

表2 皮膚腐食性/刺激性カテゴリー分類基準 (カテゴリー1: 腐食性物質)

Corrosive category (category 1) (applies to authorities not using subcategories)	Potential corrosive subclasses (only applies to some authorities)	Corrosive in ≥ 1 of 3 animals	
		Exposure	observation
corrosive	corrosive subcategory 1 A	≤ 3 minutes	≤ 1 hour
	corrosive subcategory 1 B	> 3 minutes-- ≤ 1 hour	≤ 14 days
	corrosive subcategory 1 C	> 1 hour-- ≤ 4 hours	≤ 14 days

表3 混合物の皮膚腐食性・刺激性カテゴリー分類基準

Sum of ingredients classified as:	Concentration triggering classification of a mixture as:		
	Skin		
	Corrosive	Irritant	
	category 1 (see note below)	category 2	category 3
Skin Category 1	$\geq 5\%$	$\geq 1\%$ but $< 5\%$	
Skin Category 2		$\geq 10\%$	$\geq 1\%$ but $< 10\%$
Skin Category 3			$\geq 10\%$
(10 x Skin Category 1) + Skin Category 2		$\geq 10\%$	$\geq 1\%$ but $< 10\%$
(10 x Skin Category 1) + Skin Category 2 + Skin Category 3			$\geq 10\%$

Note to Table : Only some authorities will use the subcategories of Skin Category 1 (corrosive). In these cases, the sum of all ingredients of a mixture classified as Skin Category 1A, 1B or 1C respectively, should each be $\geq 5\%$ in order to classify the mixture as either Skin Category 1A, 1B or 1C. In case the sum of the Skin Category 1A ingredients is $< 5\%$ but the sum of Skin Category ingredients 1A+1B is $\geq 5\%$, the mixture should be classified as Skin Category 1B. Similarly, in case the sum of Skin Category 1A+1B is $< 5\%$ but the sum of Category 1A+1B+1C is $\geq 5\%$ the mixture would be classified as Category 1C.

性的場合と同様に当該混合物自体について判定に使用可能なデータがある場合には、それに基づき分類を行うとの原則が記載されている。一方、含有成分から分類判定を行う必要がある場合の方式については、当該作用を示す成分の量に従って分類を行うための表が与えられている。皮膚に関する混合物分類基準を表3に示した。皮膚ではcategory 1からcategory 3までに分類される成分がどの程度含まれているかが基準となっている。例えばcategory 1に相当する成分が5%以上含まれている場合は当該混合物をcategory 1 (corrosive) に分類することとなり、1%以上5%未満の場合はcategory 2 (irritant) と分類する。

この基準の考え方は、成分の作用についての加算性を前提とするもので、当該作用を示す成分の含有量に応じて混合物の分類を決定するものである。逆にみると、希釈割合に応じて作用強度が減少することを認める考え方となる。混合物の眼刺激性・腐食性についても同様の方式で分類することとなっており、含有量基準を示す表が定められているが、皮膚の場合と異なり、眼に対する作用については、皮膚刺激性・腐食性を示す成分についても考慮することとされ、これらの成分も加算の対象となっている。

一方、作用の加算が認められない場合（酸とアルカリの混合）や希釈割合に応じた作用の減少が認められ

ないような成分が含まれる場合が想定される。酸、アルカリ、アルデヒド類、フェノール類、界面活性剤等、加算性による評価が不適と考えられるものを考慮して、もう一つの刺激性・腐食性分類基準表が作成されており、皮膚と眼を含め基準が示されている。こちらでは、全成分についての含有量加算を行わず、該当するいずれかの成分が一定の基準値以上含有されるかどうかで category 1 あるいは category 2 に分類することとされている。また、酸やアルカリを含む場合については本文で混合物の pH 値が基準（2 以下や 11.5 以上）になると記述されているが、この表では pH 値が 2 以下や 11.5 以上の酸・アルカリが成分として 1% 以上含まれている場合は当該カテゴリーに分類することを要求している。混合物自体の pH 値による判定と成分に基づく判定で齟齬が生じる場合の判断は明快ではない（成分を問題としたい立場と反対の立場があり、議論がつくされずに決着した結果である）。

2.5 呼吸器および皮膚感作性 (respiratory/dermal sensitisation)

呼吸器および皮膚に対する感作性に関しては、それぞれ当該作用に関するヒトにおける報告あるいは動物実験における結果をもとに、これらの作用の有無を判定することとなっている。特異的呼吸器過敏症を引き起こし得るとのヒトでの証拠があること、接触皮膚炎を引き起こしたとのヒトでの報告がある程度の数なされていること、適切な動物実験で陽性結果が得られていること等を基準に分類される。一旦感作を受けて当該物質に対して過敏になっているような者に対する配慮についての議論が現在も行われており、GHS 実施前に対処すべき重要事項として位置づけられているが対応は未定である。

混合物の感作性分類基準であるが、当該作用を有する成分が一定の基準値以上含有されているかどうかにより分類が行われることとなっている。OECD 文書に示されている表からは、混合物が皮膚感作性物質を基準値以上含むことによって呼吸器感作性あり（あるいは呼吸器感作性物質を基準値以上含むことによって皮膚感作性あり）といった判断をするように読めるが、これは表の誤りである。OECD 文書をもとに作成された IOMC 版 GHS 文書（2001 年 10 月版）を参照されたい。皮膚と呼吸器の基準表は分離されており、「気体」と「固体・液体」に分けて単位も明確に定められている。皮膚感作性を有する成分が 1%（固体・液体は w/w、気体では v/v）以上含まれる場合は、当該混合物を皮膚に対する感作性物質として分類し、また呼吸器に対する感作性を有する成分が、固体・液体では 1%（w/w）以上、気体では 0.2%（v/v）以上含

まれている場合に、当該混合物を呼吸器に対する感作性ありと分類することとなっている*。

2.6 生殖細胞変異原性・癌原性 (mutations in germ cells・cancer)

これらのエンドポイントでは、ともにヒトに対して当該作用を示す証拠の確実性に依拠して、category 1 および category 2 の分類が設けられており、さらに category 1 は、category 1 A と category 1 B に細分類されている。変異原性と癌原性では表現が異なるものの、category 1 は当該作用を有することがヒトで示されているもの (category 1 A) またはヒトにおいて当該作用を示すことが主として動物実験によって示唆されているもの (category 1 B)、category 2 は category 1 への分類には根拠が不十分ながら、当該作用を示す疑いがあるものとなっており、具体的には各カテゴリーは以下のように表現されている。

・生殖細胞変異原性

CATEGORY 1 A : chemicals known to induce heritable mutations in germ cells of humans.

CATEGORY 1 B : chemicals which should be regarded as if they induce heritable mutations in the germ cells of humans.

CATEGORY 2 : chemicals which cause concern for man owing to the possibility that they may induce heritable mutations in the germ cells of humans.

・癌原性

CATEGORY 1 : known or presumed human carcinogens.

CATEGORY 1 A : known to have carcinogenic potential for humans ; the placing of a chemical is largely based on human evidence.

CATEGORY 1 B : presumed to have carcinogenic potential for humans ; the placing of a chemical is largely based on animal evidence.

CATEGORY 2 : suspected human carcinogens.

ここでの変異原性は germ cell mutagenicity（生殖細胞変異原性）とされているように、ヒトの生殖細胞の遺伝子に対する作用で、子孫にその遺伝子への影響が伝わるもの (heritable mutations in germ cells in humans) となっており「遺伝毒性」を対象とする米国や EU の基準・考え方¹⁰⁾を反映している。生殖細胞を対象とした試験、哺乳類体細胞を用いる試験、パク

* 2002 年 1 月の OECD の会議で、感作性についても癌原性等と同様に、2 段階の基準値を設ける修正案が承認され、上述の基準値に加えて、皮膚・呼吸器ともに 0.1% というより厳しい基準値が追加導入される見込み。

表4 混合物の癌原性カテゴリー分類基準

Ingredient classified as:	cut-off/concentration limits triggering classification of a mixture as:	
	category 1 carcinogen	category 2 carcinogen
Category 1 carcinogen	≥0.1 %	
Category 2 carcinogen	-	≥0.1 % (note 1)
		≥1.0 % (note 2)

Note 1: If a Category 2 carcinogen ingredient is present in the mixture at a concentration between 0.1 % and 1 %, every regulatory authority would require information on the SDS for a product. However, a label warning would be optional. Some authorities will choose to label when the ingredient is present in the mixture between 0.1 % and 1 %, whereas others would normally not require a label in this case.

Note 2: If a Category 2 carcinogen ingredient is present in the mixture at a concentration of ≥1 %, both an SDS and a label would generally be expected.

