.

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as a soluble receptor. Owing to their lipophilic nature, it is presumed that ligands enter into cells by simple diffusion, and bind to the AhR. Ligand-induced conformation change of the AhR is believed to cause exposure of its nuclear localization signal and succeeding nuclear translocation of the liganded AhR. In the nucleus, the AhR forms a heterodimer with the Ah receptor nuclear translocator (Arnt), and then the heterodimer binds to a specific enhancer termed XRE (DRE or AhRE) localized in the upstream region of target genes [1]. The AhR and Arnt belong to the basic HLH–PAS domain protein family. Vertebrate PAS domains were generally composed of two imperfect repeated regions of about 110 amino acids named PAS-A and PAS-B domains. PAS and HLH domains serve as domains for dimerization with partner PAS proteins. In addition to the dimerization function, some PAS domains contain small organic compounds such as heme, probably for its sensing function [3]. The PAS-B domain of the AhR has the function of binding xenobiotic ligands [4]. The AhR homolog is also distributed in invertebrate species. Interestingly, recent studies demonstrate that *Drosophila* AhR (spineless) and *Caenorhabditis elegans* AhR (AhR-1) have no activity to bind foreign or endogenous chemicals as ligands. Although the protein has no ligand-binding activity, these AhRs heterodimerize with Arnt, binding to the DNA of which sequence is the same as XRE, and activating transcription [5,6].

In this study, we identified amino acids that play a key role in ligand binding of the AhR by several site-directed mutagenesis experiments. Furthermore, a three-dimensional model of the ligand-binding domain was constructed, which demonstrated good agreement with the results of the mutagenesis experiments.

Materials and methods

Construction of plasmids. pBOSFlag-mAhR-HA was constructed as follows. Oligonucleotides, 5'-CCACCGCCATGGACTACAAAGACGATGACGATAAAGGCATGGGCTGCA and 5'-GCCATGCCTTTATCGTCATCGTCTTTGTAGTCCATGGGCGGTGGAGCT for Flag peptide were inserted into the *SacI* and *PstI* site of pBluescript II. Full-length mouse AhR cDNA was inserted into the *HindIII* site of the generated plasmid. Using the plasmid as a template, a fragment of Flag-mAhR-HA was generated by PCR using primers, 5'-CCACCGCCATGGACTACAAAGACGATGACGATAAAGGCATGGGCTGCA (forward) and 5'-CTCGAGCTAGGCGTAGGTCGGGCACGTCGAGTTCGACACACTCTGCACCTTGCTTAGGAATGCC (reverse), and the fragment was inserted into the *XbaI* site of the pEFBOS vector. Expression plasmids for mutated AhRs were produced by site-directed mutagenesis using PCR. Construction of XRE₄-tkLuc was described previously [7]. Chimeric plasmids for pFlag-mAhR-YFP were constructed as follows. A DNA fragment containing the Flag-AhR part of pBOSFlag-mAhR-HA was amplified by PCR, digested by *Bam*HI and *Sal*I and inserted into the *Nhe*I and *Xho*I sites of pEYFPN1 (Clontech). The resultant plasmid was digested with *Eco*RI and *Bam*HI, treated with Klenow fragment and self-ligated to make the sequence in-frame.

DNA transfection and Western blotting. HeLa cells were grown in MEM supplemented with 10% fetal bovine serum. DNA transfection into HeLa cells (grown in a 60 mm dish) was carried out by the calcium phosphate method using 2 µg reporter plasmid XRE₄-tkLuc, 1 µg pBOSFlag-mAhR-HA, 1 µg pBOSmArnt, and 1 µg pBOSLacZ for internal control as

described [7]. Western blotting was performed using whole cell extracts from COS-7 cells transfected with pBOSFlag-mAhR-HA or its AhR mutants and a monoclonal anti-HA antibody (Roche, 12CA5). Because of low expression levels of the overexpressed proteins in HeLa cells, HEK293T cells were used to compare the expression levels of various mutants of the AhR, and it was found that they were relatively evenly expressed (data not shown).

Fluorescence observation of cells. CHO-K1 cells were provided by the Cell Resource Center for Biomedical Research, Institute of Development, Aging and Cancer, Tohoku University. Cells grown on the cover glass were transfected with 0.25 µg AhR–YFP fusion plasmids using FuGENE6 transfection reagent (Roche). After incubation for 40 h, cells were treated with MC or β-NF at a given concentration for 2 or 4 h, respectively. Imaging was performed with an Olympus BX50 fluorescence microscope equipped with a filter set (Olympus U-MYFPHQ) and an Olympus DP70 digital camera.

In vitro binding assay. Cytosolic extracts (1 mg protein/ml) from COS-7 cells transfected with expression plasmids for AhRs were prepared as described [8] and [³H]-labeled MC (1 µCi, 1.2 Ci/mmol, Moravak Biochemicals) was added to 450 µl of the extracts. The mixture was incubated at 4 °C for 2 h with or without unlabeled competitors, treated with dextran-coated charcoal and subjected to fractionation by 10–30% (v/v) glycerol gradient centrifugation at 50,000 rpm at 1 °C for 14 h.

Modeling the structure of PAS-B domain. The multiple alignment in the homology modeling procedure was performed based on the predicted and the observed secondary structures of the reference proteins, FixL [9], HERG [10], PHY3 [11,12], EC DOS [13], HIF-2α [14], and PAS kinase [15], while taking into consideration the sequence and structure conservation in their families. A homology model of the mAhR PAS-B domain was generated by means of the modeling module in Insight 2000 (Accelrys Inc.). The docking process was performed using the docking module of the Cerius² system (Accelrys Inc.).

Results

Transactivation activity of mutated AhR

Candidate amino acids for ligand recognition and binding were selected on the basis of the following two assumptions. (1) Amino acids are conserved among vertebrate species whose AhRs exhibit ligand-binding activity, but are not conserved in the *Drosophila* and *C. elegans* AhRs that are deficient in binding activity. (2) Interactions between ligands and amino acids include the stacking force between aromatic side chains and aromatic rings of ligands because all ligands have hydrophobic aromatic rings. There were a number of amino acids that satisfied the first criterion. Accordingly, the second criterion was placed on the amino acids. Selection of amino acids satisfying the two criteria revealed five aromatic amino acids within and near the PAS-B domain as shown in Fig. 1A. The amino acids were mutated to Ala, and the transactivation activity of the corresponding mutated AhR was assayed. As shown in Fig. 1B, activity decreased to the basal level in the presence of MC by mutation of Phe318 to Ala. This loss of activity was also seen with other inducers including TCDD and β-NF. Other mutations caused a slight decrease in the transactivation activity. The Phe318 was changed to other amino acids as shown in Fig. 1C, and the transactivation activity of the mutated AhRs was assayed. Substitution to aromatic amino acids, Tyr or Trp, showed an inducible luciferase activity by the stimulus of MC and β-NF,