テリアを用いる試験等、国際的に認められているテストガイドライン (OECD 等) に記載のある各種の変異原性・遺伝子障害性の試験法が例示されており、これらの *in vivo* および *in vitro* の (生物) 試験による評価結果から、(ヒトの) 生殖細胞遺伝子に対する影響を推定して分類を行うこととなる。

癌原性の分類にあたっては、ヒト疫学研究および動物実験で得られている証拠に基づいて判定を行うこととなるが、その際にはヒトおよび動物実験による研究で、それぞれ「十分な証拠 (sufficient evidence)」、 「限られた証拠 (limited evidence)」が存在するかどうかを吟味するとともに、因果関係等を科学的に総合判定することとなる。判断の基準として、category 1 A では「ヒトで十分な証拠」が、category 1 B では「動物実験における十分な証拠」あるいは「ヒトと動物実験でともに証拠は限定的だが科学的総合判定から分類が適当と認められる」といった条件が、category 2 では「ヒトあるいは動物実験で限定的な証拠・category 1 への分類には不十分」といった要件が示されている。なお、癌原性に関しては、国際癌研究機関 (IARC) の分類との対応が重要と思われる。IARC の総合評価では、グループ 1 (発癌物質：ヒトで十分な証拠がある)、グループ 2 A (おそらくは発癌物質：ヒトの証拠は限定的だが動物実験では十分な証拠)、グループ 2 B (発癌物質の可能性：ヒトでは限定的な証拠で動物実験データは十分ではない等) といったグループ分けがなされている¹¹⁾が、OECD・GHS の基準では「十分な証拠」、「限定的な証拠」といった用語を IARC と同様の定義で使用することとしており、両分類基準間の対応を考えることができる。

混合物に対する分類基準では、変異原性・癌原性ともに、category 1 に該当する成分を一定の基準値以上含有する場合に category 1 に、また category 2 に該

当する成分を基準値以上含有する場合に category 2 に分類する方式となっている。変異原性では category 1 物資を 0.1 % 以上含めば category 1 に、category 2 物資を 1 % 以上含めば category 2 に分類する。癌原性の場合の基準値は表 4 に示したものとなっている。癌原性に関する基準値決定にあたっては欧米間の主張に隔たりがあり、安全性データシート SDS の記載を目的とした分類には低めの (厳しい) 基準値を、ラベル表示では高めの (緩い) 基準値を設けるとの妥協方式が採用され、癌原性 category 2 には二種類の数値が示されている。すなわち、category 2 に相当する成分を 1.0 % 以上含有する混合物では SDS およびラベル表示の双方を求めるものの、0.1 % 以上 1.0 % 未満の場合にはラベルへの表示は任意 (option) とされた。このような方式は他のエンドポイントでも採用されている。

2.7 生殖毒性 (reproductive toxicity)

生殖毒性には、親世代の生殖機能・生殖能力への影響、胚・胎児への影響、次世代発達の影響を含むこととされている。当該作用有無の確からしさに準拠して分類が行われ、確実性の高い順に category 1 A (known)、category 1 B (presumed)、category 2 (suspected) となっており、category 1 A では主としてヒトにおける証拠に基づく、category 1 B は主として動物実験による証拠に基づくこととされている。動物実験データの解釈では、一般毒性発現の結果として生じた二次的影響ではないことが必要とされる。また、母乳を通じて次世代に影響を与える物質および母乳 (分泌) に影響する物質を対象とした category が別途設けられている。

混合物については、変異原性・癌原性同様、category 1 に該当する成分を基準値以上含有する場合を category 1 とし、また category 2 に該当する成分を

基準値以上含有する場合に category 2 と分類する方式となっている。また SDS とラベル表示で異なった基準値を採用する方式も採用されており、category 1 の基準値としては 0.1% (SDS 記載を要求) と 0.3% (SDS 記載とラベル表示を要求) が、category 2 の基準値としては 0.1% (SDS 記載を要求) と 3.0% (SDS 記載とラベル表示を要求) が採用されている。

2・8 標的臓器/全身毒性 (target organ/systemic toxicity : TOST)

単回曝露によるものと反復曝露によるものが章を分けて記載されているが、対象とする有害作用の範囲は同一のものであり、一つまたは複数の器官 (肝臓、腎臓等の臓器あるいは神経系、免疫系、循環器系等のシステム) において発現される標的臓器毒性あるいは (特定臓器に限定されず) 全身的レベルで生じる障害

を想定している。分類の結果の使用では該当する物質について、肝毒性物質、腎毒性物質、神経毒性物質等、標的となった臓器に因んだ呼称を用いることが求められている。

分類判定ではヒトにおける証拠や生物試験結果を使用することが示唆されているが、作用有無の確かさばかりではなく作用の強さも考慮することとなっている。生物試験に関して、単回曝露試験や 90 日間反復曝露試験といった試験方法への言及があり、毒性発現につながった投与量によって当該作用の強度を把握することも想定されている。カテゴリ分類基準は、作用有無の確かささと作用の強度の双方を含んだ記述となっており、疫学調査でヒトに対する毒性 (significant toxicity in humans) が明らかにされている場合および動物実験において比較的低い用量で毒性発現が認められる等ヒトに対する毒性が推定される場

表 5 単回曝露による標的臓器/全身毒性分類のためのガイダンス値

Route of exposure	Units	Guidance value ranges for:	
		Category 1 classification	Category 2 classification
Oral (rat)	mg/kg bw	$c \leq 300$	$2\,000 \geq c > 300$
Dermal (rat or rabbit)	mg/kg bw	$c \leq 1\,000$	$2\,000 \geq c > 1\,000$
Inhalation (rat) gas	ppm	$c \leq 2\,500$	$5\,000 \geq c > 2\,500$
Inhalation (rat) vapour	mg/l	$c \leq 10$	$20 \geq c > 10$
Inhalation (rat) dust/mist/fume	mg/l/4 h	$c \leq 1.0$	$5.0 \geq c > 1.0$

表 6 反復曝露による標的臓器/全身毒性分類のためのガイダンス値 category 1

Route of exposure	Units	Guidance values (dose/concentration)
Oral (rat)	mg/kg bw/d	10
Dermal (rat or rabbit)	mg/kg bw/d	20
Inhalation (rat) gas	ppm/6 h/d	50
Inhalation (rat) vapour	mg/litre/6 h/d	0.2
Inhalation (rat) dust/mist/fume	mg/litre/6 h/d	0.02

category 2

Route of exposure	Units	Guidance values ranges: (dose/concentration)
Oral (rat)	mg/kg bw/d	10-100
Dermal (rat or rabbit)	mg/kg bw/d	20-200
Inhalation (rat) gas	ppm/6 h/d	50-250
Inhalation (rat) vapour	mg/litre/6 h/d	0.2-1.0
Inhalation (rat) dust/mist/fume	mg/litre/6 h/d	0.02-0.2

合 (presumed to have the potential to produce significant toxicity in humans) を category 1 に、また動物実験で中程度の用量で毒性が認められる等ヒトの健康に対して有害と推定される場合 (presumed to have the potential to be harmful to human health) を category 2 に分類する方式が採用されている。生物試験に関しては、単回曝露、反復曝露 (90 日間試験) に対応して、それぞれ基準となる投与用量 (曝露濃度) の値が表 5 および表 6 に示したように指針値 "Guidance Value" として定められており、これを基準用量として特定の毒性発現が認められるかどうかで分類判断ができるようになっている。

混合物では、表 7 に示したように、category 1 に該当する成分を基準値以上含有する場合を category 1 とし、また category 2 に該当する成分を基準値以上含有する場合を category 2 とする。ラベル表示と SDS で 2 段階の基準値を採用する方式も採用されており、含有 1 % 以上では SDS 記載が求められ、10 % 以上では SDS 記載とラベル表示両方を行うこととな

っている。また、本エンドポイントでは作用の強度が分類基準の一つとなっているため、希釈の概念 (希釈によって毒性のより弱いクラスとなる) が入ることとなり、category 1 に分類される成分を 1 % 以上 10 % 未満含有する場合には、この混合物を category 2 に分類してもよいこととされている。

2.9 水生環境への有害作用 (hazardous for the aquatic environment)

水環境・水生生物への有害作用については、蓄積・分解性やミジンコ・魚・藻類を用いた試験データの組み合わせによる複雑な分類基準 (急性影響・慢性影響) が作成されている。混合物の扱いについては、当該混合物についてのデータがない場合には、各成分の有害性指標値 (LC₅₀・EC₅₀) から混合物の毒性指標値を計算式により求めて分類を行う方式 (Additivity Formula) と、どの分類カテゴリーに相当する成分がどの程度含まれているかによって分類を行う方式 (Summation Approach) を併用するようになっている。また、後者の方法を用いるに際しては、L(E)C₅₀

表 7 混合物の標的臓器/全身毒性カテゴリー分類基準

Ingredient classified as:	cut-off/concentration limits triggering classification of a mixture as:	
	category 1 Target Organ Systemic Toxicant (TOST)	category 2 Target Organ Systemic Toxicant (TOST)
Category 1 (TOST) Target Organ Systemic Toxicant	≥1.0 % (note 1)	1.0 ≤ ingredient < 10 % (note 3)
	≥10 % (note 2)	1.0 ≤ ingredient < 10 % (note 3)
Category 2 (TOST) Target Organ Systemic Toxicant		≥1.0 % (note 4)
		≥10 % (note 5)

Note 1: If a Category 1 target organ/systemic toxicant is present in the mixture as an ingredient at a concentration between 1.0 % and 10 %, every regulatory authority would require information on the SDS for a product. However, a label warning would be optional. Some authorities will choose to label when the ingredient is present in the mixture between 1.0 % and 10 % whereas others would normally not require a label in this case.

Note 2: If a Category 1 target organ/systemic toxicant is present in the mixture as an ingredient at a concentration of ≥10 %, both an SDS and a label would generally be expected.

Note 3: If a Category 1 target organ/systemic toxicant is present in the mixture as an ingredient at a concentration between 1.0 % and 10 %, some authorities classify this mixture as a Category 2 target organ/systemic toxicant, whereas others would not.

Note 4: If a Category 2 target organ/systemic toxicant is present in the mixture as an ingredient at a concentration between 1.0 % and 10 %, every regulatory authority would require information on the SDS for a product. However, a label warning would be optional. Some authorities will choose to label when the ingredient is present in the mixture between 1.0 % and 10 %, whereas others would normally not require a label in this case.

Note 5: If a Category 2 target organ/systemic toxicant is present in the mixture as an ingredient at a concentration of ≥10 %, both an SDS and a label would generally be expected.

値が1 mg/l以下の高い毒性を示す物質が含まれている場合に当該成分の濃度に重み付け係数(M値)を掛けることとなっており、L(E)C₅₀値に応じたM値が一覧表で与えられている。さらに、方法によって同一の混合物が異なるクラスに分類される場合には安全側(more conservative)を選択することとされた。

2・10 新規追加が検討されているエンドポイント

水との反応により有害気体を発生する物質(Substances Which in Contact with Water Emit Corrosive or Toxic Gases or Vapours)、液体の吸引(Aspiration)による有害作用、呼吸器への刺激性(Respiratory Tract Irritation)、麻酔作用(Narcotic Effects)について新たに検討することとなっている。これら新規エンドポイントに関する分類基準の作成および既決分類基準の改定・見直しについては、引き続きOECDにおいて作業が行われる予定である。

3. 終わりに

国際的に統一された基準によって化学品の毒性・有害性にかかわる分類・記述が行われることは、労働者の安全確保や一般消費者保護の観点から大きな意義がある。数年以内に、国内の諸制度をGHSに対応したもものとして整備する必要があると同時に、個別の製品についてGHS分類が行われSDSに記載されることになることとすると、有害化学物質の管理にかかわる専門家が本分類基準に精通することが求められよう。国内の関連業界においてもGHSの普及に対する準備・対応の動き^{12,13)}があるが、一般にはGHS導入についてあまり知られていない。本稿がGHSの理解・普及に貢献することを希望する。正確な理解のためには是非原典^{5,6)}にあたっていただきたい。

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Review

Toxicologic/carcinogenic Effects of Endocrine Disrupting Chemicals on the Female Genital Organs of Rodents

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Abstract: Toxicologic/carcinogenic effects of some representative endocrine disrupting chemicals (EDCs) having estrogenic activity, such as alkylphenols, on the female genital organs of rodents, especially rats, are reviewed and discussed, focusing on our recent research. Neonatal treatment of high-dose p-tert octylphenol (t-OP, 100 mg/kg s.c. injection every other day from postnatal day 1 (PND 1) to PND 15) induced various long-term persistent irreversible effects on the female reproductive system of Donryu rats, such as lower gonadotropin levels at prepuberty, inhibition of uterine gland genesis, persistent estrus and polycystic ovaries. The result indicates that neonatal high-dose treatment of estrogenic EDCs can affect gonadotropin secretion during the developmental period of sexual maturation with direct masculinization of the hypothalamic function. Exposure limited to the first 5 days after birth (PNDs 1–5) to 100 mg/kg t-OP, however, caused delayed influence which was characterized by accelerated appearance of atrophic ovary, manifested by early-occurring and long-term continuing persistent estrus after puberty, whereas no abnormalities could be found with regard to growth and differentiation of the reproductive organs and the hypothalamo-pituitary-gonadal control system up to maturation, the influence being caused by delayed modulation of the hypothalamo-pituitary-gonadal control system. The most notable effect on the female reproductive system when normal cycling rats were exposed to high-doses t-OP for 28 days, was disappearance of the estrous cycle, but no clear changes were detected in other parameters such as uterine weight and morphology. These results indicate that the most serious issue with EDCs is the potential effects of prenatal and/or neonatal exposure on rodents. Well or moderately differentiated adenocarcinomas increased in Donryu rats initiated with N-ethyl-N'-nitro-N-nitrosoguanidine, when high-dose t-OP was given subcutaneously during adulthood. Neonatal exposure for PNDs 1–5 to high-dose t-OP also showed promoting effects on uterine adenocarcinoma development. However, in rats given t-OP for PNDs 1–15, uterine tumor malignancy was clearly increased, although there was no significant alteration in the total incidence of adenocarcinomas. The results are very interesting in consideration of the histogenesis of uterine adenocarcinomas. However, maternal exposure to low doses of EDCs such as nonylphenol and bisphenol A at actual human exposure levels by the oral route showed no effects on growth and development of the female reproductive system or uterine carcinogenesis. These results indicate that dietary exposure to low doses of EDCs might not induce any adverse effects on the female genital system in mammals including humans. (*J Toxicol Pathol* 2004; 17: 69–83)

Key words: endocrine disrupting chemicals (EDCs), toxicity/carcinogenicity, female genital organs, rodents

Introduction

Recently, the possible adverse consequences arising from the release of man-made substances with estrogenic, anti-estrogenic or androgenic properties, so called endocrin

disrupting chemicals (EDCs), into the environment have become an important environmental problem. There is much concern that these EDCs may have the potential to disturb normal sexual differentiation and development in wild life and mammals, including humans, and exert various deleterious effects on many organs, with carcinogenic effects being particularly important in mammals. The genital organs are the obvious target organs of various EDCs, and various toxicologic changes have been reported to be induced in both male and female genital organs of rodents. Unfortunately, however, there is less information

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on females than males, although many EDCs have estrogenic properties.

In the present report, toxicologic/carcinogenic effects of some representative EDCs, such as alkylphenols, on the female genital organs of rodents, especially rats, are briefly reviewed and discussed from the point of view of extrapolation to humans, mainly focusing on our recent research. In our studies of the effects of EDCs, Donryu rats were mainly used. The Donryu rat is a unique domestic strain having a regular 4-day estrous cycle at the juvenile stage. After 5 months of age, however, persistent estrus appears and increases age-dependently, and a high occurrence of spontaneous uterine adenocarcinomas is observed at about 2 years of age or thereafter (Table 1)¹. In

this rat strain, the early appearance of persistent estrus results in an increase in the estrogen (E2):progesterone(P) ratio (E2:P ratio). In humans, it has been reported that relatively high E2:P values increase the endometrial cancer risk^{2,3}. Using this strain, we recently demonstrated effects of reproduction on uterine carcinogenesis, in line with the known lower risk of uterine adenocarcinoma in multiparous as compared to nulliparous or infertile women. The incidence of spontaneous endometrial adenocarcinomas showed a tendency to decrease in animals having three reproductive experiences, compared to the nulliparous case, although the incidence was not influenced by a single pregnancy⁴. Thus, this rat strain is a good animal model for endometrial adenocarcinoma development due to the imbalance of endogenous steroid hormones, as found in humans. We also succeeded in obtaining high induction of tumors in this strain by single intra-uterine administration of N-ethyl-N'-nitro-N-nitrosoguanidine (ENNG), thereby establishing a two-stage uterine carcinogenesis model (Fig. 1)^{1,5}. Quite recently, Vollmer⁶ reviewed experimental endometrial cancer models, including Donryu rats, and considered them useful for studies on molecular aspects of endometrial cancer and carcinogenesis.

Table 1. Persistent Estrus and Spontaneous Uterine Tumors in Donryu and F344 Rats*

Sequential Changes in Persistent Estrus Incidences in Female Donryu and F344 Rats

Strain	Incidence (%)						
	4	5	6	8	10	12	15 (Months of age)
Donryu	0	17	32	64	87	88	85
F344	0	0	2	0	6	11	4

Main Spontaneous Uterine Tumors in Donryu and F344 Rats

Uterine tumors	Incidence (%)	
	Donryu	F344
Mean survival time (weeks)	108.8 (62-120)	114.1 (60-131)
Endometrial adenocarcinoma	35	1
Endometrial stromal polyp	1	21

*: Maekawa *et al.*, *J Toxicol Pathol* 1999; 12: 1-11.

Classification of EDCs

Various chemicals have been shown to have endocrine disrupting effects not only on wildlife but also mammals including humans. Major representative EDCs are as follows, according to their use, chemical structures and/or chemical characteristics.

1. DDT (1,1,1-trichloro-2,2-bis(4-chlorophenyl)ethane), DDE (1,1'-(dichloro-ethenylidene)bis(4-chlorobenzene)), dieldrin: agricultural chemicals (insecticides) with properties of high-accumulation and resistance to degradation.

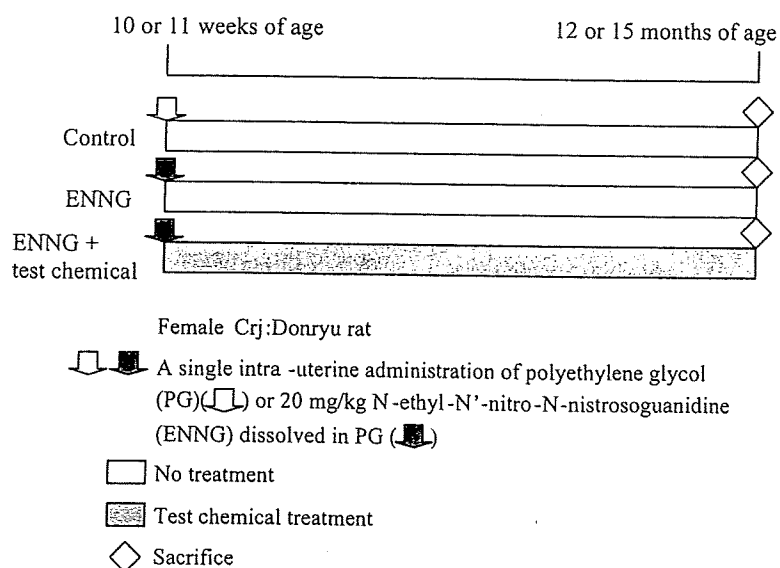


Fig. 1. Two-stage rat uterine carcinogenesis model

2. PCBs (polychlorinated biphenyl), PBB (polybrominated biphenyl): industrial chemicals (insulators etc.) which accumulate and are difficult to degrade.
3. DEHP (di(2-ethylhexyl)phthalate), alkylphenols such as nonylphenol and octylphenol, bisphenol A: industrial chemicals widely used as plasticizers or surfactants.
4. Dioxins such as TCDD (2,3,7,8-tetrachlorodibenzo-p-dioxin), dibenzofuran: chemicals naturally produced by dust-incineration.
5. TBT (tributyltin), TBTO (tributyltin oxide): industrial chemicals used for coating of ships' bottom.
6. DBCP (1,2-dibromo-3-chloropropane), atrazine, vinclozolin: agricultural chemicals.
7. DES (diethylstilbestrol), tamoxifen, oral contraceptives: medical drugs.

Development of the Female Genital Organs and Profile of Hormonal Secretions in Rodents

In general, development of the female genital organs in rats is roughly classified into 3 stages, prenatal (embryonic), neonatal/juvenile and adult/aged. The prenatal stage is from the day of fertilization, i.e., day 1 post-coitum till birth (about 20–22 days in rats). The embryonic bipotential gonad develops from mesoderm in the gonadal ridge located on the dorsal coelomic walls. The primordial germ cells and the gonadal ridge are visualized as condensations of cells localized ventral to the mesonephrons by gestational day 12 in rodents. In the embryonic developmental stage, two sets of paired tubular organs develop: the Wolffian ducts and the Mullerian ducts. It is well established that the presence or absence of functioning embryonic testes plays a major role in determining which duct system undergoes further development. In the rat, the critical time period for Mullerian duct development covers days 14–18 of gestation. On day 18, in the absence of testicular hormones including anti-Mullerian hormone (AMH) from Sertoli cells, the Mullerian ducts undergo further development and the Wolffian ducts degenerate.

The female reproductive tract of rodents is immature at birth and the developing uterus undergoes a period of rapid growth and differentiation during the first 2 weeks of postnatal life. In rats, the uterus at birth corresponds developmentally to the fetal uterus at gestation day 100 in human beings⁷. During this period, luminal epithelial cells invaginate into the underlying stroma to form uterine glands⁸. The uterine growth phase in this period coincides with an elevation of serum estradiol levels beginning on postnatal day (PND) 9. Thus, the role of endogenous estrogen (17β -estradiol, E2) and its receptor (ER) are very important for uterine growth and differentiation. In normal rats, ER expression in the uterine epithelium appears at various days from PND 7 to PND 15⁹.

In female rats, serum FSH (follicle-stimulating hormone) levels rise to a peak at PNDs 15–16 followed by an abrupt nadir, while LH (luteinizing hormone) concentrations are high at PNDs 2–10 followed by gradual decline during

sexual maturation; E2 levels also rise to a peak during the first 2 weeks of age^{10,11}. On the other hand, α -fetoprotein, the estrogen-binding protein produced in the liver, is found in very high concentrations for several weeks after birth¹². It is well known that the increase of FSH before puberty is caused by nullification of the negative feedback of estrogen because of the high concentrations of serum α -fetoprotein¹³.

A striking sexual dimorphism in gross morphology of the medial preoptic area (sexually dimorphic nucleus of the preoptic area: SDN-POA) has been recognized in the rat brain¹⁴. The development of this nucleus starts during late fetal life and depends on the hormonal environment at the critical period of sexual differentiation^{14–16}. In genetic males, the relatively high levels of perinatal testosterone are aromatized to estradiol in the nervous cells of SDN-POA and the estrogenic signals may be directly responsible for the increase of SDN-POA volume. In genetic females, in contrast, estrogenic effects on SDN-POA are prevented because estrogen is bound to the serum binding protein, α -fetoprotein, during the late embryonic and early neonatal periods. The female SDN-POA is smaller than that of males as a result of an orchestrated pattern of decreased cell proliferation and/or increased programmed cell death^{17,18}. It is well known that the SDN-POA volume of genetic females becomes larger than normal on perinatal exposure to testosterone or high amounts of some estrogenic compounds¹⁹. Analogues of SDN-POA have also been identified in various animal species such as the gerbil, Guinea pig, ferret, quail and human²⁰. Recently, another sexual dimorphism had been demonstrated in the anteroventral periventricular nucleus of the preoptic area (AVPvN-POA) and the locus coeruleus^{21,22}. The volumes of these are larger in females than males, but a direct correlation with the hormonal environment has not yet been clarified.

In the rat brain, ER α expression is found primarily in ventral midline structures such as bed nucleus of the stria terminals, hypothalamic medial preoptic area, hypothalamic ventromedial nucleus, hypothalamic arcuate nucleus, septohypothalamic nucleus, septum and central gray area of the midbrain. ER β is similarly distributed in the brain and is additionally detected in the paraventricular nucleus of the hypothalamus and the hippocampus^{23,24}.

After weaning, the first estrous cycle starts at about 36 days of age, and the minimum breeding age is about 84 days of age²⁵. The estrous cycle is characterized by cyclic changes of the epithelial surface in the vagina and the uterus. The estrous cycle in the rat lasts 4–5 days and is divided into proestrus, estrus, metestrus and diestrus. Uterine weights increase with luminal excretion from diestrus to proestrus, showing a peak at proestrus, but decrease at estrus and metestrus.

In normal cycling rats, the E2 and P levels are highest at proestrus, corresponding with the increased uterine weights. Thereafter, the E2 level drops toward estrus and slightly increases again at diestrus. The P value increases slightly at metestrus, although it is low at estrus and diestrus. At

proestrus and diestrus, and especially the former, the luminal and glandular epithelial cells along with stromal cells beneath the luminal epithelium are strongly positive for ER- α mRNA expression. At estrus, the expression is slightly diminished in luminal cells, but is almost completely lacking in glandular cells. At metestrus, positive signals appear again in the latter. In the myometrium, expression is constant in all estrous stages. Thus cell-type specific patterns of ER-mRNA expression characterize the uteri of normal estrous cycling rats²⁶.

In general, the adult stage in rats is from minimum breeding age to maximum age (about 360–450 days), and thereafter the aged stage starts²⁵. In Donryu rats, however, estrous cycle abnormalities increase age-dependently after 26 weeks of age, and almost all animals show persistent estrus at 52 weeks of age. In contrast, vaginal smears of F344 rats indicate a normal estrous cycle until 52 weeks of age²⁷. In our studies, various histological changes such as follicular cysts and atrophic changes such as absence of corpora lutea in the ovary and cornification of epithelium in the vagina in Donryu rats were observed to be linked with persistent estrus, and increased with time, especially after 10 months of age. In F344 rats, in contrast, atrophy of the ovary is observed in only a few animals at 15 months of age. As a result of ovarian atrophy, in Donryu rats, the plasma values of E2 and P, and especially the latter, decrease with age, the E2:P ratio becoming elevated, and the bromodeoxyuridine (BrdU)-labeling indices of uterine endometrial cells are age-dependently increased²⁸. This age-related hormonal imbalance and the constant high level of proliferating activity of epithelial cells are considered to play important roles in high yield development of spontaneous uterine endometrial adenocarcinomas in this rat strain^{1,28}.

Toxicologic and/or Carcinogenic Effects of EDCs on the Female Genital Organs of Rodents

Rodents in the first 2 weeks of postnatal life, termed “a critical point” or “a window of vulnerability”, are very sensitive to exogenous estrogens and androgens including EDCs, because the reproductive tract undergoes rapid growth and differentiation within this period, as mentioned above. Thus, the OECD (Organization for Economic Cooperation and Development) recently proposed the immature rat uterotrophic assay as one of the screening test methods for the detection of estrogenic or anti-estrogenic properties of chemicals²⁹. In studies using adult animals, the ovariectomized (OVX) animal model is also effective for the detection of estrogen agonists, because the effects of endogenous estrogen can be minimized³⁰. In various toxicity studies using adult animals, oral administration has generally been used to assess the toxicity of chemicals, and the OECD has also proposed a new 28-day repeated oral-dosing toxicity test protocol using adult rats, the enhanced OECD TG407 protocol, for the assessment of the toxic effects of EDCs. For the detection of endocrine disrupting activity of direct-acting chemicals, however, other administration routes such

as subcutaneous injection may provide greater sensitivity than oral administration, because this eliminates the direct effects of metabolism of the chemicals during first passage through the liver.

As mechanisms for the biological effects of EDCs on their target organs, their binding to growth factor receptors and arylhydrocarbon (Ah) receptors as well as steroid receptors has been considered to be very important. Furthermore, some chemicals have been shown to have effects on endogenous estrogen metabolism, resulting in disturbance of the hormonal milieu.

Effects of High-doses of EDCs Effects on Growth and Development of Female Reproductive Organs Prenatal and/or Neonatal Exposure

Inappropriate exposure to estrogens and also EDCs in the prenatal and/or neonatal period has been well established to exert irreversible influence directly and indirectly on the female reproductive system^{31,32}. “Androgenization” is characterized by direct modulation of the hypothalamo-pituitary-gonadal control system, resulting in lowering of gonadotropin levels and persistent estrus as an indirect effect, and abnormal uterine/vaginal development and/or growth as direct influences.

Alkylphenolic compounds are derived from biodegradation of nonionic surfactants, alkylphenol ethoxylates, which are widely used as detergents in many industries. Alkylphenol ethoxylates are also broken down in the process of sewage-treatment or in rivers into alkylphenols, such as nonyl or octylphenol (NP or OP), which are well known representative EDCs with weak estrogenic activity, acting via binding to ER. In vitro data indicate that OP has the most potent estrogenic activity of the alkylphenols (approximately 1000 times less estrogenic than E2), although NP is detected with higher levels than OP in the environment.

In our studies of the toxicologic/carcinogenic effects of EDCs on the female genital organs, OP was selected as a representative compound. It has already been reported that neonatal treatment with OP disrupts estrous cyclicity after weaning in female rats³³. We also examined the effects of neonatal exposure to a high dose of p-tert octylphenol (t-OP) on the female genital organs of Donryu rats³⁴, and the results were in line with those of other papers: long-term persistent irreversible effects such as lower gonadotropin levels at prepuberty, inhibition of uterine gland genesis, persistent estrus shown by vaginal cytology, and polycystic ovaries. In our recent study, newborn female pups were injected with 100 mg/kg t-OP subcutaneously within 24 h of birth. Administration was repeated every other day until PND 15 (PNDs 1–15), and animals were observed till PND 77. Histologically, inhibition of uterine gland genesis was apparent during the immature period before weaning. The day of vaginal opening was about 4 days earlier in OP-treated animals than in controls, and after vaginal opening,

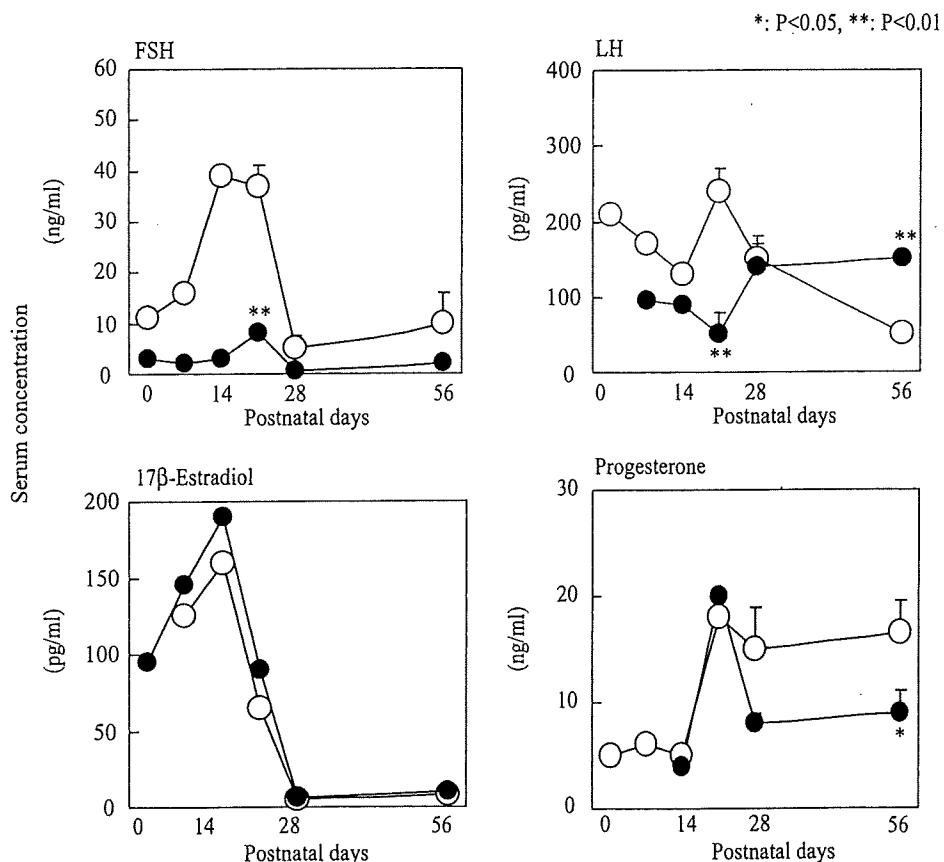


Fig. 2. Serum gonadotropins and sex steroid hormones in control and PNDs 1–15 OP-treated rats. Open circles (○), controls; black circles (●), PNDs 1–15 OP-treated. Katsuda *et al.*, *Toxicol Appl Pharmacol* 2000; 165: 217–226.

none of the OP-treated rats showed a regular estrous cycle, and persistent estrus was ultimately observed in these animals. Atrophic and polycystic ovaries without corpora lutea were anovular. In the endometrium, cell-proliferative activity and cell-death were increased and decreased, respectively, and expression of estrogen receptor alpha mRNA was apparent on in situ hybridization. At 8 weeks of age, treated animals exhibited luminal epithelial hyperplasia with overexpression of ER-mRNA. During the immature period, serum FSH and LH levels were consistently lower in OP-treated rats than in controls. In particular, serum FSH levels remained uniformly low. Serum E2 levels demonstrated essentially the same pattern as in controls, being elevated at PND 14, and then falling to low levels. After weaning before sexual maturation, FSH values in treated rats remained low, while those of control animals decreased rapidly and were maintained at the same levels as in the OP-treated case. In contrast, LH levels of treated animals increased after weaning and remained high until the end of the experiment (PND 77). Serum P levels of both OP-treated and control rats were constant, but the level in the former was only half of the latter value (Fig. 2). Serum inhibin levels of OP-treated rats were nearly the same as in controls at PND 28. The results resembled those of male or

androgenized female rats in the secretory pattern of gonadotropins at this age^{10,11}, indicating that neonatal treatment with high-dose t-OP affects gonadotropin secretion during the developmental period of sexual maturation with direct masculinization of hypothalamic function.

In another of our studies, neonatal exposure for the first 2 weeks (PNDs 1–15) to 100 mg/kg t-OP induced an early and enhanced ER expression in the luminal epithelium compared with age-matched controls, and increased proliferating cell nuclear antigen (PCNA) positive cells, though expression in the glandular epithelium was suppressed in relation to inhibited gland-genesis. Therefore neonatal exposure to high doses of EDCs with estrogenic activity can induce abnormal differentiation in the developing rat uteri via abnormal ER expression and subsequent alteration of cell proliferating activity³⁵.

Recently, however, it has been reported that prenatal and/or neonatal exposure to high doses of estrogens or EDCs with estrogenic activity also exerts a “delayed” influence, different from that of typical androgenization. The delayed influence is probably caused by delayed modulation of the hypothalamo-pituitary-ovarian control system³⁶. A number of investigators have described effects of neonatal exposure

Table 2. Uterine Gland Genesis before Puberty in Control and PNDs 1–5 or PNDs 1–15 OP-treated Rats*

	No. of uterine gland / section (Mean \pm SD)		
	Control	PNDs 1–5	PNDs 1–15
PND 10	0	0	0
PND 14	3.94 \pm 0.5	4.05 \pm 1.5	0.1 \pm 0.13**
PND 21	4.58 \pm 0.6	5.57 \pm 1.7	2.55 \pm 1.5**
PND 28	6.42 \pm 1.5	7.83 \pm 1.3	3.14 \pm 2.2**

*: Yoshida *et al.*, Carcinogenesis 2002; 23: 1745–1750.

** : Significantly different from the control value (P<0.05).

Table 3. Sequential Changes in Incidences of Persistent Estrus in Control and PNDs 1–5 or PNDs 1–15 OP-treated Rats*

Group	Incidence of persistent estrus (%)								
	1.5	2	3	4	5	6	8	10	11 (Months of age)
Control	0	0	0	2.6	17.9	30.8	64.1	85.7	100
OP-treated (PNDs 1–5)	4.9	12.2	53.7**	70.1**	87.8**	100**	100**	100	100
OP-treated (PNDs 1–15)	100	100	100	100	100	100	100	100	100

*: Yoshida *et al.*, Carcinogenesis 2002; 23: 1745–1750.

** : Significantly different from the control value (P<0.05).

to EDCs including estrogens or androgens, but information on such delayed effects is limited. In our recent study, exposure after birth to 100 mg/kg t-OP for the first 5 days (PNDs 1–5) caused a “delayed” influence which was characterized by accelerated appearance of atrophic ovary, manifested by early-occurring and long-term continuing persistent estrus, whereas no abnormalities could be found with regard to growth and development of the reproductive organs and the hypothalamo-pituitary-gonadal control system up to maturation³⁷, thus differing from the case of exposure for PNDs 1–15 to the same dose of t-OP³⁴ (Tables 2 and 3). Previously, we confirmed neonatal OP-treatment of 50 mg/kg/day every other day for PNDs 1–15 did not affect estrous cyclicity³⁴, the total administration-dose (400 mg/kg) being higher than that (300 mg/kg) in the PNDs 1–5 study. This result indicates that the differences were due to the treatment period, rather than the total dosing volume.

Postnatal Exposure

Chronic administration of OP to adult male rats causes alteration in hormonal secretions³⁸, and also induces atrophies of the testis and other genital organs³⁹. We therefore tested estrogenic effects of t-OP using adult OVX Donryu rats given daily subcutaneous injections of 6.25, 12.5, 25, 50 or 100 mg/kg for 2 or 14 days. t-OP was detected in serum at doses of 25 mg/kg and above for 2 days and of 12.5 mg/kg and above for 14 days, and uterine weights and luminal epithelial heights were increased dose-dependently. OP-treatment for 2 days caused a dose-related increase in proliferation of uterine luminal, glandular and stromal cells and vaginal epithelial cells, and the effects were fundamentally related to the serum OP levels⁴⁰.

Effects of t-OP on the female reproductive tract of

normal cycling rats were also investigated. F344 and Donryu rats were used, and t-OP was subcutaneously injected for 28 days at similar concentrations to those applied to OVX rats. The most notable changes were disappearance of normal cyclicity in 50 mg/kg or more OP-treated rats and appearance of persistent estrus in the 100 mg/kg group. In rats showing abnormal cyclicity, the uterine morphology deviated from the normal at each estrous stage of cycling rats, and cell proliferation in the endometrium was slightly increased. However, the data for uterine weights, luminal epithelial cell heights and/or numbers of epithelial cells in the endometrium demonstrated only equivocal alteration. In treated rats, the serum E2 levels were decreased with 50 mg/kg of OP or more. Donryu and F344 rats showed similar sensitivity to estrogenic effects of OP, no strain difference being evident. The results indicate that vaginal cytology may be the most sensitive endpoint for the detection of estrogenic activity of potential EDCs in studies using adult female rats⁴¹. It was also demonstrated that vaginal cytology or its morphological features might be very useful in animal toxicity studies for assessment of the individual hormonal milieu including dysfunction of the hypothalamo-pituitary-gonadal control system⁴².

The suitability of the 28-day repeated oral-dosing study for risk assessment of EDCs or strain differences was investigated in adult SD, F344 and Donryu female rats given 60 or 250 (150) mg/kg/day of NP, or 5 or 50 mg/kg/day of atrazine by stomach tube for 28 days. No morphological changes were noted in any reproductive organs of the treated animals, although abnormal estrous cycles were detected in high-dose groups of all strains, without any strain differences⁴³. The results also indicate that vaginal smear is the most sensitive parameter for detection of effects of estrogenic or anti-estrogenic chemicals, when normal

cycling animals are used. Although atrazine is an agrochemical having weak estrogen-antagonistic activity, an anti-estrogenic property was not clear in the study. However, effects were detected in the immature rat uterotrophic assay, in which atrazine alone was not associated with any changes in uterine weight, but co-treatment with atrazine and E2 reduced E2-induced increase of uterine weight⁴⁴.

Effects on Uterine Carcinogenesis

While the etiology of uterine adenocarcinomas in women is still inconclusive, hormones such as estrogens are considered to be of essential importance^{2,3}. The carcinogenic effects on the female genital tract in mammals, including humans, are considered to be one of the most important adverse consequences of EDCs with estrogenic activity. However, there have been only a few reports of unequivocal induction of carcinomas in experimental animals by EDCs, except with diethylstilbestrol (DES), as reviewed previously¹. In humans, the causation of vaginal and uterine cancers by prenatal exposure to DES is a striking example of environmental carcinogenesis⁴⁵. In experimental animals also, the effects of prenatal DES exposure have been studied in rats and mice, as reviewed by Marselos and Tomatis⁴⁶. Vaginal and uterine adenocarcinomas were induced in mice exposed prenatally to DES^{47,48}. In rats following in utero DES exposure, however, mammary and vaginal tumors, rather than uterine tumors, were observed⁴⁶. Thereafter, uterine carcinomas were also induced in Donryu rats by transplacental administration of DES⁴⁹. In the study, interestingly, data for persistent estrus incidence indicate a "delayed" influence in offspring exposed prenatally, similar to our recent report³⁷.

Tamoxifen (TAM) is a non-steroidal anti-estrogen which competes with estrogen for binding to ER. However, its pharmacology is very complex, and both estrogen agonistic and antagonistic properties have been found, depending on the species, age, exposure duration, dose, route and organs in experimental studies⁵⁰. It has been pointed out that the risk of endometrial cancer may be increased in postmenopausal women exposed to TAM for mammary cancer therapy, the agent acting on the uterus as a weak estrogen agonist^{51,52}. In experimental studies using adult rats and mice, however, it has been impossible to cause endometrial cancers by TAM treatment, although endometrial carcinomas were induced in mice treated neonatally⁵³. Also the incidences of uterine and cervical/vaginal cancers increased in rats, in the absence of any estrogen agonistic effect, when tamoxifen was administered orally on days 2–5 after birth⁵⁴. Previously we reported that TAM showed potent anti-estrogenic effects on the adult rat uterus and inhibited the development of endometrial adenocarcinomas in our two-stage uterine carcinogenesis model⁵⁵. In that study, however, the dose levels used might have been high. Quite recently, we also reported that TAM showed promotion, but not progression, effects on mouse

uterine carcinogenesis, so that the influence in the progression stage appears to be different from the estrogen agonism reported for human beings, although TAM did show estrogen agonistic effects in the promotion stage⁵⁶.

In one study, atrazine slightly increased the incidence of endometrial adenocarcinomas in female F344 rats, when given in the diet⁵⁷. Quite recently, however, it was reported that atrazine administered in diet has no modifying effects on uterine carcinogenesis in ICR mice initiated with N-ethyl-N-nitrosourea⁵⁸. Vinclozoline, a pesticide also showing an anti-estrogenic effect, induced uterine adenocarcinomas in female Wistar rats, as well as ovarian sex cord-stromal tumors, when given orally⁵⁹. The carcinogenic mechanisms of these chemicals with anti-estrogenic activity are not clear and further studies are needed to elucidate them.

Dioxin (2,3,7,8-TCDD) is known to exert its modulatory actions through the Ah receptor, and there is experimental evidence suggesting that it can also act in both estrogenic and anti-estrogenic manners, depending on the dose, species, and organ system involved. In rodents, TCDD induces mainly hepatocellular tumors. In addition, in an initiation-promotion study, morphological changes were also noted in both the uterus and the ovary. Although there is no evidence that TCDD can induce tumors in the female genital tract of rodents, it was reported to cause endometriosis in monkeys⁶⁰.

Another interesting example is ethylenethiourea (ETU), a metabolic product of ethylenebisthiocarbamate fungicides such as maneb and zineb, which are also listed as EDCs. ETU itself is a well established carcinogen, inducing thyroid tumors in rats and hepatic and lymphoid tumors in mice. In addition, it reacts with nitrite under acidic conditions in vitro and in vivo to form a mutagenic and carcinogenic compound, N-nitroso ETU⁶¹. Concurrent oral administration of ETU and sodium nitrite is reported to induce uterine endometrial adenocarcinomas in mice⁶². In our two-stage uterine carcinogenesis model using Donryu rats, concurrent oral administration of ETU (80 mg/kg) and sodium nitrite (56 mg/kg) resulted in uterine endometrial carcinomas without initiation by intrauterine administration of ENNG, and also promoted development of the tumors in animals initiated by ENNG, presumably by influencing the hormonal balance⁶³. Both ETU and nitrite are known environmental chemicals which are included in foods. Our confirmation that endometrial adenocarcinomas can be induced in this way in rats as well as mice, thus points to an importance of the oral route of exposure to these chemicals, although the doses used in the study were much higher than those in the diet.

The effects of high-dose t-OP on uterine carcinogenesis were investigated using adult Donryu rats initiated with a single intrauterine treatment of ENNG at 11 weeks of age and exposed thereafter to 100 mg/kg/day t-OP by s.c. injections until 15 months of age. Adult OVX rats were also treated in the same way. t-OP had no effect on the occurrence of persistent estrus in non-OVX rats, although uterotrophic effects were obvious in the OVX case. At the

Table 4. Uterine Adenocarcinomas in ENNG-initiated Rats with Exposure to High-dose OP*

Group	No. of rats examined	Incidence of endometrial lesions						
		Hyperplasia			total	Adenocarcinoma		
		+	++	+++		differentiation**		
					G1	G2	G3	
1. Control	23	2	8	7	4	4	0	0
2. OP-treated (Adulthood)	26	1	8	5	12***	9 (G1 and/or G2)		3
3. Control	23	3	7	5	6	6	0	0
4. OP-treated (PNDs 1-5)	28	1	3	5	18***	17	1	0
5. OP-treated (PNDs 1-15)	22	2	2	1	8	1***	3	4***

*: Groups 1-2: Katsuda *et al.*, *Jpn J Cancer Res* 2002; 93: 117-124.

Groups 3-5: Yoshida *et al.*, *Carcinogenesis* 2002; 23: 1745-1750.

** : Histological grades of uterine adenocarcinomas by tumor differentiation.

G1: well differentiated; G2: moderately differentiated; G3: poorly differentiated.

***: Significantly different from the control value ($p < 0.05$).

end of the experiment, however, development of uterine adenocarcinomas was significantly increased in animals exposed to t-OP during adulthood, but no tumors developed in OVX rats. This finding suggests that high-dose t-OP has tumor-promoting effects on the ENNG-treated endometrium of rats, possibly due to direct action on the uterus, as indicated by the uterotrophic effect of OP⁶⁴ (Table 4).

Uterine carcinogenesis in Donryu rats treated neonatally with a high-dose of t-OP has also been investigated. Female pups were subcutaneously administered 100 mg/kg/day t-OP every other day for the first 5 days after birth (PNDs 1-5), or the first 2 weeks (PNDs 1-15). Thereafter, they received a single intra-uterine injection of 20 mg/kg ENNG at 11 weeks of age and were observed until 15 months of age. PNDs 1-5 OP-treated animals showed normal development of the female reproductive system, including uterine gland genesis before weaning and normal estrous cycling immediately after vaginal opening. However, the treatment accelerated the occurrence of persistent estrus after 6 weeks of age, and increased the number of well-differentiated uterine adenocarcinomas at the end of the experiment (15 months of age), as compared with controls. This indicates that PNDs 1-5 OP-treatment resulted in delayed modulation of the hypothalamus-pituitary-ovarian hormonal control system, and thus increased the serum E2:P ratio, leading to promotion of uterine carcinoma development. On the other hand, PNDs 1-15 OP-treatment demonstrated immediate and irreversible influences on the control system, called "androgenization", and induced suppression of uterine gland genesis as well as abnormal uterine development manifested by prolonged persistent estrus immediately after vaginal opening, similar to our previous report³⁴. In addition, at the end of the experiment, uterine tumor malignancy as assessed by morphological and biological properties was clearly increased, although there was no significant alteration in the total incidence of adenocarcinomas. The total incidence of hyperplasias was significantly lowered, probably related to suppression of uterine gland genesis (Table 4). That study

provided evidence that neonatal exposure during PNDs 1-5 or 1-15 to high-dose t-OP enhances uterine carcinogenesis in ENNG-initiated rats, and that the type of uterine tumor is changed by the period of neonatal treatment³⁷.

Concerning the histogenesis of endometrial adenocarcinomas in Donryu rats, the tumors are considered to arise from hyperplasias of the luminal or glandular epithelium, especially the latter²⁸. In humans, it has been pointed out that the presence or absence of hyperplasia as the background is important for the biological behavior of endometrial adenocarcinomas. High-dose OP treatment at PNDs 1-15 induced luminal epithelial hyperplasia in the uteri of rats at 8 weeks of age³⁴, and finally increased development of undifferentiated adenocarcinomas, although the incidence of hyperplasias was decreased³⁷. Carthew *et al.* also reported that tamoxifen induced uterine adenocarcinomas, including biologically malignant examples, in rats in the absence of endometrial hyperplasia, when given on days 2-5 after birth⁵⁴. These results are very interesting in consideration of the histogenesis of uterine adenocarcinomas.

As mentioned above, estrogen and related compounds are reported to increase the risk of endometrial adenocarcinoma development in women. Estrogens occur naturally within the normal body, and are mainly metabolized in the liver by two separate pathways, producing either catechol estrogens (2- or 4-hydroxylated products) or 16 α - or 16 β -hydroxylated products. 2-Hydroxylation of estradiol or estrone to a catechol is a major metabolic pathway, and the catechol estrogens 2-OHE2 and 2-OHE1 have much weaker hormonal potency than their parent hormones, and lack carcinogenic potency when given to adult animals. On the other hand, 4-hydroxyestradiol (4-OHE2) and the two 16 α -hydroxylated forms, 16 α -OHE1 and 16 α -OHE2, retain potent hormonal activity by acting on classical estrogen receptor and also are tumorigenic⁶⁵. In fact, induction of preneoplastic and neoplastic lesions by estrogen and its steroid metabolites (16 steroids) were studied with our two-stage mouse uterine carcinogenesis

model, and 2-OHE1 or 2-OHE2 exerted promoting, but not progressing, effects, while 16 α - and 16 β -OHE1 caused both promotion and progression⁶⁶.

It is known that indole-3-carbinol binds to the Ah receptor, similar to TCDD, and induces cytochrome p450 metabolic enzymes mainly in the liver. It has been reported that this chemical shows a chemopreventive effect on spontaneous endometrial adenocarcinoma development in Donryu rats when given orally, the effect being speculated to be due to enhanced 2-hydroxylation⁶⁷. We also assessed the effect of indole-3-carbinol on uterine carcinogenesis using our two-stage rat uterine carcinogenesis model. Contrary to expectation, however, the incidences of endometrial carcinomas were increased. In rats given indole-3-carbinol, elevated liver weights and centrilobular enlargement of hepatocytes were also observed, the results indicating an effect on estrogen metabolism in the liver, and further studies are now under way, to clarify the discrepancy (Yoshida *et al.* unpublished data).

Effects of Low-doses of EDCs

The concentrations of EDCs including OP in the environment are very low, and the main exposure route is oral, rather than cutaneous, in humans. In general, the toxicokinetics of chemicals including EDCs in animals is known to be influenced by the method of administration. It has been reported that low doses of estrogens and EDCs such as OP might be removed from the blood during the first passage through the liver, when given orally^{68,69}. For risk assessment of EDCs, it is very important to investigate oral dose effects at human exposure levels and thus we have also focused on relatively low doses of OP (t-OP or n-OP) by oral administration. Female Donryu rats initiated by intrauterine administration of ENNG were given diets containing 100 or 1000 ppm t-OP (about 5 or 50 mg/kg/day) or 100 ppm n-OP (about 5 mg/kg/day) from 11 weeks of age to 15 months of age. Although the concentrations are higher than those in the environment, no significant increase in the incidences of uterine adenocarcinomas was observed in any treated group at the end of the experiment, and also there was no difference in tumor malignancy among the groups (Yoshida *et al.* unpublished data).

As detailed above, exposure to high doses of estrogens or EDCs in the fetal or new born period exerts irreversible androgenization of the female reproductive organs, because of heightened sensitivity. In addition, "delayed" influences on these organs may occur after puberty or sexual maturation. Therefore, relatively long-term comprehensive studies on the endocrinological and morphological aspects may be necessary for determination of prenatal and/or neonatal effects of low doses of EDCs regarding toxicity/carcinogenicity in the female genital organs. Low doses of EDCs such as NP or bisphenol A (BPA) were given orally to pregnant rats, and offspring were observed until 15 months of age, to investigate the prenatal and neonatal effects on growth and development of the female reproductive system

and uterine carcinogenesis. In the reproductive toxicity studies reported by others, high doses of NP caused estrogenic effects on pubertal development in male and female rats^{70,71}. However, maternal or neonatal exposure to relatively low doses demonstrated no adverse influence on the reproductive tract⁷². In our study with NP, dams were administered 0.1, 10 and 100 mg/kg daily by gavage from gestation day 2 up to the day before weaning of their offspring. Then, all female pups at 11 weeks of age were administered a single dose of 20 mg/kg ENNG into a uterine horn, and observed until 15 months of age. The low level, 0.1 mg/kg, was selected as a dose relevant to human daily intake (1 mg/kg) of isoflavones, uterotrophic activity of NP being reported to be 10 times stronger than that of daidzein, one of the major isoflavones, and the middle-dose, 10 mg/kg, was selected as near the no observed effect level in a multi-generation reproductive study using rats⁷¹. None of the treated groups demonstrated any alteration in reproductive ability. In their offspring also, uterine growth and development, vaginal opening and hormonal secretion until puberty were not changed and there were no effects on estrous cyclicity and morphology of the reproductive organs after maturation, or on uterine carcinogenesis in animals initiated with ENNG⁷³.

BPA, a volume chemical used in the manufacture of polycarbonate plastics and found in canned foods, lacquered containers and composite dental sealant, is one of the most representative EDCs with weak estrogenic activity, and uterotrophic potential has been demonstrated in the immature rat assay⁷⁴. A study conducted by the National Toxicity Program (NTP) in the USA demonstrated that maternal exposure to high doses of BPA at 0.5 or 1.0% in the diet (approximately daily intakes of 875 and 1750 mg/kg/day) reduced the number of live pups per litter and litters per pair in first generation mice⁷⁵, although pre- and/or postnatal high-dose BPA exposure did not have any apparent adverse effects on pubertal development in female rats or reproductive functions in rats and mice⁷⁶⁻⁷⁸. Recently, however, perinatal treatment with BPA at much lower doses has been described to influence male reproductive organ parameters such as weight of the testis, prostate, preputial gland and epididymis, and the efficiency of sperm production in rodents⁷⁹⁻⁸¹, and neonatal treatment advanced puberty in mice⁸², although there are also some reports of no treatment-related effects at low dose levels when given to pregnant mice and rats⁸³⁻⁸⁵, and to rats in a three-generation reproductive toxicity study⁸⁶. To further assess the risk, we also investigated effects of maternal exposure to low-doses of BPA, including a human exposure-level, on growth and development of the female reproductive system, and also uterine carcinogenesis in Donryu rats. Dams were administered BPA (0, 0.006 and 6 mg/kg/day) daily by gavage from gestation day 2 up to the day before weaning (PND 21). The concentration of 0.006 mg/kg was selected as consistent with the 63 ppb defined as the average daily intake from canned food in human beings, and 6 mg/kg was selected as appropriate to simulate the maximum dose level

Table 5. Proliferative Uterine Endometrial Lesions in Rats Given Low-doses of BPA*

Dose	Incidence of lesions (%)			Adenocarcinoma**
	Hyperplasia			
	+	++	+++	
0 mg/kg/day	21	21	13	33
0.006 mg/kg/day	20	20	17	33
6 mg/kg/day	13	47	17	20

*: Yoshida *et al.*, J Reprod Dev 2004; 50: 349–360.

**.: All adenocarcinomas in the three groups were well differentiated and limited to the uterus.

(80 ppm) detected in plastic plates⁸⁷. The treatment did not exert any influences on the reproductive system of female offspring in either treated group, in terms of prepubertal uterine growth and gland-genesis, vaginal opening and gonadotropin secretion. After maturation also, no effects were evident with regard to estrous cyclicity, age-matched sequential changes of the reproductive organs, and uterine carcinogenesis until 15 months of age (Table 5). The results demonstrated that maternal exposure to BPA at human exposure-levels did not have any adverse effects on the female reproductive organs of offspring in rats⁸⁸.

For determination of effects of EDCs on offspring by maternal treatment, biotransfer of the chemicals from dam to offspring is crucial, because the impact is fundamentally related to the serum EDC level⁴⁰. However, data for transfer of the test chemical via the placenta or milk to offspring, or toxicokinetics of low-dose EDCs are very limited⁸⁹. In one of our studies, NP at 10 and 100 mg/kg doses was transferred from dams to their offspring via the milk, but the compound could not be detected in their serum or liver⁷³. Furthermore, BPA levels in the milk of dams, and those in the serum and liver in offspring were comparable between control and treated groups, although the serum level of BPA in dams receiving 6 mg/kg was significantly elevated⁸⁸.

Quite recently, it was reported that cadmium has potent estrogen-like activity *in vivo*⁹⁰. Thus, exposure to low-dose cadmium (a single ip injection at a dose of 5 µg/kg) increased uterine net weight accompanied by proliferation of the endometrium, promoted growth and development of the mammary glands, and induced hormone-regulated genes in ovariectomized rats. *In utero* exposure to the metal (i.p. injections of 0.5 or 5 µg/kg on days 12 and 17 gestation) mimicked also the effects of estrogens, and female offspring experienced an earlier onset of puberty and an increase in the epithelial area and the number of terminal end buds in the mammary gland. The amounts of cadmium used in the study were environmentally relevant, because the WHO-recommended Provisional Tolerable Weekly Intake Level is 7 µg/kg/week. Although the administration route was intraperitoneal, not oral, the ability of environmentally relevant amounts of cadmium to mimic the effects of estradiol is very important and the metal may represent a new class of EDC.

Species Differences in Toxicologic/carcinogenic Effects of EDCs, Effects of EDCs based on the Molecular Biology and Extrapolation of the Effects to Humans

It is well known that the toxicokinetics of chemicals in animals are influenced by many factors, including species, strain, sex, age, dosage and/or administration method, as mentioned above. In particular, species differences are very important for the risk assessment in humans. Species differences in occurrence of toxicologic/carcinogenic effects of EDCs may be an indication of variation in endogenous hormonal factors, in addition to susceptibility to exogenous agents. Mice are generally more sensitive to estrogens than rats, and uterine adenocarcinomas can be induced in mice by estrogen alone, but not in rats. In mice, perinatal exposure to estrogens was found to induce ovary-independent proliferation of the vaginal epithelium, which could not be abolished by ovariectomy⁹¹. On the other hand, the vagina in rats neonatally exposed to high-dose t-OP became atrophic immediately after ovariectomy³⁴. Adenocarcinoma development in the ENNG-initiated endometrium of Donryu rats exposed to high-dose t-OP was also ovary-dependent⁶⁴. In mice, however, E2 promoted uterine adenocarcinoma development ovary-independently⁵⁶. Differences in the reaction to estrogens or EDCs with estrogenic activity may help to explain species differences in toxicity/carcinogenicity, though further studies on this point, focusing on metabolism of estrogens or EDCs and localization of ER expression, are needed.

The prenatal and/or neonatal periods are more sensitive to estrogens, and also EDCs, than the adult period. As mentioned above, it has been pointed out that tamoxifen increases the risk of endometrial cancer in women. In rodents, TAM can induce uterine carcinomas when given to newborn animals, but not adults. Similarly, 2-OHE2 has weak estrogenic, but not carcinogenic effects, although 4-OHE2 is a potent estrogenic catechol causing uterine tumors in adult mice. However, both catechols induced tumors when given on days 1–5 of neonatal life, although carcinogenic activity of 4-OHE2 is stronger than that of 2-OHE2⁹². Further studies on age-dependent differences in the mechanism of EDCs' effects on the female genital organs are also needed.

Various methods such as DNA micro-array techniques based on gene expression levels have recently been used for the evaluation of hazardous effects of various chemicals, because they should reveal very early changes. Up till now, however, there have only been a few reports concerning EDCs. The fact that changes in ontogenic expression of ER alpha and not of ER beta occur in the fetal female rat reproductive tract, provides fundamental information critical for clarifying species-specific physiological roles of ER subtypes during fetal development and for investigating the tissue-specific mechanisms underlying prenatal responses to estrogen and E2 agonists⁹³. Genome-wide analysis of early gene expression has furthermore suggested a basis for the

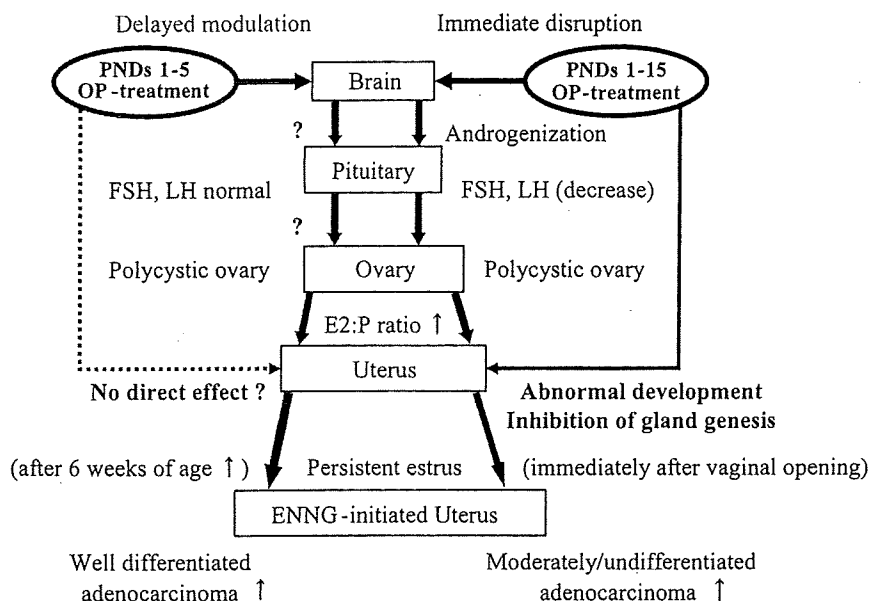


Fig. 3. Hypothesis of the scheme for "androgenization effects" or "delayed modulation effects" on the hypothalamo-pituitary-gonadal system in rats exposed neonatally to high-dose OP.

drastic uterotrophic effects following estrogen administration⁹⁴. Although DNA micro-array techniques presently demonstrate problems with reliability and reproducibility, future precise analysis should facilitate understanding of the mechanisms underlying effects of estrogenic EDCs⁹⁵.

Almost all EDCs exist at only very low concentrations in the environment, but humans may be exposed for long periods. In many animal studies, various toxicologic effects of EDCs on the female genital organs were demonstrated when very high doses were given, but no obvious effects were detected with low-doses. In humans also, low doses may show no adverse effects, because of homeostasis, although there are notable exceptions in animal and human studies. In fish, the ova-testis is known to be a good indicator of estrogenic effects of EDCs in females, but in female rodents, we are still lacking a consensus regarding equivalent reliable endpoint markers. Thus, more comprehensive studies of the endocrinological, morphological and also biomolecular aspects are necessary in animal studies using rodents for extrapolation of EDCs' effects to humans.

Conclusion

It is well known that the prenatal and/or neonatal period is particularly sensitive to various chemicals, including EDCs, in humans and rodents. Inappropriate exposure may exert irreversible influence, resulting in androgenization of the female genital system. In addition, it has also been reported that a delayed influence may be exerted. Neonatal exposure to a high dose of t-OP (100 mg/kg s.c. injection

every other day from PND 1 to PND 15) induced various long-term persistent irreversible effects on the female reproductive system of Donryu rats, such as lower gonadotropin levels at prepuberty, inhibition of uterine gland genesis, persistent estrus shown by vaginal cytology and polycystic ovaries. Neonatal treatment of high-dose EDCs having estrogenic activity can thus affect gonadotropin secretion during the developmental period of sexual maturation with direct masculinization of the hypothalamic function. Abnormal differentiation in the developing rat uteri may be induced via abnormal ER expression and subsequent alteration of cell proliferating activity. However, exposure limited to the first 5 days after birth to 100 mg/kg t-OP caused "delayed" influence which was characterized by accelerated appearance of atrophic ovary, manifested by an early-occurring and long-term continuing persistent estrus status after puberty, whereas no abnormalities could be found with regard to growth and development of the reproductive organs and the hypothalamo-pituitary-gonadal control system up to maturation. The hypothetical scheme for "androgenization effects" or "delayed modulation effects" on the hypothalamo-pituitary-gonadal system in rats exposed neonatally to high-dose OP is shown in Fig. 3.

On the other hand, the most notable effect on the female reproductive system when normal cycling rats were exposed to a high-dose of t-OP for a short time (28 days), was disappearance of the estrous cycle, and no clear changes were detected in other parameters such as uterine weight and morphology. These results indicate that the vaginal smear is the most sensitive parameter for the detection of effects of EDCs in normal cycling rats.

Well or moderately differentiated adenocarcinomas

were increased in Donryu rats initiated by ENNG, when high dose t-OP was given subcutaneously during adulthood. Neonatal exposure to a high dose of t-OP also showed promoting effects on uterine adenocarcinoma development in a two-stage rat uterine carcinogenesis model using Donryu rats, with slight to higher malignancy with more prolonged treatment.

For the risk assessment of EDCs to human health, it is very important to investigate the effects of low doses at actual human exposure levels, because the concentrations of agents, including alkylphenols and BPA, in the environment are very low. In addition, the main exposure route to EDCs is oral, not subcutaneous, in humans. Thus, we have focused on effects of maternal exposure to low doses of EDCs, such as NP and BPA, by the oral route, which have shown no effects on growth and development of the female reproductive system or uterine carcinogenesis. Transfer of low doses of BPA from dams to offspring via the placenta and/or milk was not unequivocal, although NP was transferred when relatively high doses were given.

These results indicate that dietary exposure to low doses of EDCs might not induce any adverse effects on the female genital system in mammals, including humans, because of the effects of homeostasis and clearance from the blood stream on first passage through the liver.

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